FINAL

Demonstration of Remediation by Natural Attenuation for Groundwater at Site FTA-2



Tinker Air Force Base Oklahoma City, Oklahoma

Prepared For

Air Force Center for Environmental Excellence
Technology Transfer Division
Brooks Air Force Base
San Antonio, Texas

and

Tinker Air Force Base Oklahoma City Oklahoma

DISTRIBUTION STATEMENT A

Approved for Public Release Distribution Unlimited

December 1999

20011030 090

FINAL TREATABILITY STUDY IN SUPPORT OF REMEDIATION BY NATURAL ATTENUATION FOR GROUNDWATER AT SITE FTA-2

at

TINKER AIR FORCE BASE OKLAHOMA

December 1999

Prepared for:

AIR FORCE CENTER FOR ENVIRONMENTAL EXCELLENCE
TECHNOLOGY TRANSFER DIVISION
BROOKS AIR FORCE BASE
SAN ANTONIO, TEXAS

and

TINKER AIR FORCE BASE OKLAHOMA

Prepared by:

PARSONS ENGINEERING SCIENCE, INC. 1700 BROADWAY, SUITE 900 DENVER, COLORADO 80290

EXECUTIVE SUMMARY

This report presents the results of a remediation by natural attenuation treatability study (RNA TS) performed by Parsons Engineering Science, Inc. (Parsons ES) at Fire Training Area 2 (FTA-2), Tinker Air Force Base (AFB), Oklahoma. The RNA TS was performed to evaluate the use of natural attenuation with long-term monitoring (LTM) as a remedial option for dissolved chlorinated aliphatic hydrocarbon (CAH) contamination in the upper saturated zone (USZ). The presence of groundwater contamination and soil contamination at the site was documented during remedial investigations performed by International Technology Corporation (IT) (1994) and Tetra Tech (TT) (1996). This TS focused on the impact of dissolved CAHs, primarily trichloroethene (TCE), on the USZ groundwater system at, and downgradient from, the site. Site history and the results of soil and groundwater investigations conducted previously also are summarized in this report.

It is likely that contaminants were first introduced into groundwater at this site when the FTA-2 fire pit was initially used for fire training exercises in 1962. TT (1996) reported the FTA-2 pit was operational until 1966, although aerial photos indicate the pit was present as late as 1970. Fire training exercises at Air Force bases typically use contaminated or uncontaminated jet fuel. Solvents, fuels, and other liquid wastes may have been used for training exercises or cleaning, or dumped in the general fire training area. Aerial photographs show numerous dirt access roads crossing the area through the mid-1970s, indicating unknown activities continued in the area after use of the FTA-2 pit was discontinued. Analytical data indicate that the highest concentrations of CAHs occur in the vicinity of well 2-62B, located approximately 100 to 120 feet upgradient from the former fire pit along an access road. The presence of a secondary source, located further upgradient from well 2-62B and the former fire pit along another access road near the flightline (well 2-65B), also is suggested by increasing concentrations of CAHs at that location.

Temporal data for concentrations of TCE and cis-1,2-dichloroethene (cis-1,2-DCE) in groundwater at FTA-2 indicate that concentrations in the potential source area(s) are stable or gradually increasing. Given the available data, it appears that the plume is expanding downgradient from the suspected primary FTA-2 source, and also at an upgradient location, based on increased concentration data. An increase in concentrations within the center of the CAH plume further indicate the presence of another secondary source, and/or expansion of the CAH plume in this area. Conversely, decreasing CAH concentration data for wells along the plume axis at the most northern portion of the plume indicate the plume extent may be either decreasing or approaching equilibrium in this area.

Several lines of chemical and geochemical evidence indicate that, although dissolved CAHs are undergoing biologically facilitated reductive dehalogenation, the occurrence of this process is limited and localized to the source area(s). The presence of daughter products *cis*-1,2-DCE and vinyl chloride (VC) provides strong evidence that parent solvents (TCE and/or tetrachloroethene [PCE]) are being reductively dehalogenated. However, a relative lack of VC and ethene indicates that the process does not

significantly proceed past the initial step of TCE to DCE. Molar fractions of chlorinated ethenes along the apparent CAH plume axis indicates little, if any, degradation of TCE to DCE outside the immediate source area(s). At well 2-65B, an overall decrease in the molar fraction of TCE, with an increase in molar fraction of DCE and the presence of VC indicates that TCE is degrading to DCE, and DCE to VC at that location. Geochemical data indicate that biodegradation of native or anthropogenic carbon is occurring via aerobic respiration and methanogenesis in suspected source areas; but that nitrate, ferric iron, and sulfate reduction are not significant anaerobic degradation processes at the site. Furthermore, there are an abundance of alternate electron acceptors that may inhibit use of CAHs as electron acceptors. Away from the suspected source areas, dissolved oxygen (DO) concentrations and oxidation-reeduction potentials (ORPs) indicate that environmental conditions are not sufficiently reducing for dehalogenation of CAH compounds.

An important component of this study was an assessment of the potential for groundwater contamination to migrate from the source area to potential receptors at concentrations above regulatory levels intended to be protective of human health and the environment. To accomplish this objective, the numerical model codes MODFLOW and MT3D were used to estimate the impacts of RNA and potential engineered remedial actions on the future migration and persistence of dissolved TCE within the USZ under the influence of advection, dispersion, sorption, and biodegradation. Input parameters for the numerical model were obtained from existing site characterization data, supplemented with data collected during the RNA TS. Model parameters that were not measured at the site were estimated using reasonable literature values.

The results of this study suggest that, even with the implementation of all evaluated remedial actions (Alternatives 1 through 3), dissolved TCE contamination present in groundwater north and west of the FTA-2 fire pit has the potential to migrate at least 1,000 feet beyond the August 1997 plume toe within 35 years unless geochemical conditions are encountered along the plume flowpath that promote increased biodegradation rates. In addition, the model suggests that dissolved TCE concentrations in excess of the US Environmental Protection Agency (USEPA) maximum contaminant level (MCL) groundwater quality standard of 5 micrograms per liter (µg/L) have the potential to persist for more than 100 years (the duration of the numerical model predictive period). The model predicts that source characterization and removal could potentially cause the maximum dissolved TCE concentrations at downgradient observation points during the 35-year simulation period to be reduced by approximately one-half to one-fourth.

It should be noted that the numerical model constructed for this TS is reasonably conservative; therefore, plume migration and persistence may be less than that predicted by the model. Factors that could cause plume migration and persistence to differ from model predictions include the following:

- The calibrated model is not necessarily unique, and different combinations of input parameters could potentially have been used to achieve an acceptable calibration;
- The calibrated model is limited in accuracy by inadequate characterization of the source, preferential flowpaths, and aquifer anisotropy;

- The model does not account for the potential occurrence of abiotic degradation and volatilization of TCE; and
- Hydrogeologic and/or geochemical conditions that could slow or halt plume migration may exist downgradient from the investigated area.

This report presents a conceptual model where the primary source of TCE was near the FTA-2 fire pit and well 2-62B, and the TCE plume migrated toward the northwest due to a preferential flow path and anisotropy of the aquifer matrix. This conceptual model was primarily based on the plume configuration in 1997, which suggests linear contaminant transport in a northwest direction. Fate and transport modeling results indicate that dissolved TCE contamination present in groundwater north and west of the FTA-2 fire pit has the potential to migrate significantly. However, concentrations for wells located at the northern toe of the TCE plume decreased substantially from 1997 to 1999. Alternately, the TCE plume may be migrating in the direction of maximum hydraulic The presence of multiple TCE gradient (west-southwest) from multiple sources. "hotspots" does not support a conceptual model of a single source; rather, the "hotspots" suggest multiple sources. Increases in the concentration of TCE in 1999 at wells located downgradient of suspected TCE sources suggest a more westerly migration in the direction of maximum hydraulic conductivity. Therefore, the numerical model presented in this TS should be considered as only one of multiple possible interpretations.

Natural attenuation with source characterization and source reduction (Alternative 2) is recommeded for the FTA-2 Site. Additional site characterization is necessary to accurately predict the future behavior of dissolved contaminants at FTA-2, and may include a soil gas survey, a geophysical survey, aquifer testing, and tracer tests. Future exposure to potential receptors at the site can be minimized by land use restrictions, because access to the Base (and hence the site) is controlled by Base security. The beneficial affects of natural attenuation are included in Alternative 2. The degree to which RNA will continue to control significant downgradient migration of the plume is not known with certainty, but will become more apparent as successive LTM sampling events are performed.

Regular sampling of 22 existing and 8 new LTM wells, and 2 surface water stations, is recommended to monitor the long-term migration and degradation of the dissolved CAH plumes. The plumes should be progressively tracked in the downgradient direction, and additional downgradient sentry wells should be installed/sampled as required. Regular sampling and analysis of groundwater from these wells will allow the effectiveness of RNA and engineered remedial actions to be monitored, and should allow assessment of whether additional engineering controls should be implemented. Likewise, the model can be adjusted to reflect additional source and hydrogeologic characterization.

Contaminant fate and transport model results indicate that sampling should continue on an annual basis for approximately 15 years (the anticipated maximum duration of engineered remedial actions), followed by less frequent (e.g., every other year) sampling. The LTM plan should be periodically reevaluated and modified as necessary on the basis of newly obtained data and calibration of the numerical model.

TABLE OF CONTENTS

	Page
ACRO	NYMS AND ABBREVIATIONSviii
SECTI	ON 1 - INTRODUCTION
1.1 1.2	Scope and Objectives
SECTI	ON 2 - SITE CHARACTERIZATION ACTIVITIES2-1
2.1	Soil Sample Collection and Groundwater Monitoring Well/Point Installation
2.2 2.3 2.4 2.5	Groundwater Sampling
SECT	ION 3 - PHYSICAL CHARACTERISTICS OF THE STUDY AREA 3-1
3.1 3.2	Topography, Surface Hydrology, and Climate
3.3	FTA-2 Geology and Hydrogeology 3-11 3.3.1 Site Geology 3-11 3.3.2 Site Hydrogeology 3-11 3.3.2.1 Upper Saturated Zone 3-16 3.3.2.2 Lower Saturated Zone 3-19
3.4 3.5	Groundwater Use

					Page
SECTI	ON 4 -	CONTAN	MINANT DIS	STRIBUTION AND EVIDENCE FO	R
DECT	.011 1	BIODEC	GRADATION	V	4-1
					4.1
4.1				odegradation	
	4.1.1			ation Processes	
	4.1.2			lorinated Solvents	4-3
		4.1.2.1	Electron Ac	cceptor Reactions (Reductive	4.2
		4 1 0 0	Denalogena	ation)	4-3
		4.1.2.2		onor Reactions	
		4.1.2.3		sm	
		4.1.2.4		gradation of Chlorinated Solvents	
		4.1.2.5		f Chlorinated Solvent Plumes	
			4.1.2.5.1	Type 1 Behavior	4.10
			4.1.2.5.2	Type 2 Behavior	4-10
			4.1.2.5.3 4.1.2.5.4	Type 3 Behavior Mixed Behavior	
4.2	Matura	and Exta		ination	
4.2	4.2.1			illation	
	4.2.1				
	4.2.3			in SoilGroundwater	
	4.2.3	Discolar	Chlorinoto	l Ethenes	4-13
	4.2.4			l Ethanes	
	4.2.5			zenes	
	4.2.7	Other Die	realwad Chla	rinated Compounds	4-24
	4.2.7	I S7 Grov	undwater Ou	ality	4-24
4.3				degradation	
4.3	4.3.1	Field Sco	ile Contamin	ant Mass Loss	4-26
	4.5.1	4.3.1.1		l Ethenes Over Time	
		4.3.1.2		Ethanes Over Time	
		4.3.1.3		zenes Over Time	
	4.3.2			Products and CAH Ratios	
	4.3.3			for of Dehalogenation	
	4.3.4			of Redox Processes	
	4.3.5				
	4.5.5	4.3.5.1		Total Fuel Carbon in Groundwater	
	(*)	4.3.5.2		rbon in Groundwater	
		4.3.5.3		Hs as Electron Donors	
	4.3.6			ceptors and Metabolic Byproducts	
	1.5.0	4.3.6.1		Oxygen	
		4.3.6.2		rite	
		4.3.6.3		n	
		4.3.6.4			
		4.3.6.5		thane and Ethene in Groundwater	
		4.3.6.6		Ammonium	

				Page
	4.3.7	Addition	al Geochemical Indicators	4-45
		4.3.7.1	Alkalinity and Carbon Dioxide Evolution	
		4.3.7.2	pH	4-47
		4.3.7.3	Temperature	4-47
	4.3.8		g Table for CAH Degradation	
4.4		ximation o	of Biodegradation Rates	4-47
4.5	Summ	ary		4-54
aram.	10NI 5	CDOINI	NUATED ELOMAND CONTAMINANT	
SEC1.	ION 5 -		DWATER FLOW AND CONTAMINANT	5 1
		IKANS	PORT MODEL	3-1
5.1	Canar	al Overvie	w and Model Description	5-1
5.2	Model	Objective	es	5-2
5.3	Conce	ntual Mod	lel Design and Assumptions	5-2
5.4	Initial	Model Se	tup	5-3
3.4	5.4.1		ign	
	5.4.2	Model I	ayers	5-3
	5.4.2	Groundy	vater Flow Model	5-5
	3.4.3	5.4.3.1		
		5.4.3.1	Recharge and Evapotranspiration	
			Aquifer Properties	
		5.4.3.3		
		5.4.3.4	Aquifer Stresses	
	511	5.4.3.5	Aquifer Storagenant Transport Model	
	5.4.4			
		5.4.4.1	Source	
		5.4.4.2	Dispersivity	5-9
		5.4.4.3	Retardation	
	36 11	5.4.4.4	Biodegradation	
5.5			on	
	5.5.1		vater Flow Model	
	5.5.2		alibration	
		5.5.2.1	Source Concentrations	
		5.5.2.2	Dispersivity	
		5.5.2.3	TCE Decay Rates	
		5.5.2.4	Sorption	
		5.5.2.5	Effective Porosity	5-16
		5.5.2.6	Transport Calibration Resultsysis	5-16
5.6		ivity Anal	ysis	5-18
	5.6.1		ty to Variations in Hydraulic Conductivity	
	5.6.2		ty to Variations in the Distribution Coefficient	
	5.6.3	Sensitivi	ty to Variations in Dispersivity	5-20
	5.6.4	Sensitivi	ty to Variations in the Decay Rate Constant	5-20
	5.6.5		ty to Variations in Effective Porosity	
	5.6.6	Summar	y of Sensitivity Analysis Results	5-21

	Page
5.7	Discussion of the Model and Recommendations to Reduce Model Uncertainties
SECT	ION 6 - COMPARATIVE ANALYSIS OF REMEDIAL ALTERNATIVES
6.1	Remedial Alternative Evaluation Criteria 6-1 6.1.1 Long-Term Effectiveness and Permanence 6-1 6.1.2 Implementability 6-2 6.1.3 Cost 6-2
6.2	Factors Influencing Alternatives Development 6-2 6.2.1 Program Objectives 6-2 6.2.2 Contaminant Properties 6-3 6.2.3 Site-Specific Conditions 6-4 6.2.3.1 Physical Characteristics 6-4 6.2.3.2 Geochemical Characteristics 6-5 6.2.3.3 Potential Receptor Exposure Pathways 6-5 6.2.3.4 Remediation Goals for USZ Groundwater 6-6
	6.2.4 Summary of Remedial Option Screening
6.3	Brief Description of Remedial Alternatives
	6.3.1 Alternative 1 - RNA and Institutional Controls with Long-Term Groundwater Monitoring
	6.3.2 Alternative 2 – Source Characterization and Removal, RNA, and Institutional Controls with Long-Term Groundwater Monitoring 6-13
	6.3.3 Alternative 3 - Groundwater Extraction and Treatment, Source Characterization and Removal, RNA, and Institutional Controls with Long-Term Groundwater Monitoring
6.4	Evaluation of Alternatives 6-14 6.4.1 Alternative 1 - RNA and Institutional Controls with Long-Term Groundwater Monitoring 6-14 6.4.1.1 Effectiveness 6-14 6.4.1.2 Implementability 6-16 6.4.1.3 Cost 6-17
	6.4.2 Alternative 2 – Source Characterization and Removal, RNA, and Institutional Controls with Long-Term Groundwater Monitoring6-18 6.4.2.1 Effectiveness
	6.4.3 Alternative 3 -Groundwater Extraction and Treatment, RNA, and Institutional Controls with Long-Term Groundwater Monitoring 6-20 6.4.3.1 Effectiveness

	Page
6.5	6.4.3.2 Implementability
SECTION	ON 7 - LONG-TERM MONITORING PLAN7-1
7.2 7.3 7.4 7.5	Overview
blem	
SECTI	ON 9 - REFERENCES
APPEN	NDICES
B - Geo Di Ar C - Lab D - Cos E - Sup Tr F - Mos G - Res H - Fin	ected Figures and Tables from the Phase I RCRA RFI for FTA-2 (T, 1994), and the Draft RCRA RFI for FTA-2 (TT, 1996) clogic Borehole Logs, Geophysical Borehole Logs, Well Construction agrams, Monitoring Well Development and Sampling Records, Slug Test nalyses, and Survey Data coratory Analtyical Data st Calculations oporting Model Input and Calculations for Groundwater/Fate and cansport Modeling del Input and Output Files sponses to Comments al Addendum to the Treatability Study in Support of Remediation by Natural enuation for Groundwater at Site FTA-2, Tinker Air Force Base, Oklahoma

LIST OF TABLES

No.	Title	Page
2.1	Summary of Treatability Study Activities	2-2
2.2	Monitoring Well Summary and August 1997 Groundwater Elevations	
2.3	Analytical Protocols for Groundwater and Soil Samples	2-9
3.1	Major Geologic Units in the Vicinity of Tinker AFB	

LIST OF TABLES (Continued)

No.	Title	Page
3.2	Upper Saturated Zone Hydraulic Conductivities and Average	0.10
	Groundwater Velocities	3-18
4.1	PCE, TCE, and TOC in Soil	
4.2	Fuel Hydrocarbons in Groundwater	
4.3	Chorinated Aliphatic Hydrocarbons in Groundwater	4-16
4.4	Groundwater Geochemical Data	4 40
4.5	Analytical Parameters and Weighting for Preliminary Screening	4-49
4.6	Interpretation of Points Awarded During Natural Attenuation Screening	4-51
4.7	Summary of Chlorinated Solvent Decay Rates	
5.1	Common Designations for Several Important Boundary Conditions	
5.2	Calculation of Retardation Coefficients	
5.3	Summary of Sensitive Analysis Results	3-19
6.1	Groundwater Quality Standards	0-7
6.2	Initial Technical Implementability Screening of Technologies and	60
	Process Options for Groundwater Remediation	0-0
6.3	Alternative 1 - Cost Estimate	
6.4	Alternative 2 - Cost Estimate	
6.5	Alternative 3 - Cost Estimate	6 25
6.6 7.1	Summary of Remedial Alternatives Evaluation	7.5
7.1	Long-Term Monitoring Analytical Protocol for Groundwater Estimated LTM Costs	7-7
	LIST OF FIGURES	
No.	Title	Page
1.1	Location of Tinker Air Force Base	
1.2	Location of Fire Training Area 2	1-6
1.3	Sampling Locations Fire Training Area 2	
3.1	Topographic Map of Fire Training Area 2	3-2
3.2	Surface Geologic Map Tinker Air Force Base	3-6
3.3	Locations of Hydrogeologic Cross-Sections A-A' and B-B'	3-8
3.4	Hydrogeologic Cross-Section A-A'	
3.5	Hydrogeologic Cross-Section B-B'	3-10
3.6	Cross-Section Location Map	5-12
3.7	Hydrogeologic Cross-Section C-C	5-13
3.8	Hydrogeologic Cross-Section D-D'	3-14
3.9	Hydrogeologic Cross-Section E-E'	3-13

LIST OF FIGURES (Continued)

No.	Title	Page
3.10	Upper Saturated Zone Potentiometric Surface August 1997	3-17
3.11	Lower Saturated Zone Potentiometric Surface	3-20
4.1	Anaerobic Reductive Dehalogentation	4-5
4.2	Aerobic Degradation Pathways	4-8
4.3	Concentration of TCE in USZ Groundwater August 1997	4-18
4.4	Concentration of cis-1,2-DCE in USZ Groundwater August 1997	4-19
4.5	Concentration of trans-1,2-DCE in USZ Groundwater August 1997	4-20
4.6	Detected Concentrations of PCE, 1,1-DCE, and VC in USZ Groundwater	
	August 1997	4-21
4.7	Concentration of 1,2-DCE in USZ Groundwater August 1997	4-23
4.8	Detected Concentrations of Chlorinated Benzene in USZ Groundwater	4-25
4.9	Total Chlorinated Ethenes Versus Time	4-27
4.10	TCE, cis-1,2-DCE, trans-1,2-DCE, and VC Versus Time at 2-61B	4-29
4.11	Total Chlorinated Ethanes Versus Time	4-30
4.12	Total Chlorobenzenes Versus Time	4-31
4.13	Molar Fraction of TCE to cis-1,2-DCE Versus Distance	4-32
4.14	Concentrations of Chloride in USZ Groundwater August 1997	4-35
4.15	Sequence of Microbially Mediated Redox Processes USZ Groundwater	4-37
4.16	Redox Potential of USZ Groundwater August 1997	4-38
4.17	Concentration of Dissolved Oxygen in USZ Groundwater August 1997	4-41
4.18	Concentration of Nitrate/Nitrate (as N) in USZ Groundwater August	
	1997	4-42
4.19	Concentration of Sulfate in USZ Groundwater August 1997	4-44
4.20	Concentration of Total Alkalinity in USZ Groundwater August 1997	4-46
4.21	Concentration of Carbon Dioxide in USZ Groundwater August 1997	4-48
5.1	Model Grid Superimposed on FTA-2 Site Location	5-4
5.2	Calibrated Groundwater Surface, Layer 1	5-12
5.3	Calibrated Versus Observed Groundwater Elevations	5-14
5.4	Calibrated TCE Plumes in USZ Groundwater	5-17
6.1	Predicted TCE Concentrations in USZ Groundwater After 35 Years for	
	Alternative 1	6-15
6.2	Predicted TCE Concentrations in USZ Groundwater After 35 Years for	
	Alternative 2	6-19
6.3	Predicted TCE Concentrations in USZ Groundwater After 10 Years of	
	Pumping for Alternative 3	6-22
7.1	LTM Sampling Locations Fire Training Area 2	7-3

ACRONYMS AND ABBREVIATIONS

2D two-dimensional
3D three-dimensional
°C degrees centigrade
°F degrees Fahrenheit

 $\Delta G^{\circ}r$ Gibbs free energy of the reaction

 $\begin{array}{ll} \mu g/kg & \text{micrograms per kilogram} \\ \mu g/L & \text{micrograms per liter} \end{array}$

μs/cm microsiemens per centimeter

AFB Air Force Base

AFCEE Air Force Center for Environmental Excellence

AEI Associated Environmental Industries

AOC Area of Concern

ASCII American Standard Code for Information Exchange

bgs below ground surface
BR Brown and Root

BTEX benzene, toluene, ethylbenzene, and xylenes

btoc below top of casing
CaCO₃ calcium carbonate
CB chlorobenzene
DCA dichloroethane
DCB dichloroethene
DCE dissolved oxygen

DOD Department of Defense FTA-2 Fire Training Area 2

ft/day feet per day
ft/ft feet per foot
ft/mile feet per mile
ft/yr feet per year

gpm gallons per minute

GWTP groundwater treatment plant HDPE high density polyethylene

HWBZ Hennessey Water-Bearing Zone

ID inside diameter iron (II) ferrous iron ferric iron

IRP Installation Restoration Program

IT International Technology Corporation

 K_d soil distribution coefficient K_{oc} soil sorption coefficient

kg/L kilograms per liter
L/kg liters per kilogram
LSZ lower saturated zone
LTM long-term monitoring

MCL maximum contaminant level
mg/kg milligrams per kilogram
mg/L milligrams per liter
mm Hg millimeters of mercury
MOC method of characteristics

mV millivolts N nitrogen

NAPL non-aqueous phase liquid

NGVD National Geodetic Vertical Datum

NR not recorded

NRMRL National Risk Management Research Laboratory

ORD Office of Research and Development

ORP oxidation-reduction potential

OSWER Office of Solid Waste and Emergency Response

PA/SI preliminary assessment/site inspection

Parsons ES Parsons Engineering Science, Inc.

PCA tetrachloroethane
PCE tetrachloroethene

PID photoionization detector
POC point of compliance
PVC polyvinyl chloride

RCRA Resource Conservation and Recovery Act

redox reduction-oxidation

RFI RCRA facility investigation

RMS root mean squared

RNA remediation by natural attenuation

SAP sampling and analysis plan SVOC semivolatile organic compound SWMU Solid Waste Management Unit

TCA trichloroethane
TCE trichloroethene
TeCB tetrachlorobenzene

TMB trimethylbenzene
TOC total organic carbon

TPH total petroleum hydrocarbons

TS Treatability Study
TT Tetra Tech, Inc.

USACE United States Army Corps of Engineers

USEPA United States Environmental Protection Agency

USGS United States Geological Survey

USZ upper saturated zone

VC vinyl chloride

VOC volatile organic compound

SECTION 1

INTRODUCTION

This report was prepared by Parsons Engineering Science, Inc. (Parsons ES) and presents the results of a treatability study (TS) conducted to evaluate the use of remediation by natural attenuation (RNA) for groundwater contaminated with chlorinated aliphatic hydrocarbons (CAHs) at the former Fire Training Area 2 (FTA-2) located at Tinker Air Force Base (AFB), in Oklahoma City, Oklahoma (the Base). As used in this report, RNA refers to a management strategy that relies on natural attenuation mechanisms to remediate contaminants dissolved in groundwater and to control receptor exposure risks associated with contaminants in the subsurface. The United States Environmental Protection Agency (USEPA) Offices of Research and Development (ORD) and Solid Waste and Emergency Response (OSWER) define natural attenuation as (Wilson, 1996):

The biodegradation, dispersion, sorption, volatilization, and/or chemical and biochemical stabilization of contaminants to effectively reduce contaminant toxicity, mobility, or volume to levels that are protective of human health and the ecosystem.

As suggested by this definition, mechanisms for natural attenuation of CAHs and benzene, toluene, ethylbenzene, and xylenes (BTEX) include advection, dispersion, dilution from recharge, sorption, volatilization, abiotic chemical transformation, and biodegradation. Of these processes, biodegradation is the predominant mechanism working to transform contaminants into innocuous byproducts. During natural biodegradation, indigenous microorganisms work to bring about a reduction in the total mass of contamination in the subsurface without the engineered addition of nutrients. Patterns and rates of natural attenuation can vary markedly from site to site depending on governing physical and chemical processes.

RNA is advantageous for the following reasons:

- Contaminants can be transformed to innocuous byproducts (e.g., carbon dioxide, ethene, or water), not just transferred to another phase or location within the environment;
- Current pump-and-treat technologies are energy-intensive and generally not effective in reducing residual contamination;
- The process is nonintrusive and allows continuing use of infrastructure during remediation;

- Engineered remedial technologies may pose a greater risk to potential receptors than RNA (e.g., contaminants may be transferred into another medium during remediation activities); and
- RNA is less costly than conventional, engineered remedial technologies.

A potential disadvantage of RNA is that, in some cases, natural attenuation rates are too slow to make RNA a practical remedial alternative. In addition, biodegradation of highly chlorinated compounds, such as trichloroethene (TCE), can produce vinyl chloride, which is relatively toxic. Under certain geochemical conditions, vinyl chloride may accumulate in the environment rather than being transformed to innocuous byproducts.

The main emphasis of the work described herein was to evaluate the potential for naturally occurring degradation mechanisms to reduce dissolved CAH concentrations in groundwater to levels that are protective of human health and the environment. This study is not intended to be a contamination assessment report or a remedial action plan; rather, it is provided for the use of the Base and its prime environmental contractor(s) as information to be used for future decision making regarding this site.

1.1 SCOPE AND OBJECTIVES

Parsons ES, in conjunction with researchers of the USEPA National Risk Management Research Laboratory (NRMRL), was retained by the United States Air Force Center for Environmental Excellence (AFCEE) Technology Transfer Division to conduct site characterization and groundwater modeling to evaluate the scientific defensibility of RNA with long-term monitoring (LTM) as a remedial option for contaminated groundwater at FTA-2.

The following tasks were performed to fulfill the project objectives:

- Review existing hydrogeologic and soil/groundwater quality data for the site;
- Conduct site characterization activities to more thoroughly characterize the nature and extent of groundwater contamination;
- Collect geochemical data in support of RNA;
- Develop a conceptual hydrogeologic model of the upper saturated zone (USZ), separated by a regional clay-silt aquitard from the lower saturated zone (LSZ);
- Develop an understanding of the current contaminant distribution in the USZ and LSZ;
- Evaluate site-specific data to determine whether naturally occurring processes of contaminant attenuation and destruction are occurring in USZ groundwater at the site;

- Design and execute a groundwater flow and contaminant fate and transport model for site USZ hydrogeologic conditions;
- Simulate the fate and transport of CAHs (e.g., TCE) in groundwater under the influence of advection, dispersion, adsorption, and biodegradation using the calibrated model;
- Evaluate a range of model input parameters to determine the sensitivity of the model to those parameters and to consider several contaminant fate and transport scenarios;
- Determine if natural processes are minimizing dissolved CAH plume expansion so that groundwater quality standards can be met at a downgradient point of compliance (POC);
- Conduct a preliminary exposure pathways analysis for potential current and future receptors;
- Use modeling results to asses the effect of current or planned future remedial actions on the TCE plume; and
- Provide a LTM plan that includes LTM and POC wells and a sampling and analysis plan (SAP).

Field work conducted under this program was oriented toward the collection of supplementary hydrogeologic and chemical data necessary to document and model the effectiveness of RNA with LTM for restoration of CAH-contaminated groundwater. Site characterization activities in support of RNA included static groundwater level measurement and groundwater sample collection and analysis from preexisting and recently installed site monitoring wells, and from monitoring wells and monitoring points installed by Parsons ES during the field program. Additional groundwater sampling was performed by the USEPA NRMRL in April 1999. Data for this event were not available for analysis in preparation of this TS, but have been evaluated in an addendum to this TS (Appendix H). Results of the addendum have been incorporated into the Executive Summary and Section 8, Conclusions and Recommendations.

Site-specific data were used to develop a solute fate and transport model for the site and to conduct a preliminary receptor exposure pathways analysis. The modeling effort was used to predict the future extent and concentration of the dissolved CAH plume by modeling the combined effects of advection, dispersion, sorption, and biodegradation. Results of the model were used to assess the potential for completion of exposure pathways involving groundwater, and to identify whether RNA with LTM is an appropriate and defensible remedial option for contaminated groundwater. The results will be used to provide technical support for the RNA with LTM remedial option during regulatory negotiations, as appropriate.

This TS contains nine sections, including this introduction, and eight appendices. Section 2 summarizes site characterization activities. Section 3 summarizes the physical characteristics of the study area. Section 4 describes the nature and extent of soil and

groundwater contamination, and the evidence of contaminant biodegradation in groundwater at the site. Section 5 describes the fate and transport model and design of the conceptual hydrogeologic model for the site; lists model assumptions and input parameters; and describes sensitivity analysis results. Section 6 describes the predicted effects of potential future remedial actions on the TCE plume. Section 7 presents a LTM plan for the site. Section 8 presents a summary and conclusions, and Section 9 lists the references used to develop this document. Appendix A contains pertinent figures and tables from previous reports such as the Phase I Resource Conservation Recovery Act (RCRA) facility investigation report (RFI) for FTA-2 [IT Corporation (IT), 1994], and the Draft RCRA RFI for FTA-2 [Tetra Tech, Inc.(TT), 1996]. Appendix B contains well development and sampling records, borehole logs, well construction diagrams, geophysical logs for newly installed monitoring wells, and survey data. Appendix C presents groundwater analytical results for samples collected as a part of this TS. Appendix D contains cost calculations for remedial alternatives. Appendix E contains model input parameters and supporting calculations. Appendix F contains model input and output in American Standard Code for Information Interchange (ASCII) format on a diskette. Appendix G contains responses to comments on the draft TS, and Appendix H contains the Final Addendum to the TS report.

1.2 BASE AND FTA-2 BACKGROUND

Tinker AFB covers approximately 5,000 acres in the southeastern Oklahoma City metropolitan area (Figure 1.1). The Base has supported air operations since its founding as the Midwest Air Depot in July 1941. The Base was formally activated by the Air Force in March 1942 and serves as an international repair depot for a variety of aircraft, weapons, and engines.

Tinker AFB currently operates under a RCRA Hazardous Waste Management Permit issued by the USEPA (July 1, 1991). This permit requires Tinker AFB to investigate all solid waste management units (SWMUs) and Areas of Concern (AOCs), including FTA-2, and to perform corrective action at those sites identified as posing a threat to human health or the environment. Because the Base is a Department of Defense (DOD) facility, site investigation and remediation have progressed through the stages outlined under the Installation Restoration Program (IRP).

Site FTA-2 is located in the south-central portion of the Base, west of Tower Road and south of the main northwest to southeast runway (Figure 1.2). FTA-2 was established as a temporary, unlined pit and was used infrequently for fire training exercises between 1962 and 1966. Standard operating procedures reportedly included adding water to the pit to saturate the soil and reduce fuel infiltration. Fuel was then brought in by tank truck, placed on top of the water, ignited, and extinguished using water and foam. Any residues were left in the pit to evaporate and infiltrate until the next fire training exercise. Available aerial photographs indicate that the site is currently a gently sloping, open grassy area with no visible evidence of former training practices. Records for facility operation do not exist. Therefore, data on composition, frequency, and quantity of fuel or extinguishing agents are not available.

OKLAHOMA

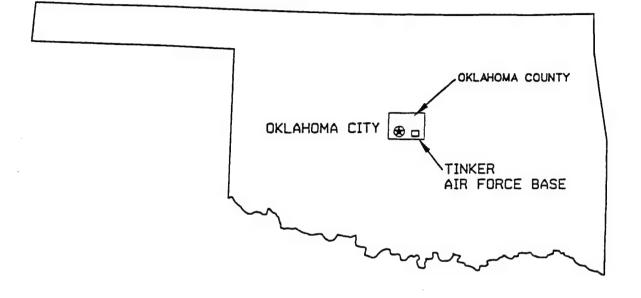




FIGURE 1.1

LOCATION OF TINKER AIR FORCE BASE

Site FTA 2 RNA TS Tinker AFB, Oklahoma

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

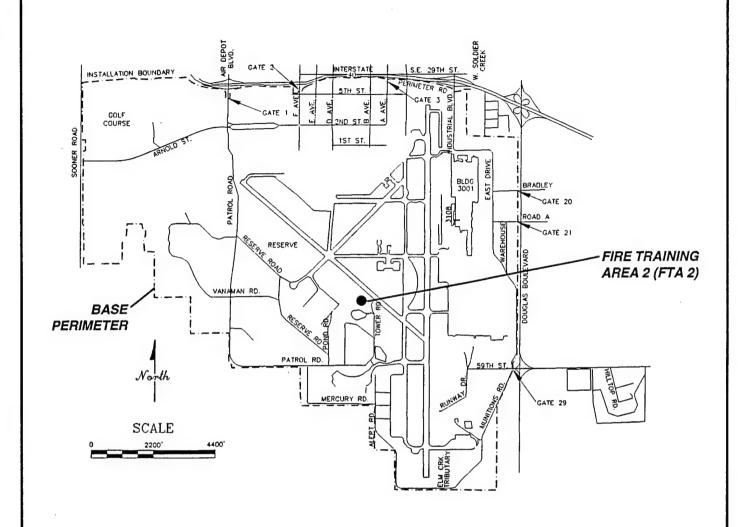


FIGURE 1.2

LOCATION OF FIRE TRAINING AREA 2

Site FTA 2 RNA TS Tinker AFB, Oklahoma

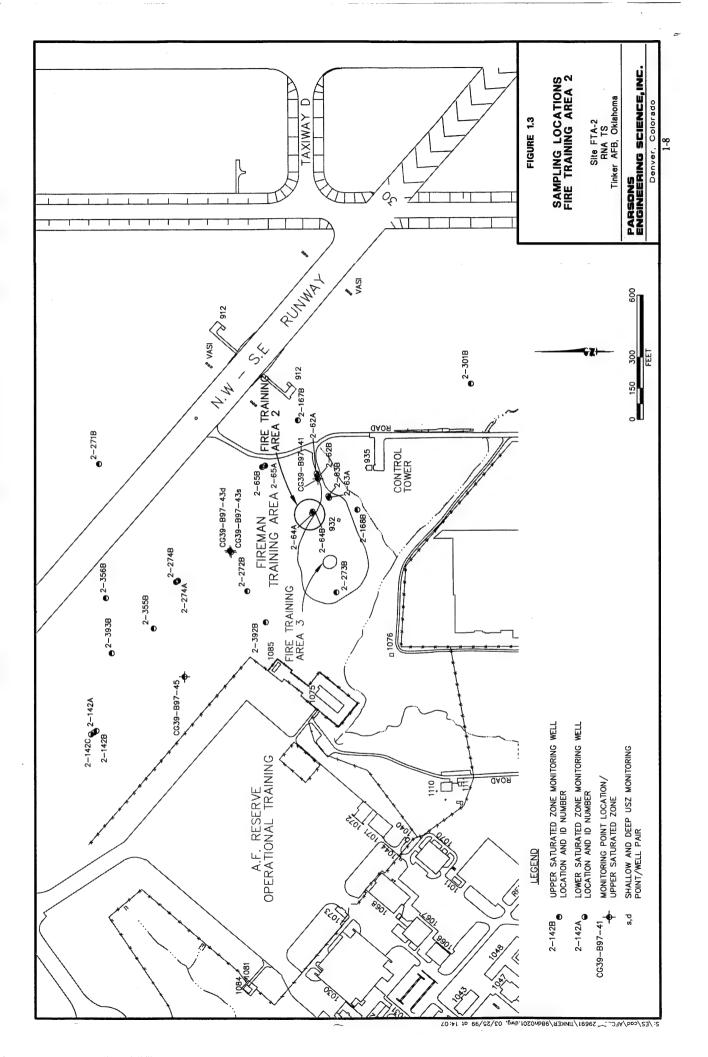
PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

In 1981, a Phase I study similar to a preliminary assessment/site inspection (PA/SI) was conducted at the Base by Engineering Science, Inc. (ES, 1982) as part of the Air Force IRP. The study identified 14 potentially contaminated sites, including Site FTA-2. Based on the findings of this study, the following tasks have been completed at Site FTA-2:

- An IRP Response Action was performed by the United States Army Corps of Engineers (USACE) in 1987 (USACE, 1988). Soil samples from seven boreholes, drilled to a maximum depth of 7 feet below ground surface (bgs), were analyzed for volatile organic compounds (VOCs), semi-volatile organic compounds (SVOCs), and total metals.
- IT conducted a Phase I RCRA RFI from October through December 1993, including soil sampling and the installation of four monitoring well pairs (IT, 1994). Soil and groundwater samples were analyzed for VOCs, SVOCs, and metals.
- A Phase II RFI was performed by TT, in 1995 (TT Draft Report, 1996). Soil samples were collected from five soil boreholes and two additional boreholes constructed for monitoring wells. Soil analytical results were included in the Draft Report. The monitoring wells were subsequently completed and sampled by Tinker AFB.
- Since 1995, six additional monitoring wells have been installed at FTA-2 and sampled for groundwater quality by Tinker AFB.

Twenty six groundwater monitoring wells and monitoring points (2-62A, 2-62B, 2-63A, 2-63B, 2-64A, 2-64B, 2-65A, 2-65B, 2-142A, 2-142B, 2-167B, 2-168B, 2-271B, 2-272B, 2-273B, 2-274A, 2-274B, 2-301B, 2-355B, 2-356B, 2-392B, 2-393B, BCG39-B97-41, CG39-B97-43S, CG39-B97-43D, and CG39-B97-45) were utilized for the RNA TS (Figure 1.3). No soil or groundwater remediation has been performed at FTA-2.



SECTION 2

SITE CHARACTERIZATION ACTIVITIES

This section presents the methods used by Parsons ES personnel to collect site-specific data at FTA-2, Tinker AFB, Oklahoma. To meet the requirements of the RNA demonstration, additional data were required to evaluate near-surface geology and geochemistry, aquifer properties, and the extent of soil, surface water, and groundwater contamination. Site characterization activities involved using a hollow stem auger drill rig for soil sample collection and groundwater monitoring well and monitoring point installation. Groundwater samples were collected during this investigation from both newly installed and previously installed monitoring wells and points. Hydraulic conductivity (slug) tests were conducted at selected site monitoring wells and points. Previously collected data and data collected under this program were integrated to develop the conceptual hydrogeologic site model and to aid with interpretation of the physical setting (Section 3) and contaminant distribution (Section 4).

Sampling locations for the RNA TS are shown on Figure 1.3, and a summary of activities conducted during the RNA TS by location is provided in Table 2.1. The following subsections summarize the procedures followed when collecting site-specific data for this TS. Additional details regarding investigative activities are presented in the work plan (Parsons ES, 1997).

Two monitoring wells (2-392B and 2-393B), and four temporary well points (CG39-B97-41, CG39-B97-43s, CG39-B97-43d, and CG39-B97-45) were installed by Parsons ES during the field program. A summary of monitoring well and monitoring point construction is presented in Table 2.2. Boring logs, well construction diagrams, geophysical logs and well development forms for the six new monitoring wells and points are presented in Appendix B. In addition to the new RNA TS sampling locations, three monitoring wells (2-274A, 2-355B, and 2-356B) were installed by Brown and Root consultants (BR) at FTA-2. Construction details for the BR wells and all wells sampled as part of the TS are provided in Table 2.2.

2.1 SOIL SAMPLE COLLECTION AND GROUNDWATER MONITORING WELL/POINT INSTALLATION

2.1.1 Pre-Drilling Activities

All subsurface utility lines or other man-made subsurface features were located, and proposed borehole locations were cleared and approved by the Base prior to drilling activities.

SUMMARY OF TREATABILITY STUDY ACTIVITIES SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA TABLE 2.1

_			_	_	_	_	_	_	_			_	_		_	_		_	_	_	_	\neg	_			_	-			
		Phenols			×																									
	Cathy	Acids			×																									
	7042	Phenols		×	×	×	×	X	Х	X	X	X	Х	X	×	×	×	×	X	X		×		X	X	X	×	×	×	
		co,"		×	×	×	×	×	X	X	×	×	X	×	x	×	×	×	×	×		×		х	Х	X	X	×	×	trogen).
		Η,,		×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×		×		×	X	X	×	×	×	te (as ni
		SO4 v		×	×	×	×	×	X	×	X	x	×	×	×	×	×	×	X	×		x		×	×	×	×	×	×	te + nitri
	+_ O.V.	NO. 5		X	Х	×	×	×	×	×	×	×	×	×	X	X	×	×	X	×		X		×	×	×	×	×	×	h/ NO." + NO." = Nitrate + nitrite (as nitrogen)
lyses		NH3 F		X	X	×	×	×	×	×	×	×	×	×	×	×	X	×	×	×		×		×	×	×	×	×	×	0. + VO
ater Ana		CI.		X	X	X	X	×	×	X	×	×	×	×	Х	X	X	X	×	×		X		×	×	×	×	×	×	N
Groundwater Analyses	Methane,	etnane, and ethene		×	×	X	×	X	×	×	×	×	×	×	×	×	×	×	×	×		×		×	×	×	×	×	X	
	1000	Mobile Lab		×	×	×	x	×	×	×	×	×	×	×	×	×	×	×	×	×		×		×	×	×	×	×	×	
		well Head Analyses		×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×		×		×	×	×	×	×	×	
		TOC																												
	Fuel Hydro-	carbons and Fuel Carbon		×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×		×		×	×	×	×	×	×	
		CAHs		×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×	×		×	×	×	×	×	×	
alyses	TCE/	Analysis		×		×																		-						
Soil Analyses	, V	Analysis		×		×		×	×																					
	Slug	Analysis Analysis		×	×	×	×	×	×											×										
	Monitoring	Well/Point Installation		×	×	×	×	×	×														ne							ic carbon
		Location	Upper Saturated Zone	CG39-B97-41	CG39-B97-43s	CG39-B97-43d	CG39-B97-45	2-392B	2-393B	2-62B	2-63B	2-64B	2-65B	2-167B	2-168B	2-271B	2-272B	2-273B	2-274B	2-355B	2-356B ^m	2-142B	Lower Saturated Zone	2-62A	2-63A	2-64A	2-65A	2-142A	2-274A	2/ TOC = Total organic carbon

a/ TOC = Total organic carbon.

b/ TCE/PCE = Trichloroethylene and Tetrachloroethylene.

c/ CAHs = Chlorinated Aliphatic Hydrocarbons.

d/ Well head analyses include dissolved oxygen, reduction-oxidation potential, pH, temperature, and conductivity.

e/ Mobile lab analyses include total alkalinity, ferrous iron, manganese, carbon dioxide, total phenols, and hydrogen sulfide.

f/ CI' = Chloride. g/ NH₃ = Ammonia.

 $h' NO_2^{-1} + NO_3^{-1} = Nitrate + nitrite (as nitrogen).$ i/ $SO_4^{-2} = Sulfate.$

j/ H* = Dissolved Hydrogen.

k/ CO₂ = Carbon dioxide.

I/ X indicates that the analysis was completed.

m/ Monitoring well 2-356B was sampled by Brown and Root on September 4, 1997.

TABLE 2.2 MONITORING WELL SUMMARY AND AUGUST 1997 GROUNDWATER ELEVATIONS FTA-2 RNA TS TINKER AFB, OKLAHOMA

			Well	Screened	Elevation		Ground	Survey	Survey	Depth to Water	Elevation of Water
Well/Borehole	Site	Completion	Diameter	Interval	Datum	Elevation	Surface	Northing	Easting	August-1997	Table - August 1997
Identification	Code	Date	(Inches)	(Feet bgs) ^{a/}	(Feet msl) ^{b/}	Datum	Elevation	(State Plane)	(State Plane)	(Feet btoc)	(Feet msl)
New Monitoring Points and Monitoring Wells - Upper Saturated Zone	oints and M	fonitoring We	alls - Upper Sa	turated Zone							
CG39-B97-41	FT22	7/28/1997	2.0	34 - 39	1245.20	Top PVC	1245.57	150440.90	2182126.70	12.33	1232.87
CG39-B97-43s	FT22	7/30/1997	2.0	30 - 35	1252.42	Top PVC	1252.72	150861.52	2181770.08	19.90	1232.52
CG39-B97-43d	FT22	7/29/1997	2.0	44.5 - 49.5	1252.54	Top PVC	1252.56	150866.94	2181777.04	20.19	1232.35
CG39-B97-45	FT22	7/31/1997	2.0	36.5 - 41.5	1241.82	Top PVC	1242.24	151085.60	2181167.63	12.11	1229.71
2-392B	FT22	7/29/1997	4.0	36 - 46	1244.99	Top PVC	1245.30	150696.27	2181430.89	15.43	1229.56
2-393B	FT22	7/30/1997	4.0	35 - 45	1245.97	Top PVC	1246.19	151434.08	2181280.87	15.86	1230.11
Existing Monitoring Wells - Upper Saturated Zone	ng Wells - I	Jpper Saturate	ed Zone								
2-62B	FT22	11/9/1993	2.0	14-24	1245.94	Top PVC	1245.94	150444.52	2182140.56	14.51	1231.43
2-63B	FT22	11/8/1993	2.0	18-23	1243.28	Top PVC	1243.28	150386.11	2182035.36	12.06	1231.22
2-64B	FT22	11/9/1993	2.0	16-26	1245.59	Top PVC	1245.59	150467.40	2181967.76	14.11	1231.48
2-65B	FT22	11/5/1993	2.0	37-47	1250.81	Top PVC	1250.81	150712.88	2182183.20	17.10	1233.71
2-167B	FT22	8/8/1995	4.0	19-29	1250.70	Top PVC	1247.70	150539.81	2182408.35	16.03	1234.67
2-168B	FT22	8/8/1995	4.0	10-20	1243.00	Top PVC	1240.10	150249.00	2181976.69	12.55	1230.45
2-271B	FT22	8/8/1996	4.0	36-46	1252.13	Top PVC	1252.41	151494.42	2182196.43	18.35	1233.78
2-272B	FT22	8/9/1996	4.0	25-35	1248.91	Top PVC	1249.38	150784.70	2181581.74	17.79	1231.12
2-273B	FT22	6/18/1996	4.0	17-27	1238.42	Top PVC	1238.92	150351.11	2181576.36	8.26	1230.16
2-274B	FT22	7/25/1996	4.0	37-42	1248.32	Top PVC	1248.77	151126.23	2181625.92	16.67	1231.65
2-355B	FT22	7/9/1997	4.0	19.9-29.9	1245.92	Top PVC	1246.61	151233.46	2181400.66	15.73	1230.19
2-356B	FT22	7/10/1997	4.0	27.8-37.8	1249.30	Top PVC	1249.75	151462.29	2181548.00	18.24	1231.06
2-142B	BW	5/19/1995	4.0	32-42	1242.09	Top PVC	1242.45	NA	NA	13.31	1228.78
2-301B	LF15	9661/11/9	4.0	16-26	1244.00	Top PVC	1244.00	149701.00	2182586.00	10.14	1233.86
Existing Monitoring Wells - Lower Saturated Zone	ıg Wells - I	Lower Saturat	ed Zone								
2-62A	FT22	11/29/1993	2.0	54-64	1246.21	Top PVC	1246.21	150447.89	2182152.17	57.32	1188.89
2-63A	FT22	11/23/1993	2.0	53-63	1243.39	Top PVC	1243.39	150389.96	2182043.02	54.70	1188.69
2-64A	FT22	11/22/1993	2.0	99-95	1246.05	Top PVC	1246.05	150470.21	2181959.37	57.34	1188.71
2-65A	FT22	11/19/1993	2.0	92-99	1250.98	Top PVC	1250.98	150698.28	2182189.88	61.48	1189.50
2-274A	FT22	7/17/1997	4.0	76-81	1248.75	Top PVC	1249.10	151115.88	2181631.94	59.20	1189.55
2-142A	BW	5/19/1995	4.0	72-82	1241.76	Top PVC	1242.30	NA	NA	53.03	1188.73
al Front Landing	oitonolo zet	- in fact halo	The three in discount of a second of the fact that have been described			C East btoc	indicates denth	C East bloc indicates denth measured below ton of well casing	on of well casing		

^{3/} Feet bgs indicates elevation in feet below ground surface.
^{b/} Feet msl indicates elevation in feet above mean sea level.

 $^{^{\}rm cl}$ Feet btoc indicates depth measured below top of well casing. $^{\rm dl}$ NA indicates the data are not currently available.

2.1.2 Hollow Stem Auger Drill Rig Field Activities

Two new monitoring wells and four new monitoring points were installed in boreholes advanced by Associated Environmental Industries Corporation (AEI) using a Failing F10 hollow stem auger rig from July 28 through July 31, 1997. All boreholes for the proposed monitoring points/wells were drilled to the top of the USZ/LSZ aquitard. Geophysical logs (including gamma ray, spontaneous-potential, and resistivity) were run from the total depth of the borehole to the ground surface. Geophysical logs were a requirement of Base installation protocols and enhance resolution of lithologic contacts. Boring logs, monitoring well installation records, and geophysical logs for the two newly installed monitoring wells and the four newly installed monitoring points are provided in Appendix B.

2.1.2.1 Equipment Decontamination

Prior to arriving at the site and after each soil boring, draw works, augers, and push rods were cleaned with a steam-cleaning system. Water used in equipment cleaning was obtained from an on-Base potable water supply. Equipment cleaning was performed at a lined decontamination pad constructed of concrete at the Base former drying beds/sewage plant area. Rinseate drained from the pad was disposed of at the Base Groundwater Treatment Plant (GWTP).

2.1.2.2 Soil Sample Collection and Analysis

The hollow stem auger rig was used to collect soil samples for visual description, VOC headspace screening, VOC analyses, and total organic carbon (TOC) analysis during the installation of monitoring wells and points. The samples were collected from multiple depths using a decontaminated continuous California sampler.

At each soil borehole location, soil samples were collected and logged continuously to the total depth of the borehole. After visual examination, selected soil samples were transferred to analyte-appropriate jars, and submitted to NRMRL for TOC analysis using methods RSKSOP-102 and RSKSOP-120, or for TCE/ and tetrachloroethene (PCE) analyses using a modification of RSKSOP-124. A portion of the soil sample from each five-foot interval was placed in an unused, clean, glass jar for photoionization detector (PID) headspace measurements of ionizable VOCs. Soil remaining in the sampler was used for stratigraphic logging. Jars containing soil samples collected for the headspace screening procedure were quickly sealed with aluminum foil and stored for at least 15 minutes at the ambient temperature. Semiquantitative measurements were made by puncturing the jar seal with the PID probe and reading the concentration of the headspace gases. The PID relates the concentration of total VOCs in the sample to an isobutylene calibration standard. The PID was also used to monitor for VOCs in the worker breathing zone. The Parsons ES scientist maintained a descriptive log of subsurface materials recovered. Final geologic borehole logs with soil sampling intervals and PID screening are presented in Appendix B.

During monitoring well/point installation, the excess soils generated by drilling were temporarily stored in a portable bin. Soils were then transferred to an on-site 20 cubic yard rolloff bin after each well/point installation to await sampling, analysis, and

disposal. After laboratory analysis, the soils were approved for disposal at the Waste Management, Inc., East Oak Landfill.

2.1.2.3 Monitoring Well and Monitoring Point Installation

Using the hollow stem auger drill rig, Parsons ES and AEI staff installed two 4-inch inside diameter (ID) monitoring wells at 2-392B and 2-393B, and four 2-inch ID monitoring points at CG39-B97-41, CG39-B97-43s, CG39-B97-43d, and CG39-B97-45.

2.1.2.3.1 Materials

Monitoring wells were constructed of flush-threaded four-inch-ID stainless steel casing and screen. Polyvinyl chloride (PVC) casing was used above the water table to the surface completion. Installed screens were 10.0 feet in length and factory-slotted with 0.010-inch wire wrap openings. Monitoring points were constructed of flush-threaded two-inch-ID stainless steel casing and screen. PVC casing was used above the USZ water table to the surface completion. Monitoring point screens were 5.0 feet in length and factory-slotted with 0.010-inch wire wrap openings. Each monitoring well/point was fitted with a locking expandable cap. Well materials were inspected for cleanliness prior to use. No glue or solvents were used with monitoring well/point materials.

2.1.2.3.2 Installation

After geophysical logging, monitoring wells/points were installed in the open borehole. No difficulties related to borehole collapse were encountered. After the monitoring well screen and riser pipe were in place, the annular backfill materials were gradually added down the borehole. Monitoring wells/points were completed by filling the borehole around the well with 20-40 graded sand, bentonite chips and bentonite/cement grout. All monitoring wells/points were completed with surface flush mount protective casing in accordance with Base well installation standards. The tops of the wells/points were fitted with expandable, locking well caps. Monitoring well/point completion logs are included in Appendix B.

2.1.2.3.3 Development

Prior to sampling, newly installed monitoring wells and monitoring points were developed. Typically, well development removes sediment from inside the well casing and flushes fines, cuttings, and drilling fluids from the sand pack and the portion of the formation adjacent to the well screen. Monitoring well/point development was accomplished using a Grunfos® submersible pump with decontaminated high-density polyethylene (HDPE) tubing. The pump was lowered to the bottom of the well/point and surged so that fines were agitated and removed from the well/point in the development water. Development continued until at least ten casing volumes of water were removed from the monitoring well/point, the temperature, pH, oxidation-reduction potential (ORP), conductivity, and dissolved oxygen (DO) concentrations of the extracted groundwater had stabilized, and turbidity stabilized at less than 10 NTU, but with a limit of 300 gallons total developed water for individual monitoring points. Monitoring well/point development records are included in Appendix B. All development water was

contained in 500 to 1,000 gallon portable tanks and was disposed at the Base GWTP by AEI.

2.2 GROUNDWATER SAMPLING

The groundwater sampling procedures described in the site work plan (Parsons ES, 1997) and summarized in the following sections were followed. Groundwater sampling took place from July 31 to August 4, 1997, and consisted of collecting groundwater samples from sixteen previously installed wells (2-62A, 2-62B, 2-63A, 2-63B, 2-64A, 2-64B, 2-65A, 2-65B, 2-142A, 2-142B, 2-167B, 2-168B, 2-271B, 2-272B, 2-273B and 2-274B), two monitoring wells recently installed by BR (2-274A and 2-355B), four newly installed monitoring points (CG39B97-41d, CG39B97-43s, CG39B97-43d, and CG39B97-45), and two newly installed monitoring wells (2-393B and 2-393B). Refer to Table 2.1 for analyses conducted at each monitoring well and point.

2.2.1 Preparation and Equipment Cleaning

All equipment used for sampling was assembled, properly cleaned, and calibrated (if required) prior to use in the field. All portions of sampling and test equipment that contacted the sample were thoroughly cleaned before use. This equipment included the water level probe and cable and equipment for measuring onsite groundwater chemical parameters. The equipment was cleaned using the following protocol:

- · Rinsed with isopropyl alcohol;
- · Rinsed with distilled or deionized water; and
- · Air dried prior to use.

Special care was taken to prevent cross contamination of the groundwater and extracted samples through contact with improperly cleaned equipment. In addition, a clean pair of new, disposable latex gloves was worn each time a different well or monitoring point was sampled. Any deviations from these procedures were documented in the field scientist's field notebook and on the groundwater sampling form. As required, field analytical equipment was calibrated according to the manufacturers' specifications prior to field use.

Prior to starting the sampling procedure, the area around the well or monitoring point was cleared of foreign materials, such as brush, rocks, and debris. These procedures prevented sampling equipment from inadvertently contacting debris around the monitoring wells and points. The integrity of the monitoring wells and points was also inspected, and any irregularities in the visible portions of the wells and points, protective covers, or concrete pads were noted.

2.2.2 Water Level and Total Depth Measurements

Prior to removing any water from the monitoring well or point, the static water level was measured. An electrical water level probe was used to measure the depth to groundwater below the well datum to the nearest 0.01 foot. After measurement of the

static water level, the water level probe was lowered to the bottom of the newly installed wells and points for measurement of total depth (recorded to the nearest 0.1 foot). The saturated casing volume for each well was calculated based on these measurements, or using total depths recorded on well construction.

2.2.3 Monitoring Well/Point Purging

Prior to sampling, each monitoring well or point was purged to remove stagnant water from the well casing. Where possible, the volume of water removed from each well/point was at least three times the calculated saturated casing volume. Where it was not possible to remove three times the calculated casing volume (i.e., the well/point was purged dry), the well/point was allowed to recharge until sufficient water was present to obtain the necessary sample quantity. Purging continued until the pH, DO concentration, ORP, conductivity, and temperature stabilized. Dedicated Grunfos® submersible pumps were used for existing wells. A decontaminated Grunfos® pump and tubing were used for newly or recently installed monitoring wells. All purge water from monitoring points and wells was collected in five-gallon buckets, transferred to a 500- or 1,000-gallon portable tank, and disposed at the Base GWTP.

2.2.4 Onsite Chemical Parameter Measurement

Groundwater samples were analyzed in the field by Parsons ES and USEPA personnel for pH, conductivity, temperature, DO, ORP, total alkalinity, ferrous iron, dissolved hydrogen, and hydrogen sulfide. Analyses for BTEX, trimethylbenzenes (TMBs), fuel carbon, CAHs, nitrate and nitrite nitrogen, ammonia nitrogen, chloride, sulfate, methane, ethane, ethene, dissolved TOC, and aliphatic and aromatic (fatty) acids were performed at the NRMRL in Ada, Oklahoma. DO measurements were taken using either an Orion® Model 840 or a YSI® Model 55 DO meter in a flow-through cell at the outlet of the purge pump. DO concentrations were recorded after the readings stabilized, and in all cases represent the lowest DO concentration observed.

Because the electrical conductivity, ORP, and temperature of the groundwater change significantly within a short time following sample acquisition, these parameters were measured in the field, in the same flow-through cell used for DO measurements. Measured values were recorded on the groundwater sampling records (Appendix B). Other time sensitive parameters, including pH, ferrous iron, alkalinity, and hydrogen sulfide, were analyzed at the USEPA mobile laboratory immediately after sample collection.

2.2.5 Sample Collection

Dedicated or decontaminated Grunfos® pumps and tubing were used to extract groundwater samples from monitoring points and wells. The samples were transferred directly into the appropriate sample containers. The water was carefully poured down the inner walls of each sample bottle to minimize aeration of the sample. Sample bottles for BTEX, TMB isomers, total fuel carbon, CAHs, methane, ethane, ethene, and volatile fatty acids were filled so that there was no headspace or air bubbles within the container. Analytical methods are summarized in Table 2.3, and groundwater sampling records are provided in Appendix B.

2.3 SAMPLE HANDLING

The fixed-base analytical laboratory, NRMRL, provided pre-preserved sample containers where appropriate. The sample containers were filled as described in the previous sections. The samples were labeled as described in the work plan. After the samples were sealed and labeled, they were transported to the onsite USEPA mobile laboratory. Sample shipment to NRMRL and the associated chain-of-custody documentation was the responsibility of NRMRL field personnel.

2.4 AQUIFER TESTING

Slug tests were performed at FTA-2 by Parsons ES on August 5, 1997. Falling and rising head slug tests were performed at monitoring wells 2-355B, 2-393B, and 2-393B, and monitoring points CG39-B97-41, CG39-B97-43s, CG39-B97-43d, and CG39-B97-45. Parsons ES personnel used testing procedures similar to those described in the work plan (Parsons ES, 1997). Slug test data were analyzed using the AQTESOLV® software package (Geraghty and Miller, Inc., 1994) and the methods of Bouwer and Rice (1976) and Bouwer (1989) for unconfined conditions. Slug test results are presented in Section 3.3 and Appendix B of this document.

2.5 SURVEYING

After completion of field work, the locations and elevations of all new monitoring points and wells were surveyed by Lemke Land Surveying, an Oklahoma licensed land surveyor from Norman, Oklahoma. The horizontal locations and elevations of the measurement datum (i.e. top of PVC well casing) and the ground surface adjacent to the well casings were measured relative to existing control points referenced to the NAD 27 (horizontal) and National Geodetic Vertical Datum (NGVD) 29 (vertical) plane coordinate systems. Horizontal locations were surveyed to the nearest 0.1 foot. Measurement datum and ground surface elevations were surveyed to the nearest 0.01 foot. Survey data are presented in Appendix B.

TABLE 2.3 ANALYTICAL PROTOCOLS FOR GROUNDWATER AND SOIL SAMPLES FTA-2 RNA TS TINKER AFB, OKLAHOMA

	ANALYTICAL LABORATORY (L)
	LABORATORY (L)
reading meter	
reading meter	
	Ţ
Direct-reading meter	Ľ
Direct-reading meter	Ľ
Direct-reading meter	ഥ
Direct reading meter	Ľ
Titrimetric, Chemetrics Method	ſĽ,
Titrimetric, Chemetrics Method	Ľ
Titrimetric, Hach Method 8221	ĹŤ
Titrimetric, Hach Method 1436-01	Ľ.
Chemetrics Method	Ľ
Titrimetric, Chemetrics Method	ír.
	J
Canillon, Electronhoronic Mothed N. 601	1
Capillary Electrophoresis Method N 601	J
Capillary Electrophoresis Method IN-001	٠, د
OP-194" /RSKSOP-175	۱,
DP-102	J
RSKSOP-133	Г
RSKSOP-148	J
RSKSOP-177	J
RSKSOP-102 & RSKSOP-120 RSKSOP-124, modified	חח
Titrimetric, His Chemetrics M Titrimetric, Chemetrics M Titrimetric, Chemetrics M Waters Capillis Waters Capillis Waters Capillis Waters Capillis RSKSOP-102 RSKSOP-1133 RSKSOP-1148 RSKSOP-177 RSKSOP-102 RSKSOP-124 RSKSOP-124	Intrimetric, Hach Method 1430-01 Chemetrics Method Titrimetric, Chemetrics Method E350.1 Waters Capillary Electrophoresis Method N-601 Waters Capillary Electrophoresis Method N-601 RSKSOP-194"/RSKSOP-175 RSKSOP-102 RSKSOP-1133 RSKSOP-177 RSKSOP-177 RSKSOP-102 & RSKSOP-120 RSKSOP-124, modified

^{3/}RSKSOP = Robert S. Kerr Laboratory standard operating procedure.

SECTION 3

PHYSICAL CHARACTERISTICS OF THE STUDY AREA

Existing site-specific data were reviewed and supplemented with data collected by Parsons ES in July and August 1997 to develop a synopsis of FTA-2 physical characteristics. In addition to field investigation results from this TS, data from the following sources are included:

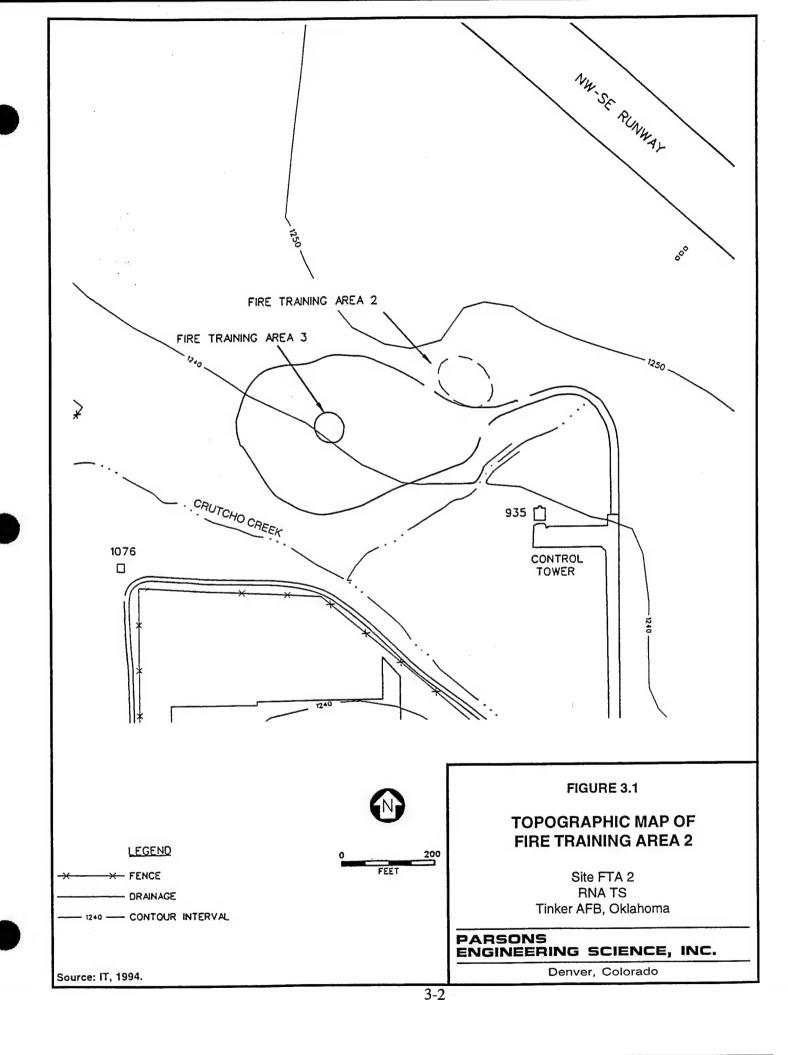
- Final Report, Phase I RCRA Facility Investigation Report for Appendix I Sites, SWMU-8, Fire Training Area 2, (IT, 1994); and
- Draft Report, RCRA Facility Investigation Report for Selected Appendix I Sites at Tinker Air Force Base, OK (TT, 1996).

3.1 TOPOGRAPHY, SURFACE HYDROLOGY, AND CLIMATE

Tinker AFB is located in central Oklahoma and topography varies from almost level to gently rolling. The local relief is a result of dissection by erosional activity and stream channel development. Ground surface elevations at Tinker AFB range from 1,190 feet above the NGVD near the northwest corner where Crutcho Creek intersects the Base boundary, to 1,320 feet NGVD at Area D, located east of the main installation. A topographic map of FTA-2 at Tinker AFB is presented as Figure 3.1.

Surface water runoff from the Base primarily drains to diversion structures and then to intermittently flowing surface streams. The north and west portions of the Base drain to Crutcho Creek, a tributary of the North Canadian River. Kuhlman Creek also collects drainage from the northwest portion of the Base and discharges into Crutcho Creek north of the Base. The northeast portion of the Base is drained primarily by tributaries of Soldier Creek, which is also a tributary of Crutcho Creek. Two small unnamed intermittent tributaries of Elm Creek cross installation boundaries south of the main runway, and generally do not receive significant quantities of Base runoff because site grading is designed to preclude such drainage. These streams, when flowing, extend to Stanley Draper Lake, approximately one-half mile south of the Base.

In the area of FTA-2, topography slopes gently to the southwest, towards Crutcho Creek, with elevations ranging from 1,240 to 1,250 feet NGVD. The sloping topography is modified by an incised channel of a southwest-flowing tributary to Crutcho Creek. The channel of this tributary is approximately 5 to 10 feet below the grade of the surrounding terrain. Surface water run-off at FTA-2 travels southeast to this southwest-draining tributary. The tributary emerges from a culvert located



approximately 200 feet east of FTA-2, after draining an area occupied by industrial facilities on the east side of the airfield.

The climate in central Oklahoma is continental, and is characterized by cold winters, hot summers, and moderate rainfall. Precipitation averages 33 inches per year. Annually, potential evaporation usually exceeds precipitation. Maximum evaporation occurs during June, July, and August.

3.2 REGIONAL GEOLOGY AND HYDROGEOLOGY

3.2.1 Regional Geology

Tinker AFB is located within the Central Redbed Plain Section of the Central Lowland Physiographic Province, which is tectonically stable. No major faults or fracture zones have been mapped near Tinker AFB. Tinker AFB is underlain by several thousand feet of sedimentary strata that range in age from Cambrian to Permian and overlie a Precambrian igneous basement. The major lithologic units in the area of Tinker AFB are relatively flat-lying and have a regional westward dip of approximately 40 feet per mile (ft/mile) (Bingham and Moore, 1975).

Geologic units that outcrop at Tinker AFB consist of, in descending order, Quaternary Alluvium, the Hennessey Group, the Garber Sandstone, and the Wellington Formation (Table 3.1, Figure 3.2). Quaternary alluvium and terrace deposits overlie bedrock in and near present-day stream valleys. Quaternary deposits consist of unconsolidated soils from weathered bedrock, eolian sands, and interfingering lenses of fluvial sands, silts, clays and gravels. Terrace deposits are exposed where stream valleys downcut through older strata, leaving them topographically higher than present day valley floors. Alluvial sediments range in thickness from less than a foot to nearly 20 feet.

The Hennessey Group, Garber Sandstone, and Wellington Formation are Permian in age (230 to 280 million years ago), and consist of a conformable sequence of sands, silts, and clays with lenticular beds that vary in thickness over short horizontal distances (IT, 1994). Because of similar lithologies and lack of fossils or key beds, the Garber Sandstone and the Wellington Formation are difficult to distinguish and are informally grouped as the Garber-Wellington Formation. Together these two units are approximately 900 feet thick beneath Tinker AFB.

The Hennessey Group outcrops on the central, southern, and western portions of the Base, generally to the west and south of Crutcho Creek (Figure 3.2). The Hennessey Group includes the Kingman Siltstone and the Fairmont Shale (Miser *et al.*, 1954; Bingham and Moore, 1975), which are composed of red shale and thin beds of fine-grained sandstone. The Hennessey Group thins from approximately 70 feet in the southwest part of the Base towards its erosional edge across the northeastern part of the Base. The conformable contact between the Hennessey group and the underlying Garber Sandstone is often difficult to distinguish.

The Garber Sandstone outcrops across the northern, central, and eastern portions of the Base, and is generally covered by a thin veneer of soil or alluvium up to 20 feet

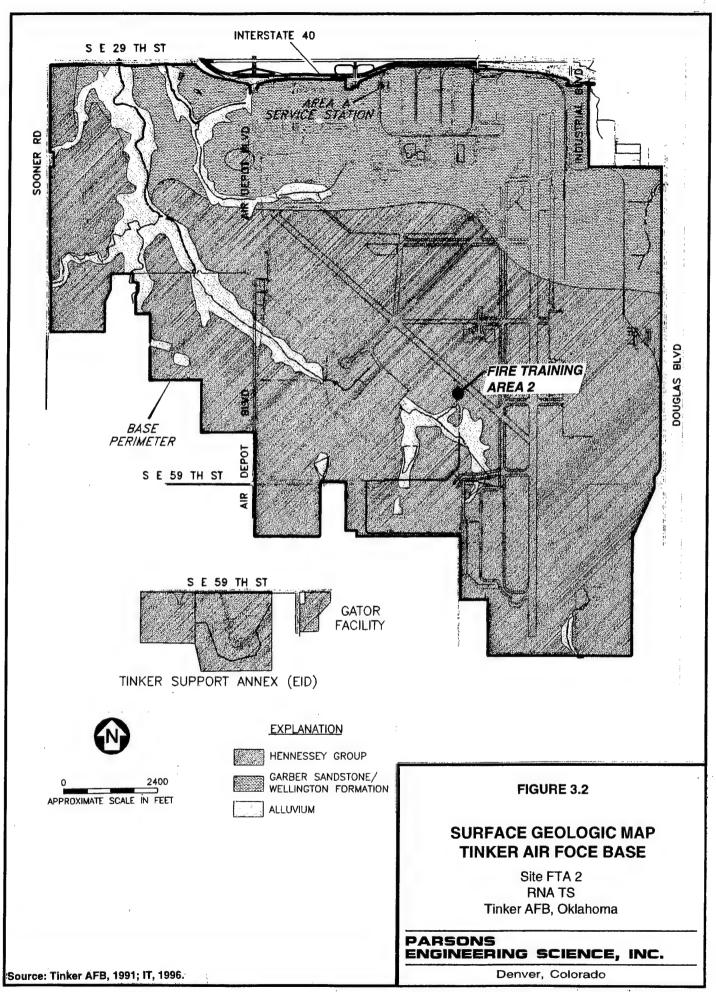
TABLE 3.1
MAJOR GEOLOGIC UNITS IN THE VICINITY OF TINKER AFB
FTA 2 RNA TS
TINKER AFB, OKLAHOMA

Water-Bearing Properties	Moderately permeable. Yields small to moderate quantities of water in	valleys of larger streams. Water is	very hard, but suitable for most uses,	unless contaminated by industrial	wastes or oil field brines							Moderately permeable. Where	deposits have sufficient saturated	thickness, they are capable of	yielding moderate quantities of water	to wells. Water is moderately hard	to very hard, but less mineralized	than water in other aquifers.	Suitable for most uses unless	contaminated by oil field brines.	
Description and Distribution	Unconsolidated, with interfingering lenses of sand, silt, clay, and gravel in flood	plains and stream channels.										Unconsolidated, with interfingering lenses	of sand, silt, gravel, and clay that occur at	one or more levels above flood plains and	principal streams.						
Thickness (feet)	0-70											0-100									
Stratigraphic unit	Alluvium											Terrace	deposits								
Series	民田	C	田	Z	П		A	Z	D		Ь	Γ	田	Н	S	L	0	ر ن	山	Z	Ε
System						0	Ω	A	L	田	R	Z	A	Υ							

TABLE 3.1 (Continued) MAJOR GEOLOGIC UNITS IN THE VICINITY OF TINKER AFB FTA 2 RNA TS TINKER AFB, OKLAHOMA

Water-Bearing Properties Poorly permeable. Yields meager quantities or very hard, moderately to highly mineralized water to shallow domestic and stock wells. In places,	sulfate.	Poorly to moderately permeable. Important source of groundwater in Cleveland and Oklahoma counties. Yields small to moderate quantities	of water to deep wells; heavily pumped for industrial and municipal uses in the Norman and Midwest City areas. Water from shallow wells hard to very hard; water from deep wells moderately hard to soft. Lower part contains water too salty for domestic and most industrial uses.
Deep-red clay shale containing thin beds of red sandstone and white or greenish bands of sandy or limey shale. Forms relatively flat to gently rolling grass-covered prairie.		Deep-red to reddish-orange sandstone, massive and cross-bedded and interfingered with red shale and siltstone.	Deep-red to reddish-orange massive and cross-bedded fine-grained sandstone interbedded with red, purple, maroon, and gray shale. Base of formation not exposed in the area.
Thickness (feet) 700		500±	200∓
Stratigraphic unit Hennessey Group (includes Kingmen Siltstone and	rannont Snate)	Garber Sandstone	Wellington Formation
Series L O W E	к ФШКХ	- 4 Z	
System P E E	X Z H K Z		

Source: Modified from Wood and Burton, 1968.



thick (Figure 3.2). The Garber Sandstone consists predominantly of fine-grained sandstone, with lesser amounts of siltstone and shale. Outcrops north of the base are characterized by small to medium channels with cross-bedded sandstones featuring cut and fill structures. The Wellington Formation, underlying the Garber Sandstone, outcrops to the east of Tinker AFB. Geophysical and lithologic logs of soil boreholes drilled on Base indicate that 65 to 70 percent of the Garber Sandstone and Wellington Formation are composed of sandstone. The sandstones are typically fine to very fine grained, friable, and poorly cemented. However, the sandstone intervals are locally cemented, typically at the base of sandstone lenses, by quartz, carbonate or iron-bearing minerals. These cemented intervals form horizons resistant to drilling. Shale intervals are generally discontinuous, and range in thickness from a few inches to 40 feet.

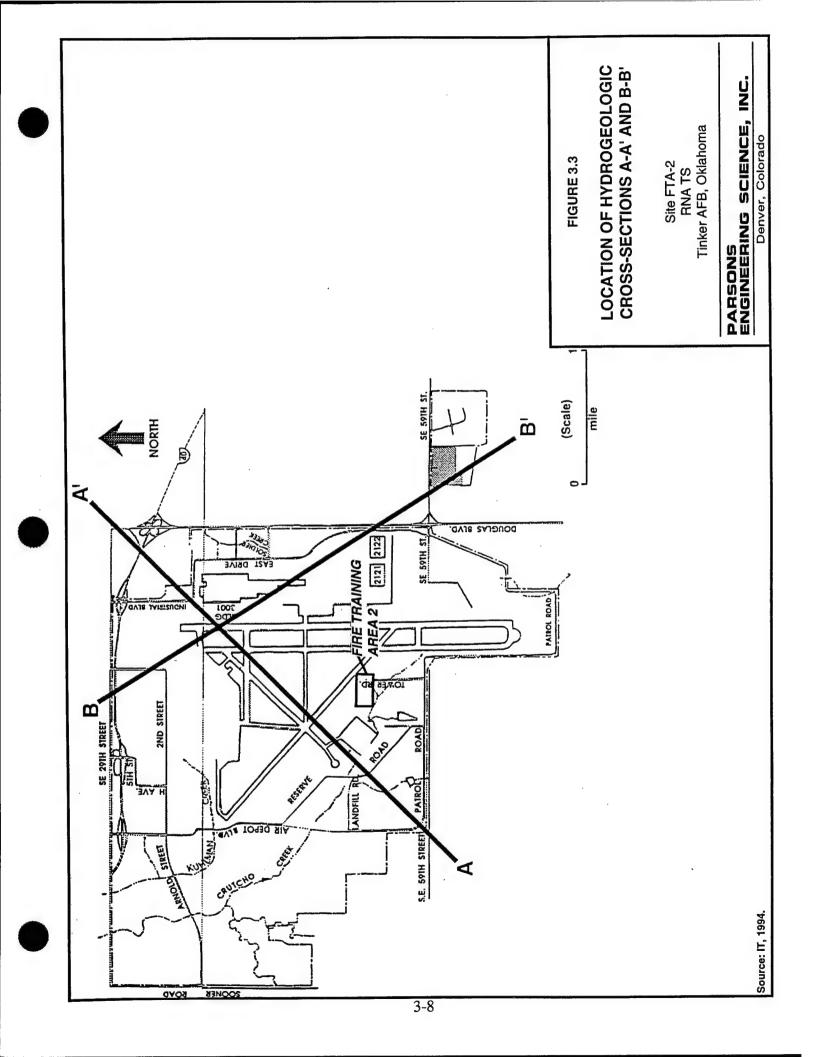
3.2.2 Regional Hydrogeology

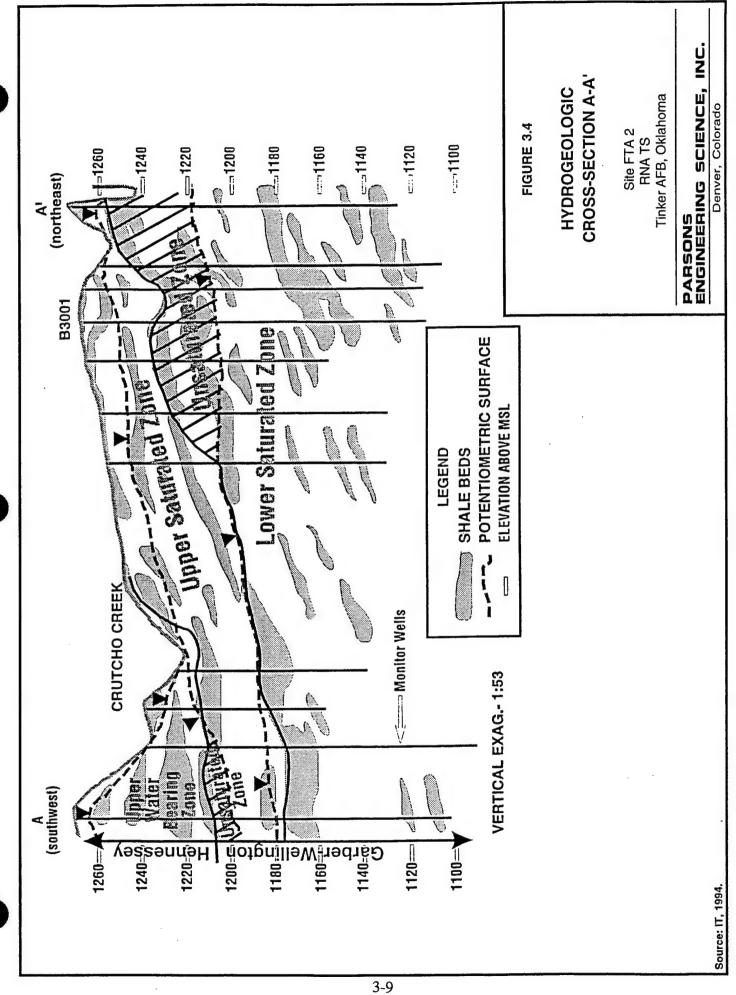
Tinker AFB overlies a regional source of potable water, the Central Oklahoma aquifer system. The productive formations of the aquifer include the Permian redbeds, including parts of the Permian Garber Sandstone and Wellington Formation, and Quaternary alluvium and terrace deposits. The Central Oklahoma aquifer has been classified as a Class IA aquifer by the State of Oklahoma, signifying that it is an irreplaceable source of public water supply (State of Oklahoma Water Resources Board, 1994). Water from the Central Oklahoma aquifer is used for municipal, industrial, domestic and agricultural purposes. Tinker AFB presently derives most of its water from a system of 26 operable water wells constructed generally along the east and west margins of the base, and from the Oklahoma City Water Department. All Base wells are completed in the Garber-Wellington Formation at depths of 400 to 1,100 feet.

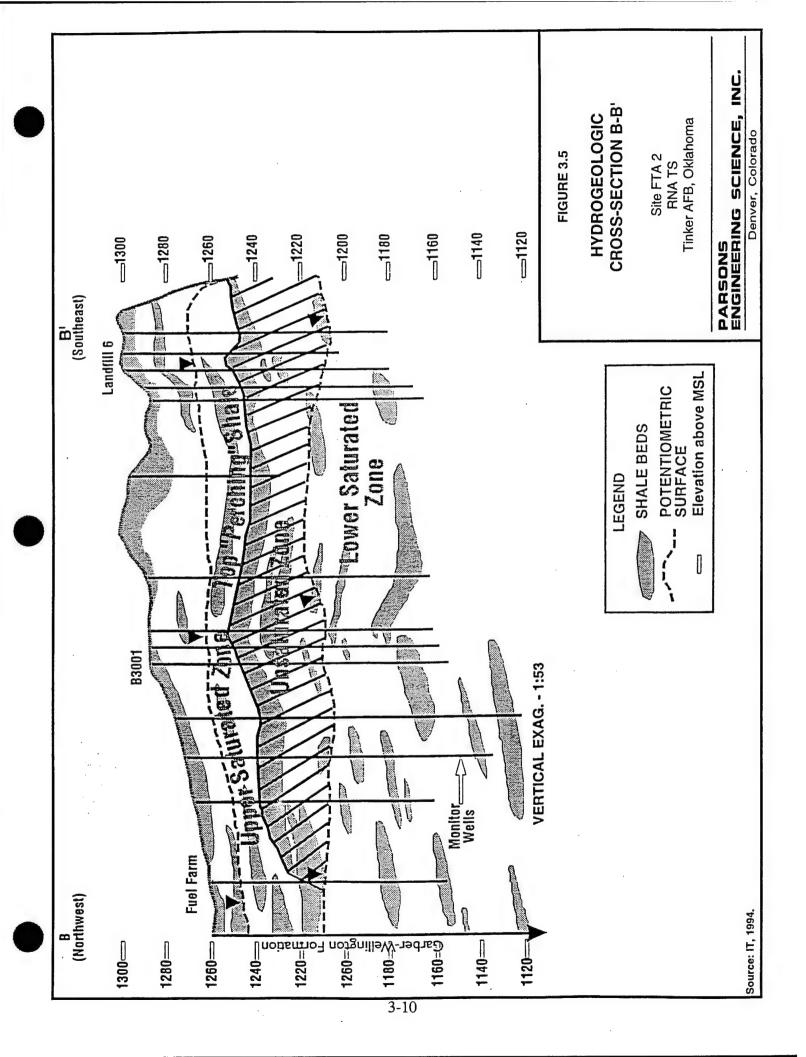
Groundwater in the Central Oklahoma aquifer is derived primarily from precipitation infiltrating outcrops of the geologic units. Infiltration of surface waters from streams crossing the outcrops is believed to be a minor source of recharge to the aquifer (Parkhurst *et al.*, 1993). Tinker AFB is located in the outcrop area of the Garber Sandstone and is therefore in the recharge zone of the aquifer.

The groundwater system at Tinker AFB has been divided into four hydrogeologic zones: the Hennessey Water-Bearing Zone (HWBZ), the USZ and LSZ, and the Producing Zone of the regional Garber-Wellington aquifer. The HWBZ occurs within the Hennessey Group and is perched above the USZ over the southwest portion of the Base. The LSZ and USZ are regionally considered to be in the upper third of the Garber-Wellington aquifer and generally are present at depths of less than 200 feet bgs. The Producing Zone generally is considered to be greater than 200 feet bgs, and is used for water supply at Tinker AFB.

The locations of regional hydrogeologic cross-sections A-A' (Figure 3.4) and B-B' (Figure 3.5) are shown on Figure 3.3. The HWBZ is perched within the Hennessey Group on the southwestern portion of the Base (Figure 3.4). The USZ generally behaves as a water table aquifer in the eastern part of the Base, but could be confined in localized portions of the Base where intra-USZ clay lenses intersect the USZ piezometric surface. The depth to the USZ water table ranges from zero feet bgs northeast and east of the Base to approximately 40 feet bgs in the southwest part of the







Base. The regional stratigraphic dip of 0.0076 feet per foot (ft/ft) to the west-southwest produces a general westward groundwater flow pattern in the USZ. Hydraulic gradients in the USZ across Tinker AFB range from 0.0034 to 0.018 ft/ft and average 0.0076 ft/ft. A low-permeability zone, characterized by high clay content and well-cemented silts and sands in the USZ, acts as an aquitard between the USZ and the LSZ.

The LSZ is considered one hydraulic unit from the USZ/LSZ aquitard to an approximate depth of 200 feet bgs. Due to variations in topography, the top of the LSZ is found at depths of 10 to 100 feet bgs. The LSZ extends east of the Base beyond the limits of the USZ and is the shallowest groundwater zone screened in off-Base wells. Across the northern and eastern portions of the Base an unsaturated zone is present between the USZ/LSZ aquitard and the top of the LSZ potentiometric surface, indicating poor hydraulic connection between the USZ and LSZ in this area. Across the central and southwestern portion of the Base, the unsaturated zone at the top of the LSZ disappears where the LSZ potentiometric surface rises above the overlying aquitard, and the LSZ is confined. A low-permeability zone at the base of the LSZ (characterized by high clay content) acts as an aquitard between the LSZ and the Producing Zone. The Producing zone extends from 200 to greater than 1,100 feet bgs and is under confined conditions.

3.3 FTA-2 GEOLOGY AND HYDROGEOLOGY

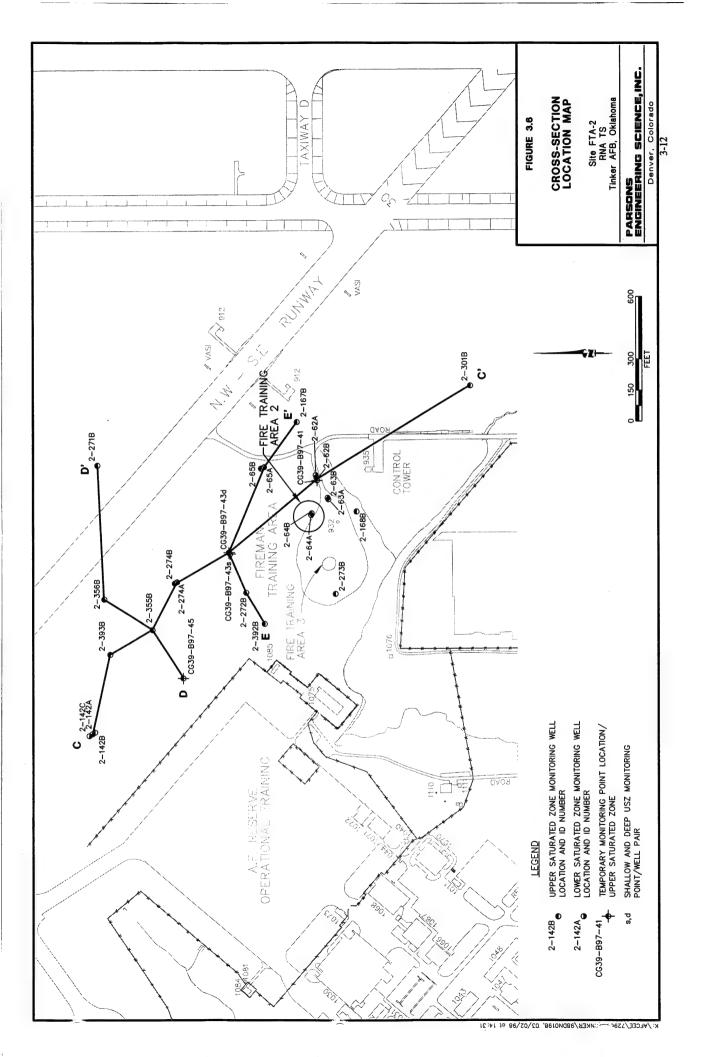
3.3.1 Site Geology

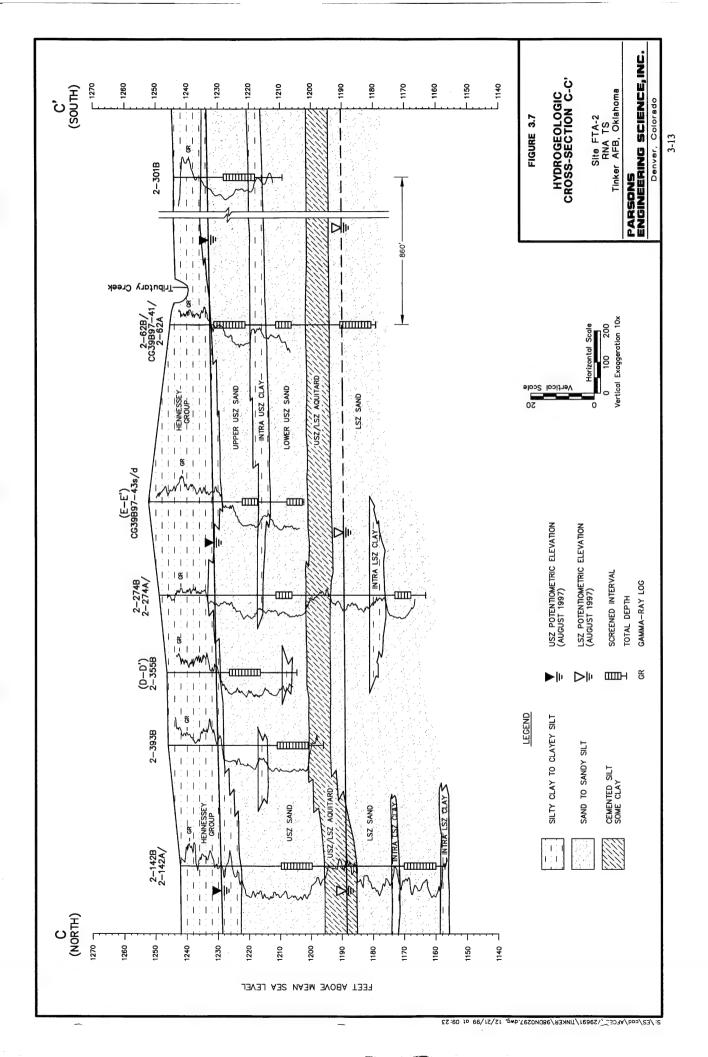
Stratigraphy through the center of the site is depicted in three hydrogeologic cross-sections through FTA-2 (Figure 3.6). Cross-section C-C' (Figure 3.7) is drawn in the approximate direction of groundwater flow, while Cross-sections D-D' (Figure 3.8), and E-E' (Figure 3.9) illustrate the stratigraphy perpendicular to Cross-section C-C' and the approximate direction of contaminant transport. Approximately 10 to 20 feet of reddish silty clay or clayey silt of the Hennessey Group overlies sands of the Garber Sandstone. The Garber Sandstone consists of fine-grained sands with silt and clay lenses. The USZ-LSZ aquitard occurs approximately 40 to 50 feet bgs and consists of a 6- to 12-foot thick interval with high clay content and cemented siltstone (Figure 3.7).

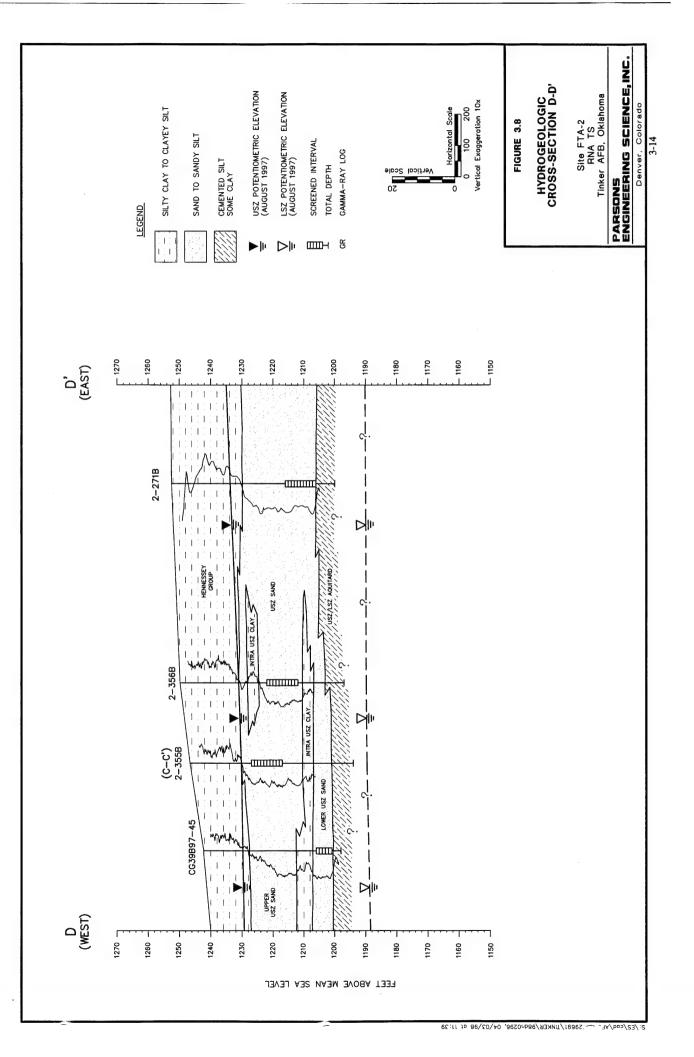
The USZ sands are primarily unconsolidated, with local lenses and intervals that are poorly to moderately cemented with silica or calcareous cement. In the southern and eastern portion of the study area a three- to six-foot thick interval of silt and clay separates the USZ into two separable sandstone intervals, herein referred to as the upper and lower USZ sand intervals (Cross-sections C-C' and E-E'). This intermediate clay and silt layer extends north and westward to well locations CG39-B97-43s,d and CG39-B97-45. To the north of these locations the intermediate clay and silt layer is discontinuous and less developed (Cross-section C-C' and D-D'). There are insufficient data to determine the extent or continuity of the intermediate clay and silt layer to the west of location CG39-B97-45.

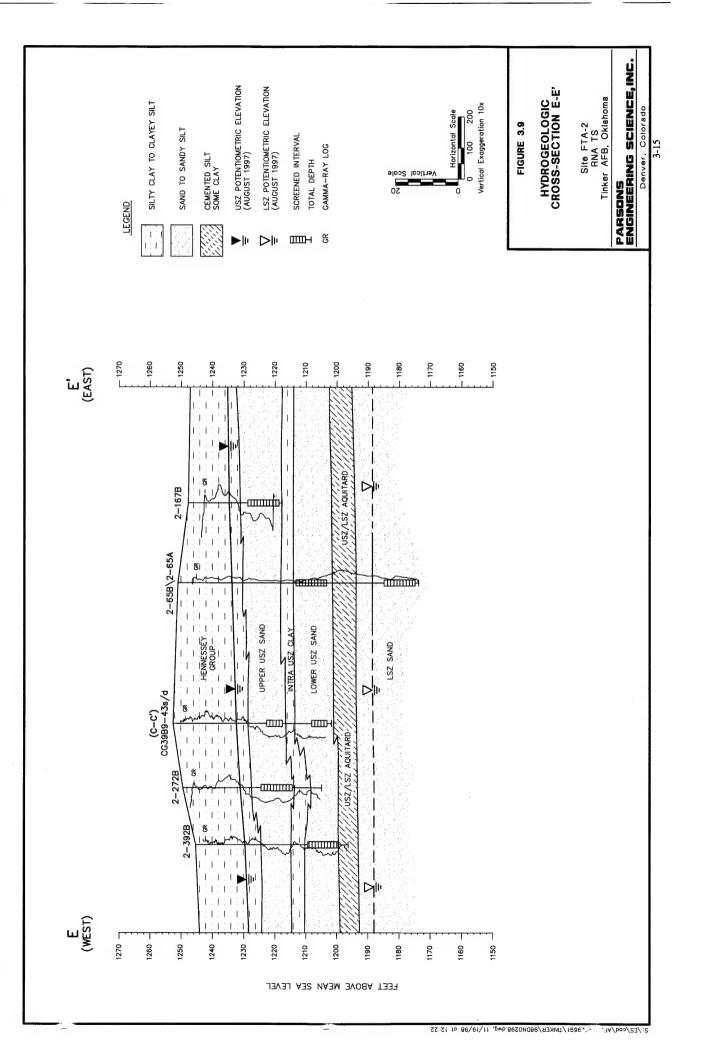
3.3.2 Site Hydrogeology

The hydrogeology at FTA-2 is characterized by groundwater flow within the USZ and the LSZ as described for regional conditions. Groundwater elevations and









monitoring well completion data are listed in Table 2.2. Geologic boring logs, monitoring well completion records, geophysical logs, and slug test data for newly installed monitoring well/points at FTA-2 are included in Appendix B.

3.3.2.1 Upper Saturated Zone

The USZ potentiometric surface generally lies above the Hennessey Group/Garber Sandstone (USZ sands) contact approximately 9 to 18 feet bgs. Figure 3.10 is a groundwater potentiometric surface map for the USZ based on data from August 1997. The horizontal hydraulic gradient in the USZ is approximately 0.003 to 0.01 ft/ft. Assuming isotropic horizontal hydraulic conductivity, horizontal flow direction in the vicinity of FTA-2 is generally towards the west-southwest. However, based on contaminant transport direction, horizontal hydraulic conductivity in the USZ may be anisotropic, with the primary direction of contaminant transport to the northwest (see Section 4.2 for further discussion of contaminant transport). The direction of maximum horizontal hydraulic gradient in the USZ for 1997 is consistent with gradient data observed in 1994 (IT, 1994), and 1996 (Parsons ES, 1997). The base of the USZ is the USZ/LSZ aquitard, a layer of hard siltstone traceable across the area at a depth of approximately 40 to 50 feet bgs. Because the underlying LSZ exists primarily as an unconfined aquifer (following Section 3.3.2.2), the USZ in the vicinity of FTA-2 is primarily a perched aquifer.

Vertical hydraulic gradients between the upper and lower USZ sand intervals were calculated for monitoring well pair CG39-B97-43s and CG39-B97-43d, and well pair 2-62B and CG39-B97-41. The vertical gradients were calculated by dividing the water level elevation difference between two clustered wells by the distance between the midpoints of the saturated portions of the well screens. The vertical groundwater gradient at well pair CG39-B97-43s and CG39-B97-43d was 0.012 ft/ft in a downward direction, typical of an unconfined/perched aquifer. The vertical groundwater gradient at well pair 2-62B and CG39-B97-41 was 0.081 ft/ft in an upward direction, indicating semi-confined or confined conditions. While the USZ is regionally considered an unconfined aquifer, semi-confined to confined conditions may exist where the potentiometric surface intercepts overlying fine-grained deposits. At FTA-2 an upward USZ groundwater gradient within the source area at well pair 2-62B and CG39-B97-41 may impact source area contaminant transport.

Hydraulic conductivity in the USZ was estimated by performing slug tests in seven newly installed and developed monitoring wells (CG39-B97-41, CG39-B97-43s, CG39-B97-43d, CG39-B97-45, 2-392B, 2-393B, and 3-355B). Hydraulic conductivity results are shown in Table 3.2. Results are segregated to compare the upper USZ sand interval and the lower USZ sand interval. The estimated hydraulic conductivity values for the upper USZ sand interval ranged from 6.5 to 28 feet per day (ft/day), and averaged 14 ft/day. The estimated hydraulic conductivity values for the lower USZ sand interval ranged from 0.9 to 44 ft/day, and averaged 15 ft/day. While the lower USZ sand interval demonstrated a greater range of hydraulic conductivities, overall values for the upper and lower USZ sand intervals are similar, corresponding to accepted literature values for fine-grained unconsolidated sand (Spitz and Moreno, 1996).

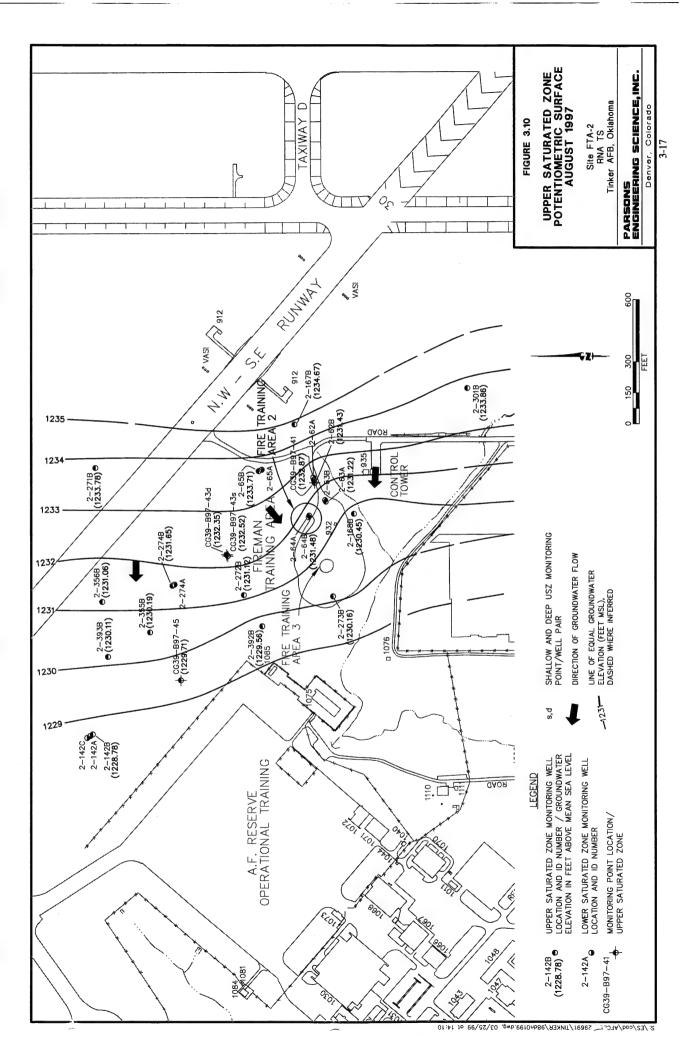


TABLE 3.2 UPPER SATURATED ZONE HYDRAULIC CONDUCTIVITIES AND AVERAGE GROUNDWATER VELOCITIES FTA-2 RNA TS

TINKER AFB, OKLAHOMA

		Screened Interval		Hydra	Hydraulic Conductivity	ivity	Estimated	Hydraulic	Average
Monitoring	Test	Elevation	Test	•	(K)		Effective	Gradient	Groundwater
Well	Date	(ft bgs) ^{a/}	Number	(ft/day) ^{b/}	(cm/sec)	(gpd/ft²) ^{d/}	Porosity	(ft/ft) ^{e/}	Velocity (ft/yr) [#]
Upper USZ Sand Interval	1 Interval								
CG39-B97-43S	Aug-97	30.0-35.0	_	18	6.4E-03	135	0.20	900'0	198
)		2	28	1.0E-02	212	0.20	900.0	311
			n	16	5.5E-03	117	0.20	0.006	171
			Average	21	7.3E-03	155	0.20	900.0	227
2-355B	Aug-97	19.9-29.9	_	7.7	2.7E-03	57	0.20	900.0	84
)		2	6.5	2.3E-03	49	0.20	900.0	72
			Average	7.1	2.5E-03	53	0.20	900.0	78
	Average of Upper	of Upper USZ Sand Interval:	Interval:	14	4.91E-3	104	0.20	9000	152
Lower USZ Sand Interval	d Interval								
CG39-B97-41	Aug-97	34.0-39.0		8.8	3.1E-03	99	0.20	900'0	62
)		2	12	4.2E-03	89	0.20	900.0	131
			Average	10	3.7E-03	78	0.20	900.0	114
CG39-B97-43D	Aug-97	44.5-49.5	-	27	9.5E-03	202	0.20	900'0	295
)		2	36	1.3E-02	272	0.20	0.006	398
			Average	32	1.1E-02	237	0.20	0.006	347
CG39-B97-45	Aug-97	36.5-41.5	1	44	1.5E-02	326	0.20	900.0	477
2-392B	Aug-97	36.0-46.0	1	1.2	4.4E-04	0.6	0.20	900'0	14
	,		2	6.0	3.2E-04	7.0	0.20	900'0	9.8
			Average	1.1	3.8E-04	8.0	0.20	0.006	12
2-393B	Aug-97	35.0-45.0	1	18	6.2E-03	132	0.20	900.0	193
			2	18	6.3E-03	134	0.20	9000	196
			Average	18	6.3E-03	133	0.20	900.0	194
	Average of Lower	of Lower USZ Sand Interval:	Interval:	15	5.37E-3	114	0.20	9000	167

^{a/} ft bgs = Feet below ground surface.

 $^{^{}b'}$ ft/day = Feet per day.

c' cm/sec = Centimeters per second.

 d^{d} gpd/ft² = Gallons per day per square foot.

e' ft/ft = Foot per foot.

 $^{^{\}prime\prime}$ ft/yr = Feet per year.

The effective porosity of a given lithology is a measure of the amount of void space available for fluid flow through the lithologic unit. Effective porosity for the USZ sand intervals is estimated to be approximately 20 to 25 percent, based on accepted literature values (Spitz and Moreno, 1996) and the high estimated hydraulic conductivities. However, effective porosity can reduced by localized intervals of cementation or fine-grained deposits. The effective porosity of the USZ/LSZ aquitard would be much less than that of the overlying USZ sands, with an estimated range of 0.5 to 5 percent based on a silt or cemented dense sandstone lithology (Spitz and Moreno, 1996).

Assuming isotropic horizontal hydraulic conductivity, the advective velocity of groundwater in the direction of maximum hydraulic gradient is given by:

$$\overline{v} = \frac{K}{n_s} \frac{dH}{dL}$$

Where: \overline{v} = Average advective groundwater velocity (seepage velocity) [L/T]

K = Hydraulic conductivity [L/T]

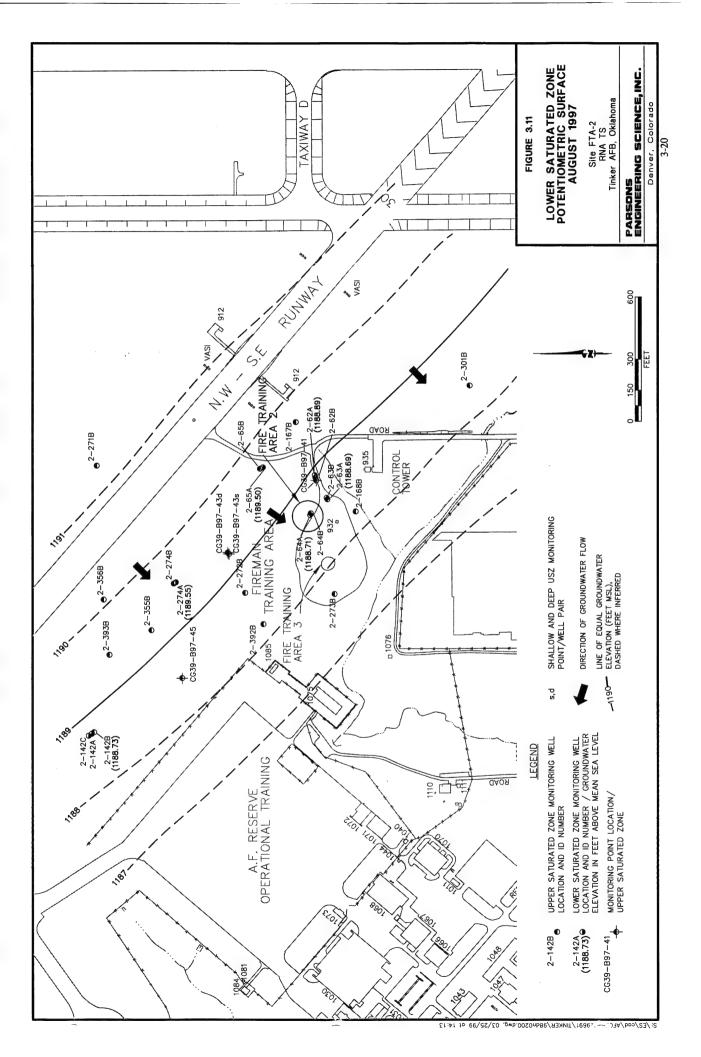
dH/dL = Lateral hydraulic gradient [L/L]

 n_e = Effective porosity.

Using the range of hydraulic conductivity (0.9 to 44 ft/day), an average horizontal hydraulic gradient of 0.006 ft/ft, and an estimated effective porosity of 0.20 for fine-grained sand, the advective groundwater flow velocity in the USZ is calculated (Table 3.2) to be 0.027 to 1.305 ft/day [9.8 to 477 feet per year (ft/yr)]. Based on average hydraulic conductivites, the average advective groundwater flow velocity in the upper USZ sand interval is calculated to be 0.417 ft/day (152 ft/yr), and the average advective groundwater flow velocity in the lower USZ sand interval is calculated to be 0.456 ft/day (167 ft/yr). Advective groundwater velocities in flow directions other than the direction of maximum hydraulic gradient would be less than these computed values.

3.3.2.2 Lower Saturated Zone

The top of the LSZ coincides approximately with the base of the USZ-LSZ aquitard at a depth of approximately 50 to 60 feet bgs. In general, the LSZ potentiometric surface lies beneath the base of the USZ-LSZ aquitard under unconfined conditions. Locally the LSZ may be confined where the LSZ potentiometric surface rises above the USZ-LSZ aquitard (monitoring wells 2-142A and 2-65A, Cross-sections C-C' and E-E'). Figure 3.11 is a groundwater potentiometric surface map for the LSZ based on data from August 1997. The direction of maximum hydraulic gradient and groundwater flow (assuming isotropic horizontal hydraulic conductivity) in the LSZ is towards the southwest at an approximate horizontal hydraulic gradient of 0.0033 ft/ft. The flow direction and gradient in the LSZ for 1997 are consistent with groundwater flow direction and gradient data observed in 1994 (IT, 1994), and 1996 (Parsons ES, 1997).



While the upper and lower USZ sand intervals are hydraulically connected, where it is developed the intermediate clay and silt interval may act as a significant vertical permeability barrier to local groundwater flow and contaminant transport.

3.4 GROUNDWATER USE

Groundwater from the USZ and LSZ at Tinker AFB is not extracted for beneficial use. Water used at the Base is supplied by deep (>400 feet) wells in the Garber-Wellington Formation, supplemented by water purchased from Oklahoma City. Midwest City also obtains potable water from deep wells screened in the Central Oklahoma Aquifer, including the Garber-Wellington Formation.

3.5 POTENTIAL PATHWAYS AND RECEPTORS AT FTA-2

Precipitation runoff at FTA-2 primarily flows to a tributary of Crutcho Creek or the Base storm sewer system. Groundwater at FTA-2 flows generally west-southwest, with local variations due to lithologic, topographic, or surface (tributary) features. A light industrial area is present west of FTA-2. Because the site is located near an industrial area on a secured military Base, Base workers are the most probable potential receptors that could be exposed to any site-related contamination. Other potential receptors include Tinker AFB water supply wells located on the western portion of the Base, over one mile away in a downgradient direction, and screened at a depth of over 400 feet bgs.

SECTION 4

CONTAMINANT DISTRIBUTION AND EVIDENCE FOR BIODEGRADATION

Soil and groundwater have been contaminated as a result of fire training activities at FTA-2. Work during the RFI (IT, 1994, and TT, 1996) focused on defining the nature and extent of contamination at the site, and annual groundwater monitoring is conducted (Tinker AFB, 1997). As needed, the data collected during those efforts are used to supplement the TS field investigation results presented in the following subsections. In particular, this section focuses on data useful for evaluating and modeling natural attenuation of CAHs dissolved in groundwater following a summary of hydrocarbon degradation.

4.1 OVERVIEW OF HYDROCARBON BIODEGRADATION

Primary mechanisms for natural attenuation of fuel hydrocarbons and CAHs include biodegradation, dispersion, dilution from recharge and upgradient flow, sorption, and volatilization. Of these processes, biodegradation is the only mechanism working to transform contaminants into ultimately innocuous byproducts. When indigenous microorganisms work to bring about a reduction in the total mass of contamination in the subsurface without the addition of nutrients, these biodegradation processes are considered intrinsic.

To provide a foundation for site interpretations, the following subsections review the major bioremediation processes that act upon CAHs and fuel hydrocarbons. Chlorinated solvents are the primary groundwater contaminants at FTA-2. Total fuel carbon, benzene, toluene, and chlorobenzenes are generally limited to the FTA-2 source area (monitoring wells 2-62B, 2-63B, 2-64B, and 2-65B). Because the terminology of describing biodegradation of CAHs and fuel hydrocarbons is similar, a generalized review of the processes is first presented. Subsequent sections focus on the biodegradation of CAHs.

4.1.1 Review of Biodegradation Processes

Microorganisms obtain energy for cell production and maintenance by facilitating thermodynamically advantageous reduction/oxidation (redox) reactions involving the transfer of electrons from electron donors to available electron acceptors. This results in the oxidation of the electron donor and the reduction of the electron acceptor. Electron donors may be natural organic carbon, fuel hydrocarbon compounds (BTEX), and less-chlorinated solvents [e.g., vinyl chloride (VC), dichloroethene (DCE), dichloroethane (DCA), or chlorinated benzenes ranging from chlorobenzene (CB) to tetrachlorobenzene].

Fuel hydrocarbons or solvents are completely degraded or detoxified if they are utilized as the primary electron donor (i.e., as a primary substrate or carbon source) for microbial metabolism (Bouwer, 1992). Electron acceptors are elements or compounds that occur in relatively oxidized states, and include oxygen, nitrate, ferric iron, sulfate, manganese, carbon dioxide, and highly chlorinated solvents [e.g., PCE, TCE, tetrachloroethane (PCA), trichloroethane (TCA), and polychlorinated benzenes].

The driving force of biodegradation is electron transfer, which is quantified by the Gibbs free energy of the reaction (ΔG°_{r}) (Stumm and Morgan, 1981; Bouwer, 1994; Godsey, 1994). The value of ΔG°_{r} represents the quantity of free energy consumed ($\Delta G^{\circ}_{r} > 0$) or yielded ($\Delta G^{\circ}_{r} < 0$) to the system during the reaction. Although thermodynamically favorable, most of the reactions involved in biodegradation of fuel hydrocarbons or CAHs cannot proceed abiotically due to a lack of activation energy. Microorganisms are capable of providing the necessary activation energy; however, they will facilitate only those redox reactions that have a net yield of energy (i.e. $\Delta G^{\circ}_{r} < 0$). Most reactions involving biodegradation of contaminants do yield energy to the microbes. However, specific geochemical conditions are often necessary for biodegradation reactions, which allow the appropriate microbial population to develop and grow.

Microorganisms preferentially utilize electron acceptors while metabolizing hydrocarbons (Bouwer, 1992). DO is utilized first as the prime electron acceptor where present. It is under aerobic conditions that fuel hydrocarbons and the less chlorinated solvents are most commonly used as electron donors. After the DO is consumed, anaerobic microorganisms use native electron acceptors (as available) in the following order of preference: nitrate, manganese, ferric iron hydroxide, sulfate, and finally carbon dioxide. Under anaerobic conditions, BTEX compounds and other fuel hydrocarbons are still used as electron donors. Chlorinated solvents that are amenable to reductive dehalogenation are generally used as electron acceptors when aquifer conditions are such that sulfate or carbon dioxide is the preferred electron acceptor. Because the biodegradation of fuel hydrocarbons and CAHs can deplete the concentrations of electron acceptors, examining the distribution of electron acceptor concentrations can provide evidence of whether biodegradation is occurring and the degree to which it is occurring.

The expected sequence of redox reactions in an aquifer is also a function of the ORP of the groundwater. ORP is a measure of the relative tendency of a solution or chemical reaction to accept or transfer electrons, and it can be used as an indicator of which redox reactions are operating at a site. As each subsequent electron acceptor is utilized, the groundwater becomes more reducing, and the redox potential of the water decreases. Microbial mediated redox reactions are the main driving force for ORP decreases.

Depending on the types and concentrations of electron acceptors present (e.g., nitrate, sulfate, carbon dioxide), pH conditions, and redox potential, anaerobic biodegradation can occur by denitrification, manganese reduction, ferric iron reduction, sulfate reduction, or methanogenesis. Other, less common anaerobic degradation mechanisms such as manganese or nitrate reduction may dominate if the physical and chemical conditions in the subsurface favor use of these electron acceptors. Vroblesky and Chapelle (1995) show that the dominant terminal electron accepting process can vary both temporally and spatially in an aquifer with fuel hydrocarbon contamination.

4.1.2 Biodegradation of Chlorinated Solvents

Chlorinated solvents can be transformed, directly or indirectly, by biological processes (e.g., Bouwer et al., 1981; Miller and Guengerich, 1982; Reineke and Knackmuss, 1984; Wilson and Wilson, 1985; de Bont et al., 1986; Nelson et al., 1986; Spain and Nishino, 1987; Bouwer and Wright, 1988; Little et al., 1988; Mayer et al., 1988; Arciero et al., 1989; Cline and Delfino, 1989; Freedman and Gossett, 1989; Folsom et al., 1990; Harker and Kim, 1990; Alvarez-Cohen and McCarty, 1991a, 1991b; DeStefano et al., 1991; Henry, 1991; Sander et al., 1991; McCarty et al., 1992; Hartmans and de Bont, 1992; Chapelle, 1993; McCarty and Semprini, 1994; Vogel, 1994; Suflita and Townsend, 1995; Bradley and Chapelle, 1996; Klier et al., 1996; Spain, 1996). Biodegradation of chlorinated solvents (or CAHs) and chlorinated benzenes results from the same general processes that bring about biodegradation of fuel hydrocarbons. However, a more complex series of processes often is involved, and CAHs and chlorinated benzenes may act as either substrates (electron donors) or electron acceptors depending upon what geochemical conditions prevail.

Whereas fuel hydrocarbons are biodegraded in essentially one step by acting as an electron donor/carbon source, CAHs and chlorinated benzenes may undergo several types of biodegradation involving multiple steps. CAHs may undergo biodegradation through three different pathways: use as an electron acceptor, use as an electron donor, or cometabolism, which is degradation resulting from exposure to a catalytic enzyme fortuitously produced during an unrelated process. At a given site, one or all of these processes may be operating, although at many sites the use of CAHs as electron acceptors appears to be the most likely.

In a pristine aquifer, native organic carbon is utilized as an electron donor and DO is utilized first as the prime electron acceptor. Where anthropogenic carbon (e.g., fuel hydrocarbons, less-chlorinated CAHs, or chlorinated benzenes with four or fewer chlorines) is present, it may also be utilized as an electron donor. Most chlorinated solvents that can act as electron donors have thus far only been demonstrated to do so under aerobic conditions, with the notable exception of VC (Bradley and Chapelle, 1996). After the DO is consumed, anaerobic microorganisms typically use native electron acceptors (as available) in the following order of preference: nitrate, ferric iron oxyhydroxide, sulfate, and finally carbon dioxide. Evaluation of the distribution of these electron acceptors can provide evidence of where and how CAH biodegradation is occurring. In addition, because CAHs may be used as electron acceptors or electron donors (in competition with other acceptors or donors), maps showing the distribution of these compounds also will provide evidence of the types of biodegradation processes acting at a site. A more complete description of the main types of biodegradation reactions affecting CAHs dissolved in groundwater is presented in the following subsections.

4.1.2.1 Electron Acceptor Reactions (Reductive Dehalogenation)

Under anaerobic conditions, biodegradation of chlorinated solvents usually proceeds through a process called reductive dehalogenation. During this process, the halogenated hydrocarbon is used as an electron acceptor, not as a source of carbon, and a chloride atom is removed and replaced with a hydrogen atom. The transformation of chlorinated

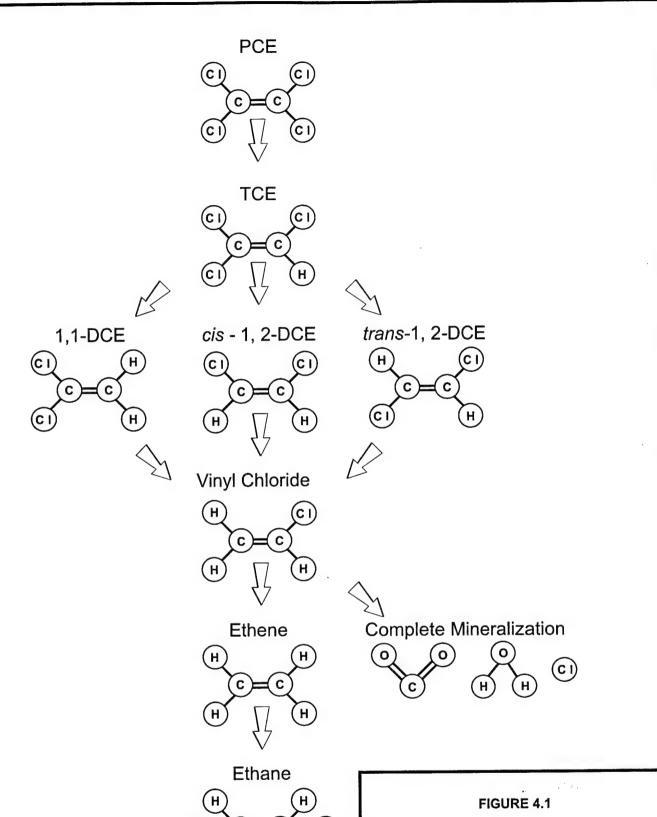
ethenes via reductive dehalogenation is illustrated in Figure 4.1. In general, reductive dehalogenation of chlorinated ethenes occurs by sequential dehalogenation from PCE to TCE to DCE to VC to ethene. An analogous pattern for chlorinated ethanes might be PCA to TCA to DCA to CA, and for chlorinated benzenes the pattern might be tetrachlorobenzene (TeCB) to trichlorobenzene (TCB) to dichlorobenzene (DCB) to CB. Depending upon environmental conditions, these sequences may be interrupted, with other processes (e.g., aerobic or abiotic degradation) then acting upon the products. Reductive dehalogenation of chlorinated solvent compounds is associated with the accumulation of daughter products and an increase in chloride concentrations.

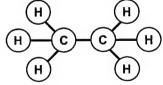
Reductive dehalogenation affects each of the chlorinated ethenes differently. Of the ethenes, PCE is the most susceptible to reductive dehalogenation because it is the most oxidized. Conversely, VC is the least susceptible to reductive dehalogenation, because it is the least oxidized of the ethene compounds. Likewise, hexachlorobenzene is more susceptible to reductive dehalogenation than less-chlorinated benzenes. In general, the rate of reductive dehalogenation of chlorinated solvents has been observed to decrease as the degree of chlorination decreases (Vogel and McCarty, 1985; Bouwer, 1994). Murray and Richardson (1993) have postulated that this rate decrease may explain the accumulation of VC in PCE and TCE plumes that are undergoing reductive dehalogenation.

Reductive dehalogenation can also be controlled by the redox conditions of the site groundwater system. In general, reductive dehalogenation has been demonstrated under anaerobic nitrate- and sulfate-reducing conditions, but the most rapid biodegradation rates, affecting the widest range of compounds, occur under methanogenic conditions (Bouwer, 1994). Dehalogenation of PCE and TCE to DCE can proceed under mildly reducing conditions such as nitrate reduction or ferric iron [iron (III)] reduction (Vogel et al., 1987). However, the transformation of DCE to VC, or the transformation from VC to ethene requires more strongly reducing conditions (Freedman and Gossett, 1989; DeStefano et al., 1991; De Bruin et al., 1992).

Reductive dehalogenation of some compounds can also preferentially produce specific daughter compounds. For example, during reductive dehalogenation of TCE or PCE, all three isomers of DCE can theoretically be produced. However, Bouwer (1994) reports that under the influence of biodegradation, *cis*-1,2-DCE is a more common intermediate than *trans*-1,2-DCE, and that 1,1-DCE is the least prevalent intermediate of the three DCE isomers.

When chlorinated compounds are used as electron acceptors, there must be an appropriate source of carbon for microbial growth in order for reductive dehalogenation to occur (Bouwer, 1994). Potential carbon sources/electron donors can include low-





ANAEROBIC REDUCTIVE DEHALOGENATION

Site FTA-2 RNA TS Tinker AFB, Oklahoma

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

molecular-weight compounds (e.g., lactate, acetate, methanol, or glucose) present in natural organic matter, fuel hydrocarbons, or less-chlorinated solvents (as discussed below).

4.1.2.2 Electron Donor Reactions

Under aerobic conditions some CAHs can be utilized as the primary substrate (i.e., electron donor) in biologically mediated redox reactions (McCarty and Semprini, 1994). In this type of reaction, the facilitating microorganism obtains energy and organic carbon from the degraded compound. In contrast to reactions in which the chlorinated compound is used as an electron acceptor, only the least oxidized chlorinated solvents (e.g., VC, DCE, or CB) may be utilized as electron donors in biologically mediated redox reactions. Also, chlorinated benzenes with up to four chlorines (i.e., CB, DCB isomers, TCB isomers, and TeCB isomers) can act as electron donors under aerobic conditions.

Microorganisms are generally believed to be incapable of growth using TCE and PCE, although other less chlorinated CAHs have been documented as substrates (Murray and Richardson, 1993). For example, Davis and Carpenter (1990) describe the aerobic oxidation of VC in groundwater, and McCarty and Semprini (1994) describe investigations in which VC and 1,2-DCA were shown to serve as primary substrates. Dichloromethane has the potential to function as a primary substrate under either aerobic or anaerobic environments (McCarty and Semprini, 1994), and all three DCE isomers can be mineralized under aerobic conditions (Klier *et al.*, 1996). In addition, Bradley and Chapelle (1996) show evidence of oxidation of VC under iron-reducing conditions as long as there is sufficient bioavailable iron (III). Aerobic metabolism of VC may be characterized by a loss of VC mass, a decreasing molar ratio of VC to other CAH compounds, and rarely, the presence of chloromethane.

CB and polychlorinated benzenes (up to and including TeCB) have been shown to biodegrade under aerobic conditions. Several studies have shown that bacteria are able to utilize CB (Reineke and Knackmuss, 1984), 1,4-DCB (Reineke and Knackmuss, 1984; Schraa et al., 1986; Spain and Nishino, 1987), 1,3-DCB (de Bont et al., 1986), 1,2-DCB (Haigler et al., 1988), 1,2,4-TCB (van der Meer et al., 1987; Sander et al., 1991), and 1,2,4,5-TeCB (Sander et al., 1991) as primary growth substrates in aerobic systems. Nishino et al. (1994) note that aerobic bacteria able to grow on chlorobenzene have been detected at a variety of chlorobenzene-contaminated sites but not at uncontaminated sites. Spain (1996) notes that this provides strong evidence that the bacteria are selected for their ability to derive carbon and energy from chlorobenzene degradation in situ. The pathways for all of these reactions are similar and are also similar to that of benzene (Chapelle, 1993; Spain, 1996). The only significant difference between aerobic oxidation of chlorobenzenes and benzene is the required elimination of chlorine. Anaerobic oxidation of chlorobenzenes has not been conclusively documented, although little work has been done on the subject (Spain, 1996).

4.1.2.3 Cometabolism

When a chlorinated solvent is biodegraded through cometabolism, it serves as neither an electron acceptor nor a primary substrate in a biologically mediated redox reaction. Instead, an enzyme or cofactor that is fortuitously produced by organisms for other purposes catalyzes the degradation of the compound. The organism receives no known benefit from degradation of the solvent; rather the cometabolic degradation of the solvent may in fact be harmful to the microorganism responsible for the production of the enzyme or cofactor (McCarty and Semprini, 1994).

Cometabolism is best documented for chlorinated ethenes in aerobic environments, although it potentially could occur with other chlorinated solvents or under anaerobic conditions. Aerobic degradation pathways for chlorinated ethenes are illustrated in Figure 4.2. It has been reported that under aerobic conditions chlorinated ethenes, with the exception of PCE, are susceptible to cometabolic degradation (Murray and Richardson, 1993; Vogel, 1994; McCarty and Semprini, 1994). Vogel (1994) further elaborates that the cometabolism rate increases as the degree of dehalogenation decreases.

In the cometabolic process, bacteria indirectly transform TCE while they use BTEX or another carbon source to meet their energy requirements. TCE does not enhance the degradation of carbon sources, nor does its cometabolism interfere with the use of electron acceptors involved in the oxidation of those carbon sources. Given this relationship, it would follow that depletion of suitable substrates (BTEX or other organic carbon sources) likely limits cometabolism of CAHs.

4.1.2.4 Abiotic Degradation of Chlorinated Solvents

Chlorinated solvents dissolved in groundwater may also degrade by abiotic mechanisms, although the reactions may not be complete and often result in the formation of a toxic intermediate. The most common abiotic reactions affecting chlorinated solvents are hydrolysis and dehydrohalogenation. Hydrolysis is a substitution reaction in which a halogen substituent is replaced with a hydroxyl (OH) group from a water molecule. Dehydrohalogenation is an elimination reaction in which a halogen is removed from a carbon atom, followed by removal of a hydrogen atom from an adjacent carbon atom, with a double bond between the carbon atoms being produced. Other possible reactions include oxidation and reduction, although no abiotic oxidation reactions involving common halogenated solvents have been reported in the literature (Butler and Barker, 1996).

Hydrolysis of chlorinated methanes and ethanes has been well demonstrated in the literature (e.g., Vogel et al., 1987; Jeffers et al., 1989; Vogel, 1994; Butler and Barker, 1996). The likelihood that a solvent will hydrolyze depends partly on the number of halogen substitutes, typically with fewer halogens resulting in more rapid hydrolysis. Dehydrohalogenation, on the other hand, is more likely to take place as the number of halogen substitutes increases. One common solvent for which abiotic degradation reactions are well documented is 1,1,1-TCA. 1,1,1-TCA may be transformed through a series of abiotic processes (including hydrolysis) to acetic acid. 1,1,1-TCA can also dehydrohalogenate to 1,1-DCE (Vogel and McCarty, 1985). and

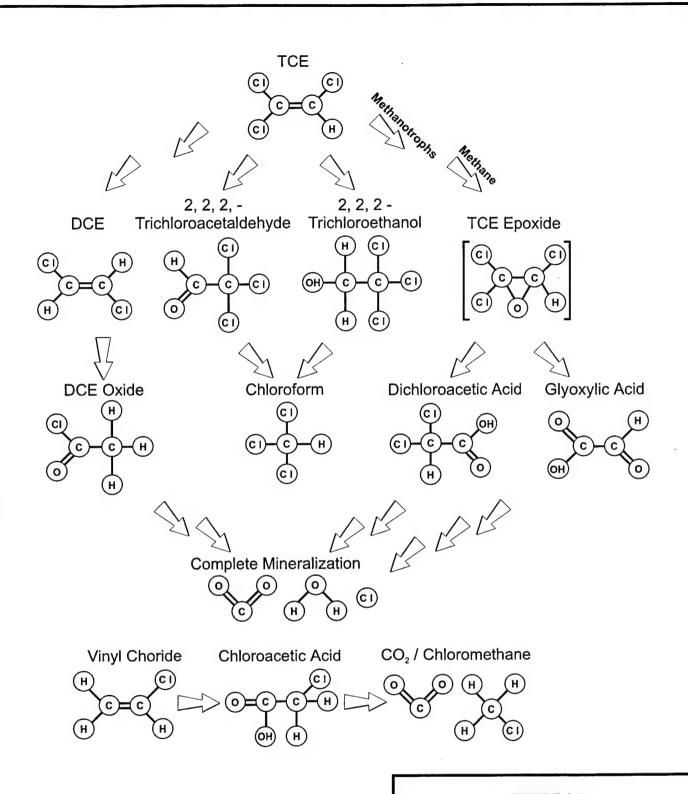


FIGURE 4.2

AEROBIC DEGRADATION PATHWAYS

FTA-2 RNA TS Tinker AFB, Oklahoma

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

1,1,2-TCA can react to form 1,1-DCE (Jeffers et al., 1989). Once TCA is reductively dehalogenated to chloroethane (CA), it can then hydrolyze to ethanol (Vogel and McCarty, 1985) or dehydrohalogenate to VC (Jeffers et al., 1989).

Attributing changes in the presence, absence, or concentration of halogenated solvents to abiotic processes is usually difficult, particularly on the field scale (Butler and Barker, 1996). Solvents may undergo both biotic and abiotic degradation, and discerning the effects of each mechanism (on the field scale), if possible, is difficult. Also, the breakdown products of some reactions such as hydrolysis (e.g., acids and alcohols) may be further degraded (biotically or abiotically) to products, which require additional analyses that may not be feasible for a field investigation (Butler and Barker, 1996). CAH hydrolysis has not been successfully demonstrated at the field scale (Butler and Barker, 1996). Evidence of dehydrohalogenation is also difficult to collect, although the presence of 1,1-DCE in conjunction with 1,1,1-TCA can provide a tentative indication that the process is ongoing. The presence of 1,1-DCE provides strong evidence of dehydrohalogenation where it is known that no DCE has been released.

4.1.2.5 Behavior of Chlorinated Solvent Plumes

Chlorinated solvent plumes can exhibit three types of behavior depending on the amount of solvent, the amount of organic (native and/or anthropogenic) carbon in the aquifer, the distribution and concentration of natural electron acceptors, and the types of electron acceptors being utilized. Individual plumes may exhibit all three types of behavior in different portions of the plume.

4.1.2.5.1 Type 1 Behavior

Type 1 behavior occurs where the primary substrate is anthropogenic carbon (e.g., BTEX or landfill leachate), and this anthropogenic carbon drives reductive dechlorination. When evaluating natural attenuation of a plume exhibiting Type 1 behavior the following questions must be answered:

- 1) Is the electron donor supply adequate to allow microbial reduction of the chlorinated organic compounds? In other words, will the microorganisms "strangle" before they "starve" [i.e., will they run out of CAHs (electron acceptors) before they run out of primary substrate (anthropogenic carbon)]?
- 2) What is the role of competing electron acceptors (e.g., DO, nitrate, iron (III) and sulfate)?
- 3) Are VC and other less-chlorinated solvents being oxidized, or are they being reduced?

Type 1 behavior results in the rapid and extensive degradation of the highly chlorinated solvents such as PCE, TCE, or polychlorinated benzenes.

4.1.2.5.2 Type 2 Behavior

Type 2 behavior dominates in areas characterized by relatively high concentrations of biologically available native organic carbon. This natural carbon source drives reductive dehalogenation (i.e., the primary substrate for microorganism growth is native organic carbon). When evaluating natural attenuation of a Type 2 chlorinated solvent plume, the same questions as those posed in the description of Type 1 behavior must be answered. Type 2 behavior generally results in slower biodegradation of the highly chlorinated solvents than Type 1 behavior, but under the right conditions (e.g., areas with high natural organic carbon contents), this type of behavior also can result in rapid degradation of CAH compounds.

4.1.2.5.3 Type 3 Behavior

Type 3 behavior dominates in areas characterized by low concentrations of native and/or anthropogenic carbon, and concentrations of DO that are greater than 1.0 mg/L. Under these aerobic conditions reductive dehalogenation will not occur. Thus there is little or no removal of PCE and TCE. Biodegradation may proceed via the much slower process of cometabolism, but will be limited by the low concentrations of native or anthropogenic carbon. The most significant natural attenuation mechanisms for CAHs will be advection, dispersion, and sorption. However, VC can be rapidly oxidized under these conditions, DCE may be oxidized, CBs may be oxidized, and cometabolism also may occur.

4.1.2.5.4 Mixed Behavior

A single chlorinated solvent plume can exhibit all three types of behavior in different portions of the plume. This mixed behavior can be beneficial for natural biodegradation of CAH plumes. For example, Wiedemeier et al. (1996b) describe a CAH plume at Plattsburgh AFB, New York that exhibits Type 1 behavior in the source area and Type 3 behavior downgradient along the contaminant flowpath from the source. The best scenario involves a plume in which PCE, TCE, and DCE are reductively dehalogenated (Type 1 or Type 2 behavior), then VC is oxidized (Type 3 behavior), either aerobically or anaerobically (via iron reduction). VC is oxidized to carbon dioxide in this type of plume and does not accumulate. The following sequence of reactions occurs in a plume that exhibits mixed behavior:

$PCE \rightarrow TCE \rightarrow DCE \rightarrow VC \rightarrow Carbon Dioxide$

In general, the TCE, DCE, and VC may attenuate at approximately the same rate, and the reductive dehalogenation reactions may be confused with simple dilution. Note that no ethene is produced, and VC is removed from the system much faster under oxidizing conditions than under VC-reducing conditions.

A less desirable scenario, but one in which all contaminants may be entirely biodegraded, involves a plume in which all CAHs are reductively dehalogenated via Type 1 or Type 2 behavior. VC is reduced to ethene, which may be further reduced to ethane or methane. The following sequence of reactions occur in this type of plume:

$PCE \rightarrow TCE \rightarrow DCE \rightarrow VC \rightarrow Ethene or Ethane$

In a reductive plume, VC degrades more slowly than TCE, and tends to accumulate (Freedman and Gossett, 1989).

4.2 NATURE AND EXTENT OF CONTAMINATION

The first step for evaluating the occurrence and methods of CAH and fuel hydrocarbon biodegradation is to look at the distribution of contaminants and the biodegradation products of those compounds. At site FTA-2, dissolved groundwater contamination consists of a mixture of compounds, some of which may be electron donors, some of which may be electron acceptors, and some of which may be either donors or acceptors. It is therefore important to evaluate the distribution of all relevant compounds and the spatial relations between their distributions.

The following sections present contaminant and daughter product distributions at FTA-2, providing the initial evidence that dissolved fuels and chlorinated solvents are biodegrading. Geochemical data presented in later sections also provides evidence of contaminant biodegradation.

4.2.1 Contaminant Sources

Sources of dissolved groundwater contamination at FTA-2 are not clearly defined. It is likely that residues from fuel or extinguishing agents have leached into soil and groundwater at the fire training pit and are a primary source of groundwater contamination. However, the exact locations and nature of the releases is not clear from the available data. Residual non-aqueous phase liquid (NAPL) has not been identified during either previous or current subsurface investigations.

In a 1987 soil investigation by the USACE (USACE, 1988), seven soil boreholes were drilled to a maximum depth of 7 feet and soil samples were collected and analyzed for VOCs and metals. From October to December 1993, 21 soil samples were collected from four locations at depths from 2 to 30 feet bgs and analyzed for VOCs, SVOCs, total petroleum hydrocarbons (TPH), and metals (IT, 1994). In 1995, an additional 37 soil samples were collected from five soil boreholes and two monitoring well boreholes (TT, 1996). Low concentrations of TPH, VOC, and SVOC compounds were detected, including methylene chloride, 1,1,1-TCA, butybenzyphthalate, di-n-butylphthalate, and bis(2-ethylhexyl)phthalate. TCA was detected at concentrations ranging from 0.07J (estimated concentration) to 6.7 micrograms per kilogram (μg/kg). Metals detected at the site were within background concentration ranges.

A total of six soil samples were collected from four soil boreholes (CG39-B97-41, CG-B97-43d, 2-392B, and 2-393B) in July 1997. Two samples were analyzed for TCE and PCE, and all six samples were analyzed for TOC. The results are summarized in Table 4.1. TCE was detected at concentrations of 0.019 and 0.070 milligrams per kilogram (mg/kg) in the soil sample and its replicate collected at a depth of 18 to 19 feet bgs from the borehole for CG39-B97-41. PCE was not detected in either sample.

PCE, TCE AND TOC IN SOIL^{b/} TABLE 4.1 JULY 1997

TINKER AIR FORCE BASE, OKLAHOMA FTA-2 RNA TS

Sample	Sample	Sample Depth	Sample	TCEb'	PCE ^{b/}	TOC ^{b/}
Location	Date	(feet bgs) ^{a/}	Replicate	(mg/kg) ^{c/}	(mg/kg)	(percent)
CG39-B97-41	7/28/1997	18-19	Replicate 1	0.019	ND®	0.952
CG39-B97-41	7/28/1997	18-19	Replicate 2	0.07	ON	0.816
CG39-B97-41	7/28/1997	18-19	Replicate 3	NA®	NA	0.819
					Mean TOC:	0.862
CG39-B97-43d	7/29/1997	29-30	Replicate 1	QN	QN	0.894
CG39-B97-43d	7/29/1997	29-30	Replicate 2	QN	QN	0.832
CG39-B97-43d	7/29/1997	29-30	Replicate 3	NA	NA	698'0
					Mean TOC:	0.865
2-392B	7/29/1997	29-30	Replicate 1	NA	NA	696'0
2-392B	7/29/1997	29-30	Replicate 2	NA	NA	0.632
2-392B	7/29/1997	29-30	Replicate 3	NA	NA	0.888
					Mean TOC:	0.830
2-392B	7/29/1997	44-45	Replicate 1	NA	NA	908'0
2-392B	7/29/1997	44-45	Replicate 2	NA	NA	098'0
2-392B	7/29/1997	44-45	Replicate 3	NA	NA	0.790
					Mean TOC:	0.819
2-393B	7/30/1997	34-35	Replicate 1	NA	NA	0.857
2-393B	7/30/1997	34-35	Replicate 2	NA	NA	0.638
2-393B	7/30/1997	34-35	Replicate 3	NA	NA	0.697
					Mean TOC:	0.731
2-393B	7/30/1997	44-45	Replicate 1	NA	NA	0.684
2-393B	7/30/1997	44-45	Replicate 2	NA	NA	0.755
2-393B	7/30/1997	44-45	Replicate 3	NA	NA	0.873
					Mean TOC:	0.771
				Average of Mo	Average of Mean TOC Values:	0.813
2 624 622 624 621						

² feet bgs = feet below ground surface.

^{b/} PCE = tetrachloroethene, TCE = trichloroethene, and TOC = Total Organic Carbon.

c/ mg/kg = milligrams per kilogram.

 $^{^{}d/}$ ND = not detected.

e' NA = compound was not analyzed for.

The presence of TCE and TCA in soil does suggest the use of chlorinated solvents at this site, though the low detected concentrations of TCE and TCA in soils do not confirm that the FTA-2 fire pit is the primary source for chlorinated solvents in groundwater. The presence of TCE in soil at borehole CG39-B97-41 coincides with the maximum concentrations of total fuel carbon, benzene, and TCE detected in USZ groundwater at monitoring well 2-62B, located immediately adjacent to CG39-B97-41. These wells are located approximately 75 feet upgradient from, and along the access road to, the site identified from aerial photographs as FTA-2 (Figure 1.3). Therefore, the source area for FTA-2 may be larger than estimated from the actual fire pit area identified from aerial photographs, or the access road may have been used for other purposes presenting additional sources (i.e., cleaning or dumping).

4.2.2 Total Organic Carbon in Soil

TOC concentrations are used to estimate the amount of organic matter sorbed to soil particles or trapped in the interstitial passages of a soil matrix. The TOC concentration in the saturated zone is an important parameter used to estimate the amount of contaminant that could potentially sorb to organic carbon in the aquifer matrix. Sorption results in retardation of the contaminant plume relative to the average advective groundwater velocity. In addition, TOC can be used as a gross indicator of organic compounds that are available as a source of carbon and electron donors (or substrate) for a microbial population.

Soil TOC concentrations were measured in six samples from four borehole locations in 1997 (Table 4.1). All samples were collected from below the water table. Soil TOC concentrations ranged from 0.632 to 0.969 percent, and average 0.813 percent. For most of the CAHs detected at FTA-2, contaminant retardation due to sorption is significant when TOC concentrations are greater than approximately 0.01 percent. Because TOC concentrations at FTA-2 are significantly greater than 0.01 percent, sorption of dissolved CAHs onto organic carbon may play an important role in contaminant retardation at FTA-2.

4.2.3 Fuel Hydrocarbons in Groundwater

Total fuel carbon was detected in two of eighteen groundwater samples collected from the USZ aquifer at FTA2. Total fuel was not detected in the LSZ aquifer. Results for fuel hydrocarbon analyses performed on samples collected in August 1997 are presented in Table 4.2. Total fuel carbon was detected at concentrations of 1,140 and 5.8 micrograms per liter (μ g/L) for wells 2-62B and 2-168B, respectively. Compound specific fuel hydrocarbons detected in groundwater at FTA-2 are benzene and toluene. Benzene was detected at two locations, 2-62B and 2-168B, at concentrations of 6.0 and 2.5 μ g/L, respectively. Only the detected concentration at 2-62B exceeds the USEPA maximum contaminant level (MCL) of 5 μ g/L. Toluene was only detected at location 2-62B at a concentration of 1.2 μ g/L.

Based on the location of fuel hydrocarbon detections in groundwater, the contamination source at FTA-2 is in a small isolated area near 2-62B and 2-168B. This area is located hydraulically upgradient and cross-gradient of the FTA-2 pit. Detection of fuel hydrocarbons at location 2-62B coincides with the highest detected concentration

FUEL HYDROCARBONS IN GROUNDWATER FTA-2 RNA TS TINKER AIR FORCE BASE, OKLAHOMA AUGUST 1997 TABLE 4.2

Cample	Comple	Total Fuel	Donzana	Toluene	Ethyl-	n Vylana	m. Vylana	o. Vulana	1 3 5_TMB	1.2.4.TMB	1 2 3-TMR
Location	Date	$(\mu g/L)^{a'}$	μg/L)	(µg/L)	(µg/L)	μ-λεγιοπο (μg/L)	m-Ayreme (μg/L)	μg/L)	(μg/L)	(μg/L)	(μg/L)
	MCL ^{b/}	در	5	1000	700	10,000 ^{d/}	10,000 ^{d/}	10,000 ^{d/}	-		-
Ilnnar Coturated Zone	4 Zone										
CC30 D07 41	9/4/1007	NIDe/	21/2	div	CN.	Š	C.N	CIN	div	CIN	CN
CG39-B97-41	8/4/1007	CIN CIN	2 5	S S	2 2	S S	C. C.	CN CN	CN CN	GN CN	CN CN
CG39-B97-438	8/4/1007	GN ON	S S	S S	2 5	CN CN	2 2	QV CN	CN CN	QN CN	GN CN
CG39-B97-45	8/4/1997	QX	2	R	QN	2	Q.	QN	Q	QN	S
2-392B	8/3/1997	ND	QN	QN	QN	QN	QN	ND	QN	ND	QN
2-393B	8/4/1997	BLQ ⁶	BLQ	QN	QN	QN	QN	QN	ON	ND	QN
2-62B	7/31/1997	1140	9	1.2	ND	ND	QN	BLQ	BLQ	ND	ND
2-63B	7/31/1997	QN	QN	QN	QN	QN	QN	QN	ND	ON	ND
2-64B	8/1/1997	QN	QN	QN	QN	QN	QN	ND	ND	ND	ND
2-65B	8/1/1997	ND	QN	ND	ND	ND	QN	ND	ND	ND	ND
2-167B	8/1/1997	ND	QN	QN	QN	QN	QN	ND	ND	QN	ND
2-168B	7/31/1997	5.8	2.5	ND	ND	QN	ND	ND	ND	ND	ND
2-271B	8/3/1997	QN	QN	ND	QN	ND	ND	QN	QN	ND	ND
2-272B	8/1/1997	QN	QN	QN	QN	QN	QN	ND	ND	QN	ND
2-273B	2661/1/8	QN	QN	ND	QN	ND	ND	QN	QN	ND	QN
2-274B	8/3/1997	ND	ND	ND	ND	QN	ND	ND	ND	ND	ND
2-355B	8/3/1997	ND	ND	QN	QN	QN	QN	ND	ND	ND	QN
3-356B ^{g/}	9/4/1997	ηVN	QN	ON	QN	QN	ND	QN	QN	ND	NA
2-142B	8/2/1997	QN	ND	QN	ND	QN	QN	QN	QN	ND	QN
Lower Saturated Zone	d Zone										
2-62A	7/31/1997	QN	ND	QN	QN	QN	ND	QN	QN	ND	QN
2-63A	7/31/1997	QN	ND	QN	QN	QN	ND	QN	QN	ND	ND
2-64A	2661/1/8	QN	ND	QN	QN	QN	QN	QN	QN	ND	ND
2-65A	2661/1/8	QN	QN	QN	QN	ND	ND	QN	QN	ND	ND
2-274A	8/3/1997	QN	QN	QN	ND	ND	ND	QN	QN	ND	ND
2-142A	8/2/1997	ND	ND	ND	ND	ND	ND UN	ND	ND	ND	ND
$\mu g/L = micro$	$\mu g/L = micrograms per liter.$.:					$^{c'}$ ND = Not D	betected.			

 $\mu g/L = \text{micrograms per niter.}$ b' MCL = Maximum Contaminant Level, State of Oklahoma OAC 165:25-3-65+A21, or

U.S. EPA Primary Drinking Water Regulations, May 14, 1996.

 c^{\prime} -- = no standard listed. ^{d/} Standard listed is for total xylenes.

" BLQ = below limit of quantification.

g/ Monitoring well 3-356B was sampled on 9/4/97 by Brown and Root.

h' NA = not analyzed.

of TCE (Section 4.3.2), suggesting fuel hydrocarbons originated from the same source as CAHs at the site. Because of the limited occurrence and low concentrations of fuel hydrocarbons detected at the site, natural attenuation of benzene and toluene is not further analyzed. However, the presence of low levels of anthropogenic carbon (fuel hydrocarbons) within the source area may impact biodegradation of CAHs.

4.2.4 Dissolved Chlorinated Ethenes

Chlorinated ethenes detected in groundwater at FTA-2 include PCE, TCE, DCE, and VC. Data for chlorinated ethenes and other chlorinated compounds are presented in Table 4.3. Distributions of TCE, cis-1,2-DCE and trans-1,2-DCE for both the upper and lower USZ sand intervals are shown on Figures 4.3 through 4.5, respectively. Concentrations for PCE, 1,1-DCE and VC are shown on Figure 4.6.

As measured in August 1997, PCE was detected in three of the nineteen groundwater samples collected from the USZ aquifer at FTA-2 (Figure 4.6). PCE concentrations detected in groundwater ranged from 1.9 µg/L at 2-63B to 52 µg/L at 2-65B. PCE was also detected at a concentration of 6.1 µg/L at 2-62B. The detected concentration at monitoring well 2-65B is above the USEPA MCL for PCE of 5 µg/L. All three locations correspond to an area approximately 100-200 feet east and north of the FTA-2 fire pit area, in an USZ upgradient direction. Locations 2-62B and 2-63B are screened in the upper USZ sand interval, while location 2-65B is screened in the lower USZ sand interval. It is unclear whether the presence of PCE is directly associated with the former FTA-2 fire pit location.

As measured in August 1997, TCE was detected in thirteen of the nineteen groundwater samples collected from the USZ aquifer at FTA-2 (Figure 4.3). TCE concentrations detected in USZ groundwater ranged from 1.4 μ g/L at 2-168B to 9,440 μ g/L at 2-62B. Of these detections, 11 locations exceed the USEPA MCL of 5 μ g/L. The highest detection of TCE at 2-62B suggests that there may be a source of TCE associated with the area located to the east of the FTA-2 fire pit in an upgradient direction. The concentration of TCE within the FTA-2 fire pit area was 914 μ g/L at 2-64B. Based on the distribution of TCE, the source of CAHs (PCE and TCE) may be the FTA-2 fire pit, or upgradient locations near 2-62B and 2-65B located along access roads in the area. Although undocumented, the data suggest possible cleaning or dumping operations may also be potential sources in the general fire training area. TCE was detected in one of the six groundwater samples collected from the LSZ aquifer at FTA-2. At 2-62A, TCE was detected at 10.9 μ g/L, which is above the USEPA MCL for TCE.

Cis-1,2-DCE was detected at several wells (Figure 4.4), with the highest concentration of 1,200 µg/L detected at 2-62B. Concentrations of cis-1,2-DCE exceed the USEPA MCL of 70 µg/L at four locations. The two locations with the highest concentrations of cis-1,2-DCE coincide with the highest TCE detections at 2-62B and CG3-B97-43s in the upper USZ sand interval. In addition, another apparent plume of cis-1,2-DCE was detected downgradient in the lower USZ sand interval extending from well 2-65B to 2-393B. The concentrations at 2-393B and 2-274B in the lower USZ sand interval were slightly higher than measured at 2-355B located between the two wells, suggesting the lower USZ downward into the cis-1,2-DCE plume is migrating

TABLE 4.3
CHLORINATED ALIPHATIC HYDROCARBONS IN GROUNDWATER
AUGUST 1997
FTA-2 RNA TS
TINKER AIR FORCE BASE, OKLAHOMA

Sample	Sample	PCE ^{a/}	TCE*	1,1-DCE ^{3/}	cis-1,2-DCE	trans-1,2-DCE	VC*	1,1,1-TCA ^{a/}	1,1-DCA ^{2/}
Location	Date	(μg/L) ^{b/}	(µg/L)	(µg/L)	. (µg/L)	(µg/L)	(µg/L)	(μg/L)	(μg/L)
	MCΓα	5	5	7	70	100	2	200	Φ
Upper Saturated Zone	one								
CG39-B97-41	8/4/1997	ND	3.6	ND	<i>u</i>	QN	ND	ND	ND
CG39-B97-43s	8/4/1997	1	1490	1.1	183	6.1	ND	ND	ND
CG39-B97-43d	8/4/1997	QN	169	ND	31.8	2.4	ND	ND	ND
CG39-B97-45	8/4/1997	ND	153	ND	22.6		ND	ND	ND
2-392B	8/3/1997	ND	1	ND	ND	QN	ND	ND	ND
2-393B	8/4/1997	ND	406	1	54.3		ND	ND	ND
3-393B-Dup ^{g/}	8/4/1997	ND	425	1	54.5		ND	ND	ND
2-62B	7/31/1997	6.1	>2000	9.5	1200	76.2	1.7	ND	ND
2-62B-Dup	7/31/1997	!	9440	1	1110	11	ND	ND	ND
2-63B	7/31/1997	1.9	75.4	_	191	3.6	9.9	ND	4.6
2-64B	8/1/1997		914	ı	154	17.6	•••	ND	DN
2-65B	8/1/1997	52	06		37	2.5	1.3	QN	!
2-167B	8/1/1997	ND	ND	ND	ND	ND	ND	QN	ND
2-168B	7/31/1997		1.4	***	27.6	ND	ND	QN	1.2
2-271B	8/3/1997	ND	QN	ND	ND	ND	ND	QN	ND
2-272B	8/1/1997	ND	28.1	ND	2.1	QN	ND	ND	QN
2-273B	8/1/1997	QN	QN	ND	ND	ND	ND	ND	ND
2-274B	8/3/1997	1	518	1	51.7	2.2	ND	ND	ND
2-274B-Dup	8/3/1997		909		53.4	2.1	ND	QN	ND
2-355B	8/3/1997	QN	344	QN	50		ND	ND	ND
2-356B ^{IV}	9/4/1997	ND	ND	ND	ND	ND	ND	ND	QN
2-142B	8/2/1997	ND		ND	ND	ND	ND	QN	QN
Lower Saturated Zone	one								
2-62A	7/31/1997	QN	6'01	ND	gas are ten	ND	ND	ND	QN
2-63A	7/31/1997	ND		ND	ND	ND	ND	ND	ND
2-64A	2661/1/8	ND	ND	QN	ND	ND	ND	ND	ND
2-65A	8/1/1997	QN	ND	ND	ND	ND	ND	ND	ND
2-274A	8/3/1997	QN	ND	ND	ND	ND	ND	QN	ND
2-142A	8/2/1997	ND	QN	ND	QN	ND	ND	ND	ND
					(Continued)				

(Continued)

TABLE 4.3 (Continued) CHLORINATED ALIPHATIC HYDROCARBONS IN GROUNDWATER AUGUST 1997 FTA2 RNA TS TINKER AIR FORCE BASE, OKLAHOMA

Sample	Sample	1,2-DCA	Chlorobenzene	1,2-DCB	1,3-DCB ^{2/}	1,4-DCB	Carbon Tetrachloride	Chloroform
Location	Date	(μg/L)	(µg/L)	(µg/L)	(μg/L)	(μg/L)	(μg/L)	(µg/L)
	MCL ^o	5	100	009	009	75	5	
Upper Saturated Zone	ne .							
CG39-B97-41	8/4/1997		ND	-	ND	ND	ND	ND
CG39-B97-43s	8/4/1997	2.4	_	QN	1.6	QN	ND	-
CG39-B97-43d	8/4/1997		ND	ND		QN	ND	ND
CG39-B97-45	8/4/1997	1.4	QN	ND	****	QN	ND	ND
2-392B	8/3/1997	QN	ND	ND	ND	ND	ND	ND
2-393B	8/4/1997	45.4		ND		QN	ND	1
2-393B-Dup	8/4/1997	49.2	1	QN		ND	ND	9 9 9 9
2-62B	7/31/1997	260	117	950	34.8	168	ND	5.3
2-62B-Dup	7/31/1997	239	104	696	25	123	ND	1
2-63B	7/31/1997	:	QN	-	QN	QN	ND	ND
2-64B	8/1/1997	1.3	-	1	1.2	•••	ND	
2-65B	8/1/1997	3.2	9.5	3.6		1.8	ND	ND
2-167B	8/1/1997	QN	QN	ND	QN	QN	ND	ND
2-168B	7/31/1997	QN	QN	ND	QN	ND	ND	ND
2-271B	8/3/1997	QN	ND	ND	ND	ND	ND	QN
2-272B	1661/1/8	I	QN	ND	GN	ND	ND	QN
2-273B	8/1/1997	QN	ND	ND	QN	ND	ND	ND
2-274B	8/3/1997	1.1	1	ND		ND	ND	1
2-274B-Dup	8/3/1997	1.4		ND	-	ND	ON	1
2-355B	8/3/1997	1.8		ND	-	QN	QN	***
2-356B	9/4/1997	ND	ND	ND	ND	QN	QN	QN
2-142B	8/2/1997	ND	ND	ND	ND	QN	ND	ND
Lower Saturated Zone	one							
2-62A	7/31/1997	ND	ND		ND	UD	ND	QN
2-63A	7/31/1997	ND	ND	*****	ND	ND	QN	QN
2-64A	4661/1/8	ND	ND	ND	ND	QN	QN	ND
2-65A	8/1/1997	QN	ND	ND	ND	ND	QN	QN
2-274A	8/3/1997	ND	ND	QN	QN	ND	ND	ND
2-142A 8/2/1997 ND	8/2/1997	ND	QN	QN	QN	ND	ND	ND
VC = vinvl chlor	ide DCE = dichi	Inroethene DCA	= dichloroethane. TCA = trichloroethane. TCE = trichloroethene	4 = trichloroethane	TCE = trichloroet	hene,		

VC = VINY = VI

PCE = tetrachloroethene, DCB = dichlorobenzene. bV $\mu g/L$ = micrograms per liter. cI MCL = Maximum Contaminant Level, U.S. EPA Primary Drinking Water Regulations, May 14, 1996.

 ω' -- = no standard listed.

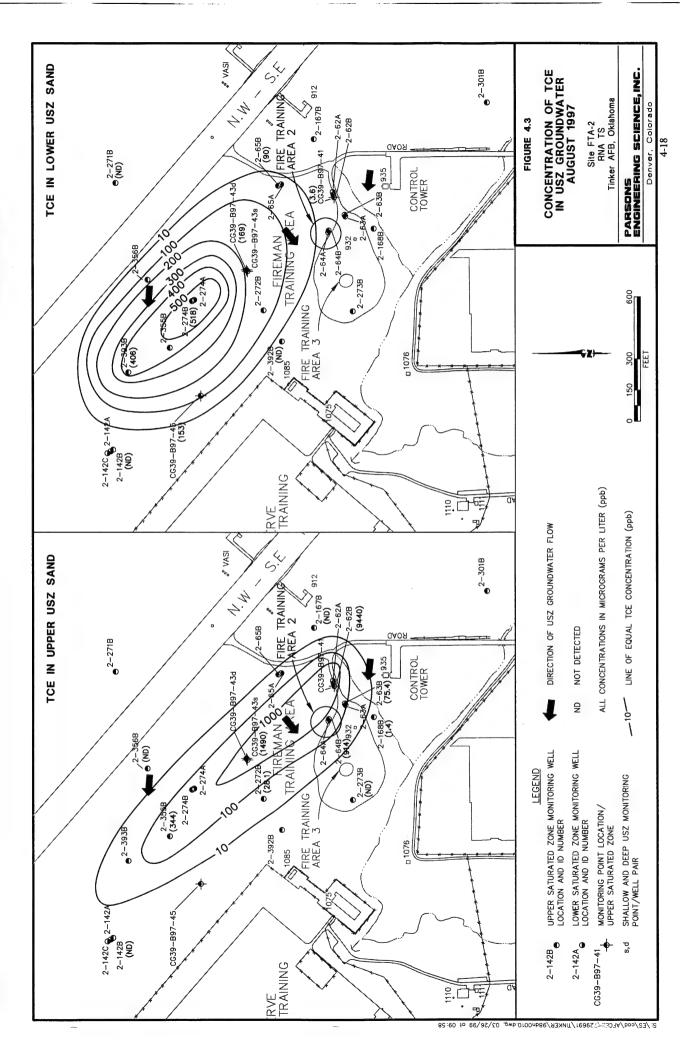
e' ND = not detected.

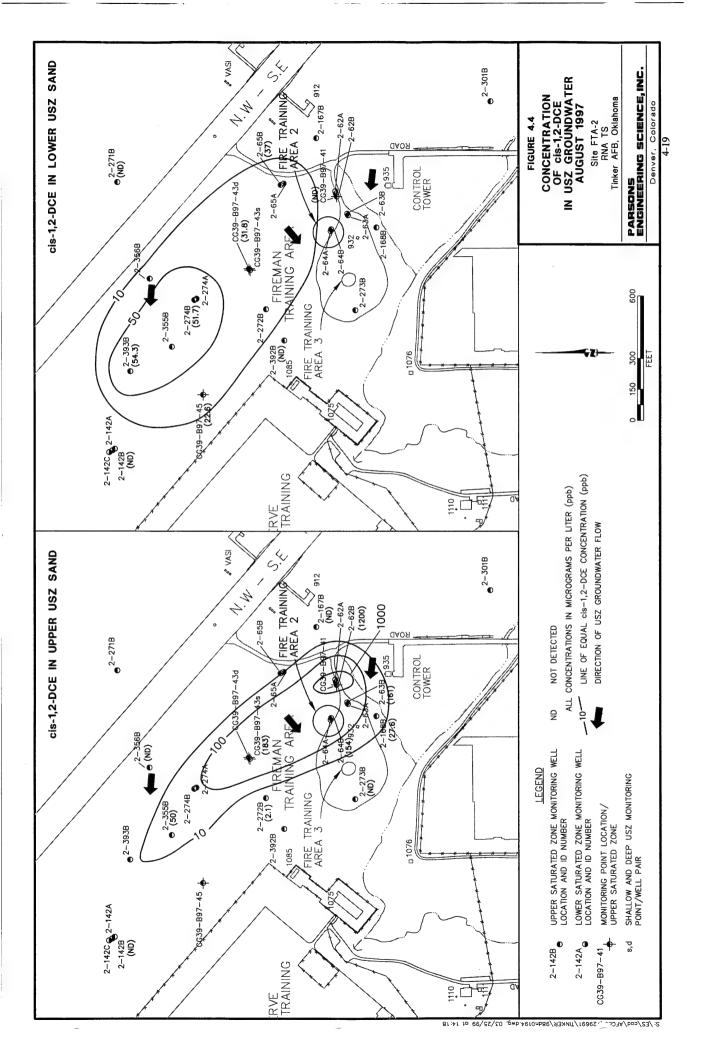
 θ --- = below calibration limit (1 part per billion).

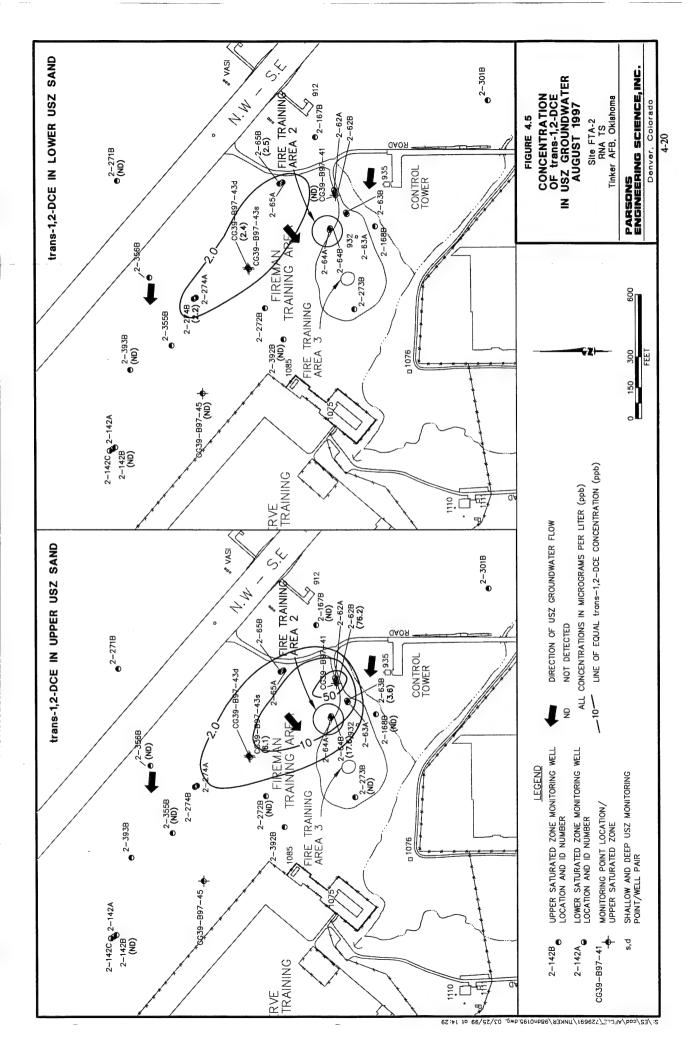
 $^{g'}$ Dup = field duplicate of preceding sample.

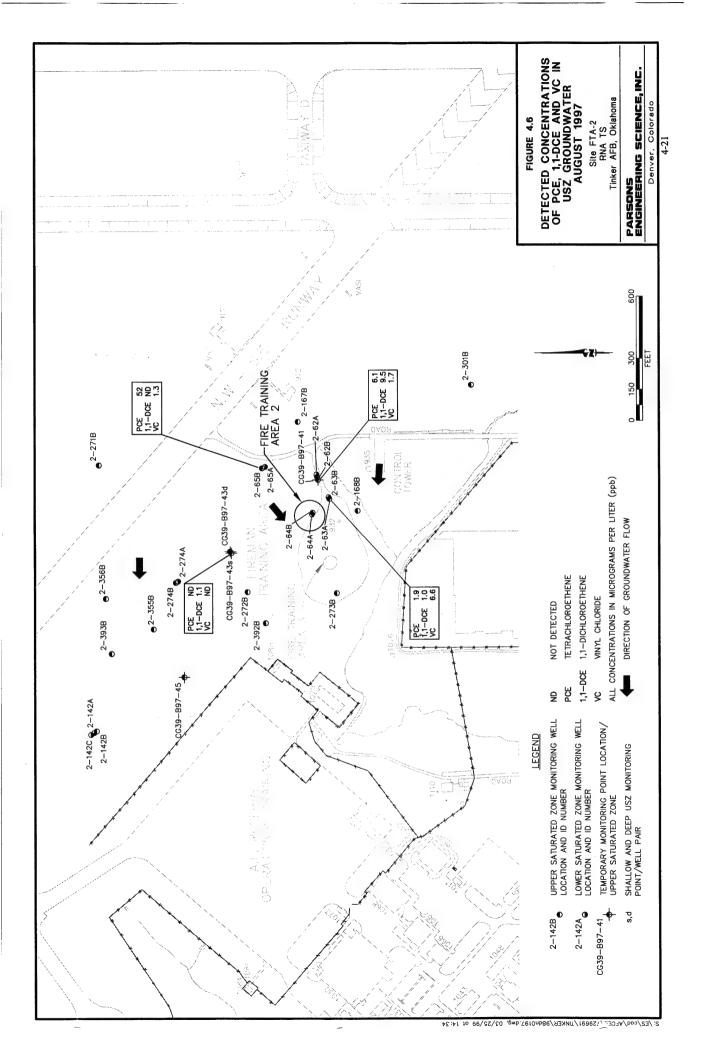
^{b/} Monitoring well 3-356B was sampled on 9/4/97 by Brown and Root.

4-17









sand interval. The clay layer that separates the upper and lower USZ sands is thinner and less continuous in the area of well locations 2-274B, 2-355B, and 2-393B relative to the source area near well location 2-62B.

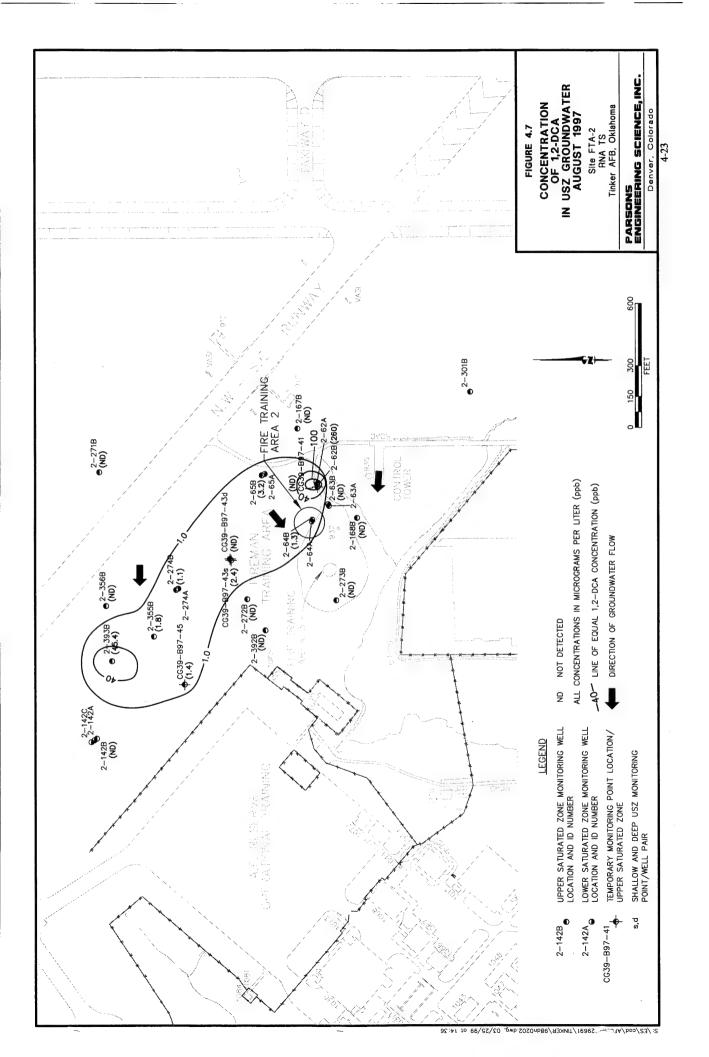
Trans-1,2-DCE was detected at seven locations at concentrations ranging from 2.1 to 77 μg/L (Figure 4.5). No concentrations exceeded the USEPA MCL of 100 μg/L. At all wells/points sampled, trans-1,2-DCE was detected at concentrations significantly lower than cis-1,2-DCE (Figures 4.4 and 4.5). This trend would be expected if TCE is being degraded to DCE via reductive dehalogenation. As noted before, under the influence of biodegradation, cis-1,2-DCE is a more common daughter product than the other DCE isomers. Furthermore, the trans isomer is more common when DCE is used for commercial applications. The presence of cis-1,2-DCE is therefore a good indicator that the initial step of TCE reductive dehalogenation is taking place in groundwater at FTA-2.

1,1-DCE was detected in samples from three of the 19 USZ groundwater monitoring wells at FTA-2 (Figure 4.6). Concentrations of 1.1 μ g/L, 9.5 μ g/L and 1.0 μ g/L were detected in samples from wells CG39-B97-43s, 2-62B and 2-63B, respectively. Only the concentration detected at 2-62B exceeds the MCL of 7 μ g/L. TCE and the other isomers of DCE were also detected in these wells suggesting that 1,1-DCE could have been produced by reductive dehalogenation of TCE.

VC was detected in samples from three of the 19 USZ groundwater monitoring wells sampled at FTA-2 (Figure 4.6). Concentrations of 1.7 μ g/L, 6.6 μ g/L and 1.3 μ g/L were detected in samples from 2-62B, 2-63B and 2-65B, respectively. Only the concentration detected at well 2-62B exceeds the MCL of 2 μ g/L. The distribution of VC is similar to that for PCE. In fact, VC was detected at all three locations where PCE was detected. The presence of VC in groundwater is also a good indicator of reductive dehalogenation. Because VC is typically a gas at most working temperatures [the boiling point of VC is about 13.5 degrees centigrade (°C) or 7 degrees Fahrenheit (°F) (Montgomery, 1996)], VC is used for plastic manufacturing instead of being used as a solvent. In addition, when liquid VC is spilled, most of the mass is likely to volatilize before reaching groundwater unless there is a large-volume release.

4.2.5 Dissolved Chlorinated Ethanes

Chlorinated ethanes detected in groundwater at FTA-2 include 1,1-DCA and 1,2-DCA (Table 4.3). Relatively low concentrations of 1,1-DCA were detected, with the maximum 1,1-DCA concentration of 4.6 μ g/L occurring at well 2-63B. There is no USEPA MCL listed for 1,1-DCA. The distribution of 1,2-DCA is shown on Figure 4.7, with 1,2-DCA detected in eight of the 19 USZ groundwater samples analyzed. The maximum concentrations of 1,2-DCA were 260 μ g/L at 2-62B and 49.2 μ g/L at 2-393B, both exceeding the MCL of 5 μ g/L. These two locations are at opposite ends of the overall CAH contaminant plume. Concentrations of 1,2-DCA in the intermediate area are substantially less, ranging from 1.1 to 3.2 μ g/L. 1,1-DCA is not commonly used as a solvent, whereas 1,2-DCA is used as a solvent. Therefore, the presence of 1,2-DCA rather than 1,1-DCA suggests that 1,2-DCA is a result of source release, rather than a biodegradation byproduct. Furthermore, the data indicate the possibility



of a secondary 1,2-DCA source located northwest of the FTA-2 fire pit in the vicinity of 2-393B.

4.2.6 Dissolved Chlorobenzenes

Chlorobenzenes detected in groundwater at FTA-2 include CB, 1,2-DCB, 1,3-DCB, and 1,4-DCB (Table 4.3, Figure 4.8). These compounds were detected in four wells associated with the general CAH source area (2-62B, 2-64B, and 2-65B), and the preferential plume migration pathway (CG39-B9743s). A CB and 1,2-DCB plume appears to be originating from the vicinity of 2-62B and migrating to the northwest. Detected CB concentrations ranged from 1.0 μ g/L in CG39-B97-43s, to 117 μ g/L in 2-62B. CB was also detected at a concentration of 9.5 μ g/L in 2-65B. Only the highest CB concentration detected in 2-62B exceeds the MCL for CB of 100 μ g/L. 1,2-DCB was detected in two wells at concentrations of 3.6 μ g/L in 2-65B, and 969 μ g/L in 2-62B. Only the 1,2-DCB concentration detected in 2-62B exceeds the MCL of 600 μ g/L. Similarly, 1,4-DCB exceeded its MCL of 75 μ g/L at well 2-62B. 1,3-DCB was not detected above its MCL of 600 μ g/L.

4.2.7 Other Dissolved Chlorinated Compounds

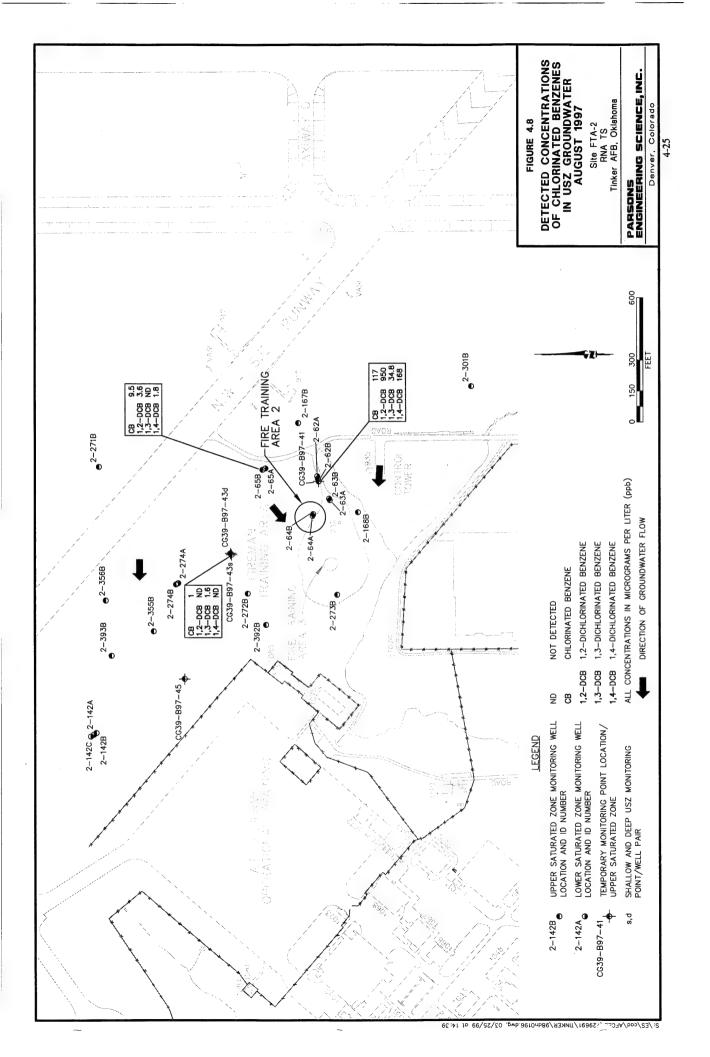
Chloroform was the only other chlorinated compound detected in groundwater at FTA-2. Chloroform was detected at concentrations of 1.0 μ g/L in CG39-B97-43s, and 5.3 μ g/L in 2-62B. These locations coincide with the maximum detected concentrations of TCE and cis-1,2-DCE. The detected concentrations do not exceed the MCL of 100 μ g/L. Chloroform may form as the end product of aerobic degradation of TCE through intermediate byproducts of 2,2,2-trichloroacetaldehyde and 2,2,2-trichloroethanol (Figure 4.2). Therefore, the formation of chloroform in association with the presence of TCE may indicate an aerobic biodegradation pathway from TCE to chloroform.

4.2.8 LSZ Groundwater Quality

The only contaminant detected in LSZ groundwater was TCE at a concentration of $10 \,\mu g/L$ in the groundwater sample collected from monitoring well 2-62A. The TS sampling event is the first time CAHs have been detected in LSZ groundwater at FTA-2. Monitoring well 2-62A is located beneath the primary source area, immediately adjacent to well 2-62B, the USZ well with the highest detected concentration of TCE in USZ groundwater at FTA-2. TCE is the most mobile (lowest retardation) of the CAHs detected at FTA-2, and its detection indicates that CAH contamination is starting to migrate vertically through the USZ/LSZ aquitard to the LSZ.

4.3 ANALYSIS OF HYDROCARBON BIODEGRADATION

Available information indicates that the CAH plume primarily exhibits Type 1 behavior in the immediate source area delimited by monitoring wells 2-63B, 2-168B, 2-62B, and 2-64B. Fuel hydrocarbons and chlorobenzenes remain in this area, and the electron donor supply may be adequate to allow continued microbial reduction of CAHs. The role of competing electron acceptors (e.g., DO, nitrate, ferric iron, and sulfate) is described later in this subsection. The Type 1 conditions are resulting in



extensive degradation of the highly chlorinated solvents PCE and TCE in the source area. Degradation of DCE may also be occurring, as evidenced by the presence of VC dissolved in groundwater in this area. The dehalogenation of DCE to VC and of VC to ethene occurs at progressively slower rates, because the rate of reductive dehalogenation decreases as the degree of chlorination decreases. DCE (particularly *cis*-1,2-DCE) is being produced from TCE and PCE faster than it can be dehalogenated to VC, explaining the abundance of DCE, the relative scarcity of VC, and the lack of ethene in site groundwater.

The downgradient portion of the CAH plume appears to primarily exhibit Type 3 behavior. There is little or no anaerobic reductive dehalogenation of TCE and DCE occurring; however, DCE may be aerobically degraded (oxidized). VC is being rapidly oxidized as it migrates into the more aerobic environment as evidenced by the general lack of VC detections north and west of the general source area. The rapid decrease, or lack of, chlorobenzenes in the downgradient portions of the plume may indicate that CB and DCB are being utilized as electron donors in the source or downgradient portion of the CAH plume.

In addition to the distribution of contaminants and daughter products, other evidence can be used to support the interpretation that biodegradation of fuel hydrocarbons and chlorinated solvents is ongoing at FTA-2. Analysis of temporal changes in contaminant concentrations over time may indicate field-scale contaminant mass loss as a result of biodegradation processes. By comparing concentrations and distributions of electron donors, electron acceptors, and byproducts of microbially mediated reactions, types of bidegradation processes operating at a site can be identified. Geochemical parameters, including redox potential, alkalinity, and other changes in groundwater chemistry can also provide supporting evidence and are presented in the following subsections.

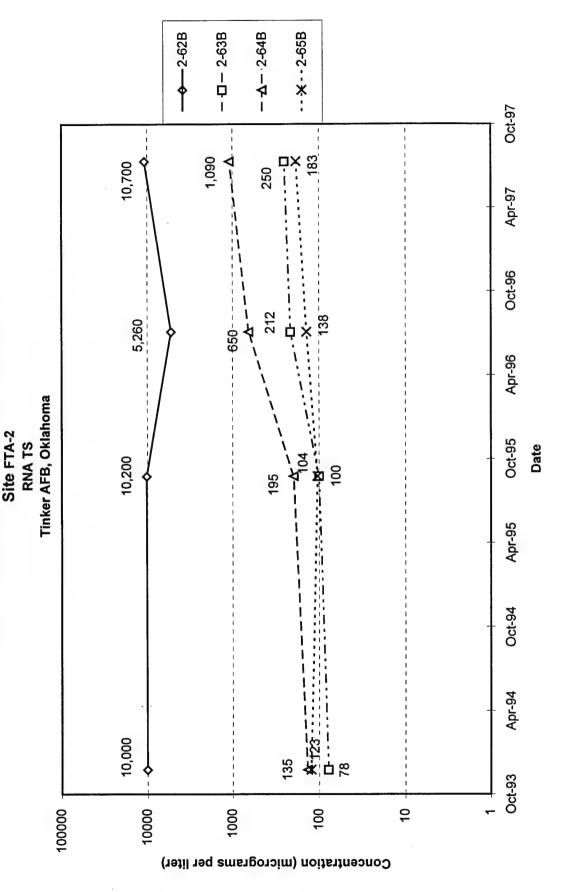
4.3.1 Field-Scale Contaminant Mass Loss

Groundwater quality data collected for four sampling events between December 1993 and August 1997 from monitoring wells 2-62B, 2-63B, 2-64B, and 2-65B were reviewed to assess temporal changes in contaminant concentrations over time. Decreasing contaminant concentration trends reflect natural attenuation trends. In a stable or attenuating contaminant plume, decreasing concentration trends may be used to support loss of contaminant mass through destructive processes such as biodegradation.

4.3.1.1 Chlorinated Ethenes Over Time

Available concentration-versus-time data for chlorinated ethenes in samples from source area wells 2-62B, 2-63B, 2-64B, and 2-65B were examined to reassess whether or not long-term trends in chlorinated ethene concentrations are apparent. Total chlorinated ethene concentrations at these wells are relatively stable (well 2-62B), or gradually increase over time (wells 2-63B, 2-64B, and 2-65B) (Figure 4.9). Review of this plot indicates that the source of the CAH contaminant plume has not attenuated over the period from 1993 to 1997.

FIGURE 4.9
TOTAL CHLORINATED ETHENES VERSUS TIME



Concentrations of TCE, cis-1,2-DCE, trans-1,2-DCE, and VC versus time for well 2-62B are shown on Figure 4.10. Well 2-62B has consistently exhibited the highest detected concentrations at the site, and it is assumed to represent dissolved contaminant concentrations in the primary source area. Concentrations at well 2-62B do not show a discernible trend, indicating that the source of chlorinated ethenes at FTA-2 has not attenuated. TCE concentrations have increased from 9.0 to 28 µg/L, and from 260 to 518 µg/L, respectively, at two wells further down the plume migration pathway (2-272B and 2-274B) from August 1996 and August 1997. An accurate assessment regarding the degree to which the chlorinated ethene plume is at equilibrium (neither expanding nor receding) cannot be made on the basis of the available data. However, given the indications of increasing concentrations at several wells, it is probable that the plume is expanding.

4.3.1.2 Chlorinated Ethanes Over Time

Concentration-versus-time data for chlorinated ethanes detected in samples from wells 2-62B, 2-63B, 2-64B, and 2-65B were examined to reassess whether or not long-term trends in chlorinated ethanes concentrations are apparent. Total chlorinated ethane concentrations increase through time at wells 2-63B, 2-64B, and 2-65B; however, total chlorinated concentrations decrease through time at well 2-62B (Figure 4.11). Results of this analysis are inconclusive, but the data from source area well 2-62B may indicate slight attenuation of chlorinated ethanes over the period from 1993 to 1997.

4.3.1.3 Chlorobenzenes Over Time

Available concentration-versus-time data for chlorobenzenes in samples from source area wells 2-62B, 2-64B, and 2-65B were examined to reassess whether or not long-term trends in chlorobenzene concentrations are apparent. Total chlorobenzene concentrations at these wells are erratic (well 2-62B) or gradually increasing over time (wells 2-64B and 2-65B) (Figure 4.12). Results of this analysis indicate that the source of the chlorobenzene contaminant plume has not significantly attenuated over the period from 1993 to 1997.

4.3.2 Presence of Daughter Products and CAH Ratios

The presence of daughter products that were not used in Base operations, particularly cis-1,2-DCE and VC, provides strong evidence that the parent solvents (TCE and/or PCE) are being reductively dehalogenated. However, the anaerobic dehalogenation processes occurring in FTA-2 groundwater are limited to the primary source area.

If reductive dehalogenation of TCE were occurring along the contaminant flow path, then the mass fraction of TCE would be expected to decrease relative to that of DCE and VC. However, the molar ratio of TCE relative to *cis*-1,2-DCE does not decrease with distance from the source area (well 2-62B), remaining relatively constant at approximately 85 percent TCE of the TCE plus *cis*-1,2-DCE molar concentration (Figure 4.13). The rapid decrease in VC with distance from the source area (Table 2.3) indicates that, beyond the source area, VC is used as an electron donor and is aerobically degraded.

TCE, CIS-1,2-DCE, TRANS-1,2-DCE, AND VC VERSUS TIME AT 2-62B FIGURE 4.10 Site FTA-2

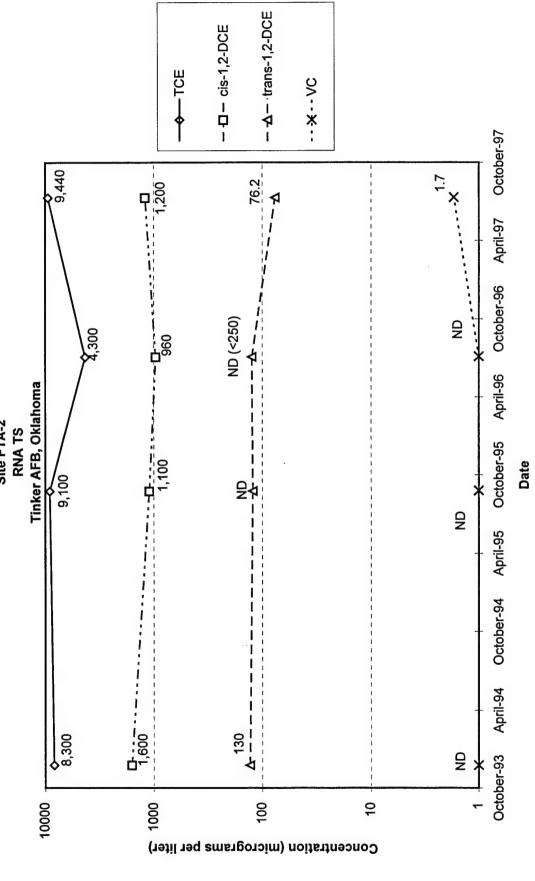


FIGURE 4.11
TOTAL CHLORINATED ETHANES VERSUS TIME
Site FTA-2

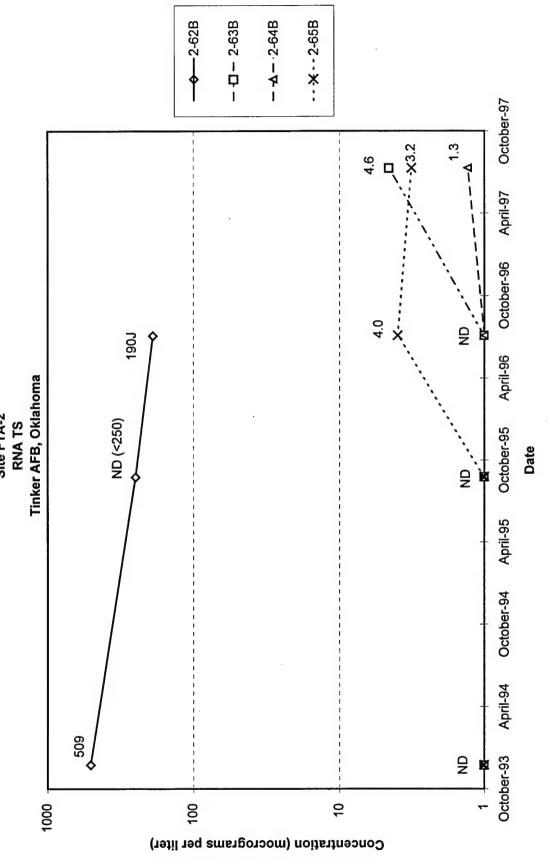


FIGURE 4.12
TOTAL CHLOROBENZENES VERSUS TIME
Site FTA-2

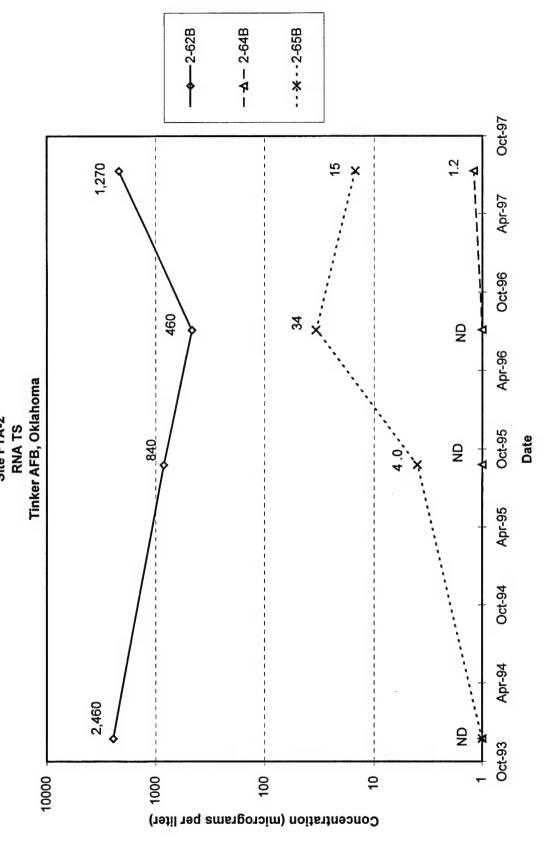
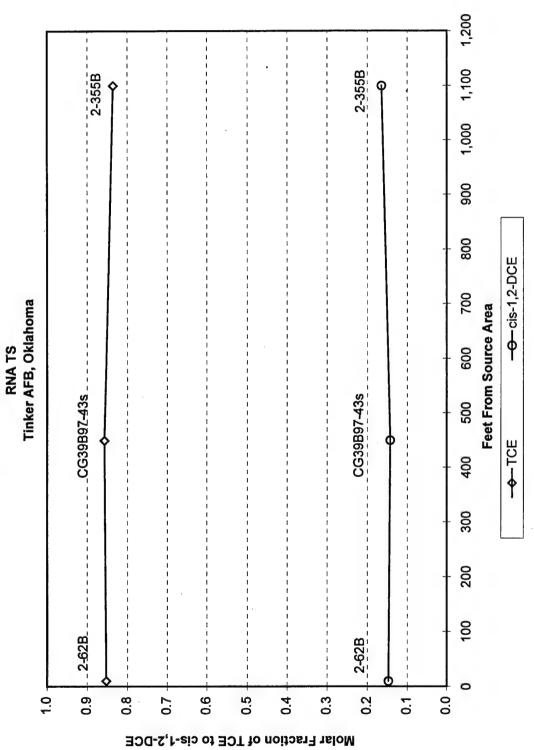


FIGURE 4.13
MOLAR FRACTION OF TCE TO cis-1,2-DCE VERSUS DISTANCE
Site FTA-2



The geochemical conditions conducive to aerobic degradation of VC do not at the same time support anaerobic reductive dechlorination of TCE or DCE. DCE is most likely also aerobically degraded, but at a slower rate than VC.

Data from wells located near the source area (2-62B, 2-63B, 2-64B, and 2-65B), and along the approximate CAH plume axis (CG39B97-43s and 2-355B) were used to assess the magnitude of the *cis*-1,2-DCE to *trans*-1,2-DCE ratio throughout the CAH plume. Because the *cis*- isomer is preferentially produced by microbial reductive dehalogenation, the ratio of *cis*-1,2-DCE to *trans*-1,2-DCE should be relatively high, especially in areas where the parent solvent(s) is being transformed (source area). The reported concentrations of *trans*-1,2-DCE at well 2-355B were less than 1 µg/L, so the concentration was assumed to be 0.5 µg/L for calculation purposes. The magnitude of the *cis*-1,2-DCE to *trans*-1,2-DCE ratio varies from 8.75 at well 2-64B to approximately 100 at well CG39B97-43s, indicating that the *cis*- isomer predominates throughout the plume. These ratios are a strong indication that TCE is being reductively dehalogenated, at least in the source area. According to Weidemeier, *et al.* (1996), if *cis*-1,2-DCE is greater than 80 percent of total DCE present, then it is a likely daughter product of TCE. In all cases at FTA-2, *cis*-1,2-DCE is present at greater than 80 percent of total DCE.

4.3.3 Chloride as an Indicator of Dehalogenation

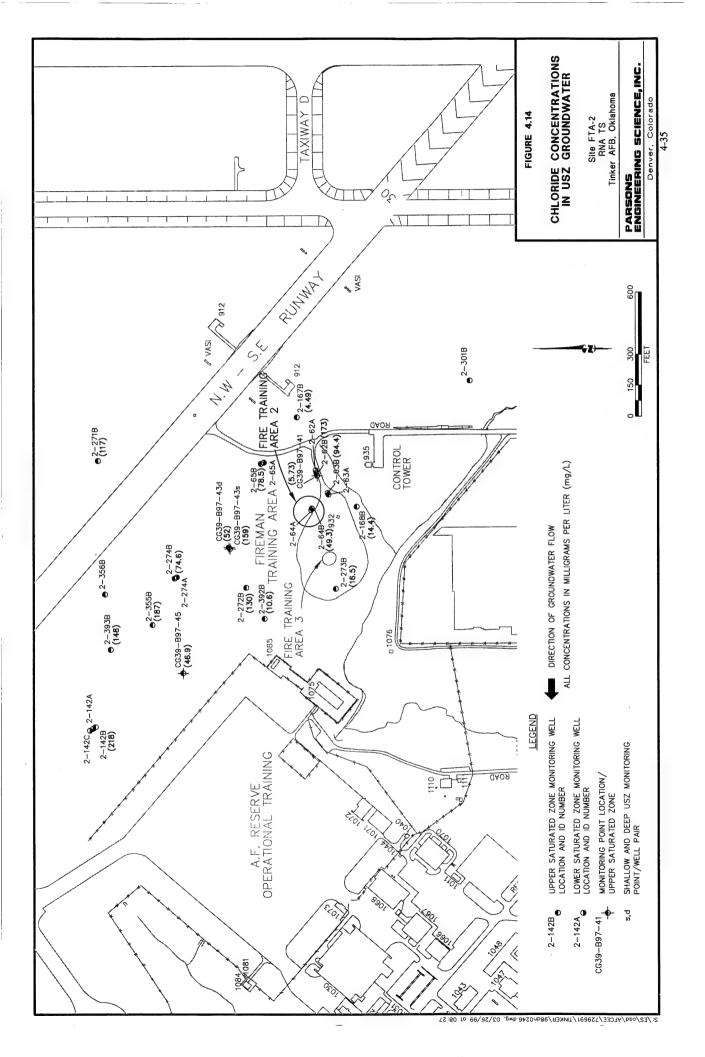
Chloride ions are removed from chlorinated solvents and enter solution during biodegradation, whether via reductive dehalogenation or aerobic oxidation. Therefore, chloride concentrations in groundwater should increase above background levels in areas where reductive dehalogenation or aerobic biodegradation are taking place. Chloride concentrations are presented in Table 4.4, and the distribution of chloride in USZ groundwater at the site is shown on Figure 4.14. Background chloride concentrations (measured in wells without contamination in USZ groundwater) range from about 4.49 milligrams per liter (mg/L) (2-167B) to about 17 mg/L (2-273B), with two exceptions being 117 mg/L at 2-271B and 218 mg/L at 2-142B. It is not clear why these higher chloride concentrations occur at wells 2-271B and 2-142B, although these wells are located farther to the north, and chloride concentrations may represent upgradient or cross-gradient water quality. Most background chloride concentrations are less than 20 mg/L.

Chloride concentrations within the boundaries of the chlorinated solvent plume are slightly elevated above background, with a maximum concentration of 187 mg/L at 2-355B. Within areas where biodegradation is indicated by other chemical data, most chloride concentrations are greater than 40 mg/L. The slightly elevated chloride concentrations observed at FTA-2 are a moderate indication that CAHs are undergoing microbially mediated degradation when reviewed in conjunction with the distributions of CAHs (Figures 4.4 through 4.6). Given the presence of daughter products such as cis-1,2-DCE, trans-1,2-DCE, and VC, it is likely that some of the chloride is a product of reductive dehalogenation, at least within the primary source area. Chloride concentrations within the dissolved CAH plume may also be related to aerobic degradation of DCE and VC.

TABLE 4.4 GROUNDWATER GEOCHEMICAL DATA AUGUST 1997 FTA-2 RNA TS TINKER AFB, OKLAHOMA

Sample				Dissolved	Redox	Total			Ferrous	Nitrite +		Hydrogen		Total	Carbon				
Location	Temp		Conductivity	Oxygen	=	Alkalinity		Chloride Ammonia	Iron	Nitrate	Sulfate	Sulfide	Manganese	Phenols	Dioxide	TOC	Methane	Ethane	Ethene
	(°C),	pH _b /	(μs/cm) ^{c/}		(mV) ^{e/}	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)
Upper Saturated Zone	d Zone																		
CG39-B97-41	22.0	88.9	861	3.9	136.2	520	5.73	<0.05	< 0.01	2.01	15.4	< 0.01	< 0.01	< 0.01	09	<0.4	< 0.001	< 0.003	< 0.002
CG39-B97-43s	21.5	08.9	1278	4.8	116.7	330	159	<0.05	<0.01	4.22	49.0	< 0.01	< 0.01	0.2	48	0.539	< 0.001	< 0.003	< 0.002
CG39-B97-43d	18.4	7.07	1112	3.3	129.8	450	52.0	< 0.05	< 0.01	2.55	44.1	< 0.01	< 0.01	0.2	75	0.545	< 0.001	< 0.003	< 0.002
CG39-B97-45	19.4	6.91	872	6.5	103.2	380	46.9	<0.05	< 0.01	4.80	11.3	< 0.01	< 0.01	< 0.01	75	<0.4	< 0.001	< 0.003	< 0.002
2-392B	18.5	7.18	697	0.9	138.7	350	9.01	<0.05	< 0.01	2.57	5.91	< 0.01	< 0.01	0.2	09	<0.4	< 0.001	< 0.003	< 0.002
2-393B	18.2	7.23	1184	7.0	110.9	190	148	< 0.05	< 0.01	3.70	14.1	< 0.01	< 0.01	< 0.01	110	<0.4	< 0.001	< 0.003	< 0.002
2-62B	18.8	7.34	1441	0.2	50.7	260	173	<0.05	<0.01	3.92	127	< 0.01	< 0.01	< 0.01	55	3.30	1.05	< 0.003	< 0.002
2-63B	19.9	7.35	1516	9.0	79.4	460	4.46	<0.05	<0.01	2.39	65.7	< 0.01	< 0.01	< 0.01	75	11.2	0.002	< 0.003	< 0.002
2-64B	20.6	NR ^{t/}	848	1.8	123.5	260	49.3	<0.05	< 0.01	3.42	27.0	< 0.01	< 0.01	< 0.01	40	0.795	< 0.001	< 0.003	< 0.002
2-65B	18.8	7.13	1166	2.3	91.2	350	78.5	<0.05	<0.01	3.18	44.4	< 0.01	< 0.01	< 0.01	75	1.96	0.002	< 0.003	< 0.002
2-167B	17.3	7.11	1008	2.3	9.08	999	4.49	<0.05	0.1	0.27	18.1	< 0.01	< 0.01	0.1	75	0.488	0.022	< 0.003	< 0.002
2-168B	18.3	7.17	993	0.7	-21.7	380	14.4	<0.05	< 0.01	0.58	42.4	< 0.01	< 0.01	< 0.01	99	3.01	0.02	< 0.003	< 0.002
2-271B	16.9	7.19	1238	3.8	8.79	420	117	0.15	9.0	1.11	29.7	< 0.01	< 0.01	0.3	65	0.939	< 0.001	< 0.003	< 0.002
2-272B	17.8	7.24	1075	4.1	30.2	280	130	< 0.05	< 0.01	1.95	37.6	< 0.01	< 0.01	< 0.01	45	0.448	< 0.001	< 0.003	< 0.002
2-273B	18.3	7.20	761	0.2	-105.1	280	16.5	<0.05	0.2	0.35	12.1	< 0.01	< 0.01	< 0.01	45	0.638	0.005	< 0.003	< 0.002
2-274B	17.7	7.20	919	5.1	50.0	310	74.6	< 0.05	0.2	3.08	24	< 0.01	< 0.01	0.2	45	<0,4	< 0.001	< 0.003	< 0.002
2-355B	17.8	7.23	1490	7.9	116.7	380	187	< 0.05	0.4	3.28	48.8	< 0.01	< 0.01	0.3	50	0.741	< 0.001	< 0.003	< 0.002
2-142 B	16.9	7.36	1480	3.1	52.6	380	218	< 0.05	8.0	2.00	76.1	< 0.01	< 0.01	0.2	55	0.929	< 0.001	<0.003	< 0.002
Lower Saturated Zone	d Zone																		
2-62A	21.5	7.42	782	5.6	97.2	310	7.54	< 0.05	0.1	4.02	6.51	< 0.01	<0.01	<0.01	45	1.53	< 0.001	< 0.003	< 0.002
2-63A	18.9	7.31	654	7.2	47.8	380	20.5	< 0.05	1.0	1.91	8.46	< 0.01	< 0.01	< 0.01	20	0.745	< 0.001	< 0.003	< 0.002
2-64A	20.2	69.7	999	6.1	121.9	270	21.5	< 0.05	0.1	1.29	7.96	< 0.01	<0.01	< 0.01	45	0.152	< 0.001	< 0.003	< 0.002
2-65A	19.9	79.7	576	4.2	-38.3	190	25.6	< 0.05	0.3	0.72	68.6	< 0.01	< 0.01	<0.01	25	<0.4	< 0.001	< 0.003	< 0.002
2-274A	20.1	7.60	516	7.4	63.1	210	18.0	< 0.05	< 0.01	0.62	10.4	< 0.01	< 0.01	0.4	25	<0.4	< 0.001	<0.003	< 0.002
2-142A	17.5	7.75	620	1.2	26.3	270	28.4	< 0.05	< 0.01	< 0.05	9.72	< 0.01	< 0.01	0.2	45	2.14	< 0.001	< 0.003	< 0.002
,																			

a/°C = degrees Centigrade.
b/ well head measurement not made; value reported is measurement made in on-base field laboratory.
c/ µs/cm = microsiemens per centimeter.
d/ mg/L = milligrams per liter.
e/ mV = millivolts.
f/ NR = not recorded.



4.3.4 ORP as an Indicator of Redox Processes

Redox potential is a measure of the relative tendency of a solution to accept or transfer electrons. The redox potential of a groundwater system depends on which electron acceptors are being reduced by microbes during oxidation of organic compounds. For example, by coupling the oxidation of fuel hydrocarbon compounds (or native organic carbon), which requires energy, to the reduction of other compounds (e.g., oxygen, nitrate, manganese, ferric iron, sulfate, carbon dioxide, and possibly *cis*-1,2-DCE), which yields energy, the overall reaction will yield energy.

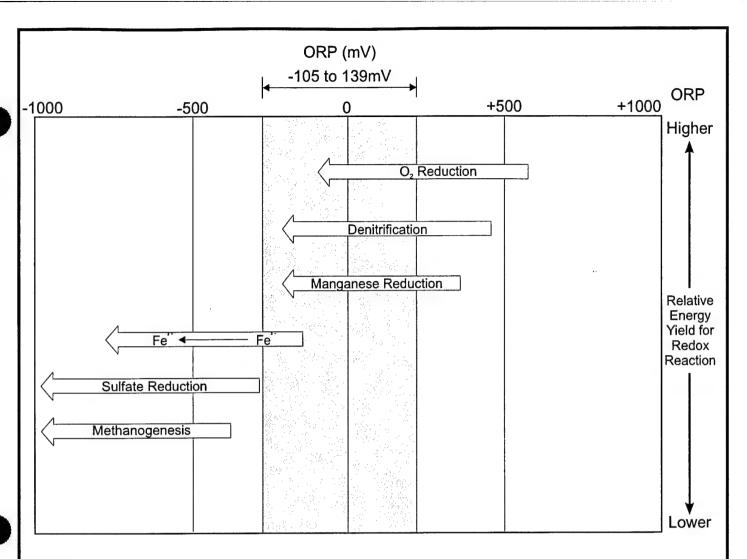
The sequence of microbially mediated redox processes and the approximate ranges of ORPs that are favorable for each process are presented in Figure 4.15. In general, reactions yielding more energy tend to take precedence over processes that yield less energy (Stumm and Morgan, 1981; Godsey, 1994; Reinhard, 1994). Oxygen reduction would be expected in an aerobic environment with microorganisms capable of aerobic respiration, because oxygen reduction yields significant energy (Bouwer, 1992; Chapelle, 1993). However, once the available DO is depleted and anaerobic conditions dominate the interior regions of a contaminant plume, anaerobic microorganisms can utilize other electron acceptors in the following order of preference: nitrate (denitrification), manganese (manganese reduction), ferric iron (iron reduction), sulfate (sulfate reduction), and finally carbon dioxide (methanogenesis). Each successive redox reaction provides less energy to the system, and each step down in redox energy yield is paralleled by an ecological succession of microorganisms capable of facilitating the pertinent redox reactions.

Redox potentials were measured at groundwater monitoring wells and points in August 1997. Redox potentials for USZ groundwater at the site range from 138.7 millivolts (mV) to -105.1 mV, and from 121.9 mV to -38.3 mV in the LSZ (Table 4.4 and Figure 4.16). In some locations, low redox potentials coincide with areas of elevated solvent concentrations, decreased DO, nitrate, and sulfate concentrations, and elevated ferrous iron (iron [II]) and methane concentrations. However, this is not always the case and the correlation between redox potential and geochemical indicators is not strong.

The lowest redox potentials were detected in the vicinity of 2-273B and 2-168B. The redox potentials measured at those two locations were -105.1 mV and -21.7 mV, respectively. This area is immediately downgradient of the FTA-2 fire pit and the CAH source area (2-62B). This area of low redox potential also coincides with Fire Training Area 3, and may reflect reducing conditions caused by biodegradation of hydrocarbons released by historical practices at either site.

4.3.5 Electron Donors

The distribution of potential electron donors (carbon sources) such as BTEX, native organic carbon, or CAHs themselves are useful for evaluating the feasibility of reductive dehalogenation, aerobic biodegradation, or cometabolism (i.e., biodegradation reactions involving another substrate).



Notes

ORP = Oxidation Reduction Potential

Range of ORP measured at FTA-2

- 1. These reactions would be expected to occur in sequence if the system is moving toward equilibrium.
- 2. These redox processes occur in order of their energy-yielding potential (provided microorganisms are available to mediate a specific reaction). Reduction of a highly oxidized species decreases the ORP of the system.
- 3. The ORP of the system determines which electron acceptors are available for organic carbon oxidation.
- 4. Redox reaction sequence is paralleled by an ecological succession of biological mediators.

FIGURE 4.15

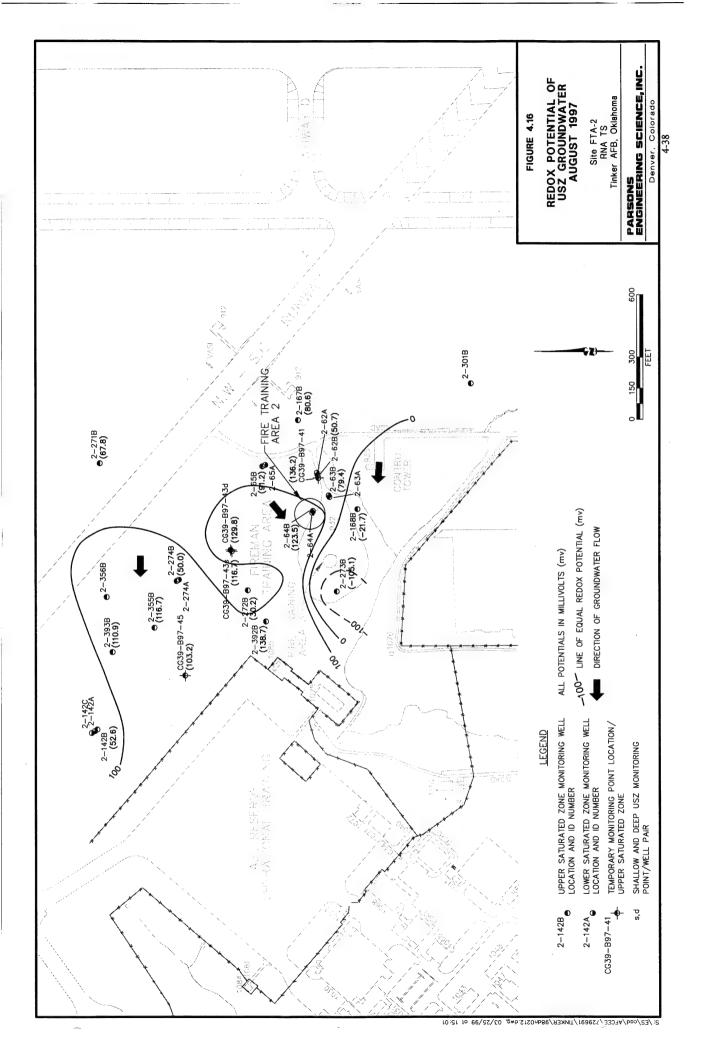
SEQUENCE OF MICROBIALLY MEDIATED REDOX PROCESSES USZ GROUNDWATER

Site FTA-2 RNA TS Tinker AFB, Oklahoma

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

Adapted from Stumm and Morgan, 1981.



4.3.5.1 BTEX and Total Fuel Carbon in Groundwater

The presence of BTEX and fuel carbon in the same area as a CAH plume creates favorable conditions for reductive dehalogenation, because the petroleum compounds provide a source of electron donors and facilitate microbial reactions that lower the groundwater ORP. It is likely that the introduction of petroleum hydrocarbons due to fire training and other waste disposal activities stimulated microbial activity and reducing conditions ensued. Given these conditions, the presence of BTEX and fuel carbon in site groundwater is considered a favorable factor in the biodegradation of CAHs.

The coincidence of the dissolved BTEX and fuel carbon plumes with the highest concentrations of the reductive dehalogenation daughter products 1,1-DCE and VC (presented in Sections 4.3 and 4.4.1) establishes the importance of petroleum hydrocarbons in the effective reductive dehalogenation of chlorinated compounds near the general source area. However, BTEX and fuel carbon concentrations are much lower than CAH concentrations, and the extent and concentrations of the BTEX/fuel carbon plumes are such that detectable concentrations of these compounds do not appear to have significantly migrated beyond the general source area of the CAH contaminant plumes.

4.3.5.2 Organic Carbon in Groundwater

Dissolved organic carbon can also act as a source of electron donors during the reductive dehalogenation of CAHs. Dissolved TOC concentrations can be used as an indicator of the presence of such native carbon compounds in wells outside of the area containing dissolved contamination (anthropogenic organic compounds, such as CAHs, petroleum hydrocarbons, or landfill leachate also are measured by the TOC analytical method). Dissolved TOC concentrations in excess of 20 mg/L are desirable to drive dehalogenation reactions (Wiedemeier et al., 1996a).

TOC dissolved in groundwater was measured in groundwater samples collected in August 1997 (Table 4.4). Dissolved TOC concentrations detected in shallow groundwater at FTA-2 range from less than 0.4 mg/L to 11.2 mg/L. The highest concentrations were measured in 2-63B and 2-62B, and may reflect the presence of contamination. Elsewhere, the TOC concentrations did not exceed 3.0 mg/L. Native organic carbon concentrations of this magnitude are insufficient to drive reductive dehalogenation of CAHs.

4.3.5.3 Use of CAHs as Electron Donors

Less-chlorinated CAHs (e.g., DCE and VC) can be used as electron donors in aerobic environments, resulting in biodegradation of CAHs. Shallow groundwater throughout most of the FTA-2 contaminant plume is relatively aerobic (DO >1 mg/L), and the silty to sandy alluvial sediments do not have a particularly high native organic carbon content (Table 4.1). In these areas, TCE, DCE, and VC are probably being degraded through use as electron donors in microbially mediated redox reactions.

4.3.6 Alternate Electron Acceptors and Metabolic Byproducts

Biodegradation of organic compounds, whether natural or anthropogenic, brings about measurable changes in the chemistry of groundwater in the affected area. Concentrations of compounds used as electron acceptors (e.g., DO, nitrate, and sulfate) are depleted, and byproducts of electron acceptor reduction (e.g., iron (II), methane, and sulfide) are enhanced. By measuring these changes, it is possible to evaluate the importance of natural attenuation at a site.

Results of alternate electron acceptor and metabolic byproduct analyses for the TS are presented in Table 4.4. The following paragraphs discuss those parameters most useful in evaluating site biodegradation processes.

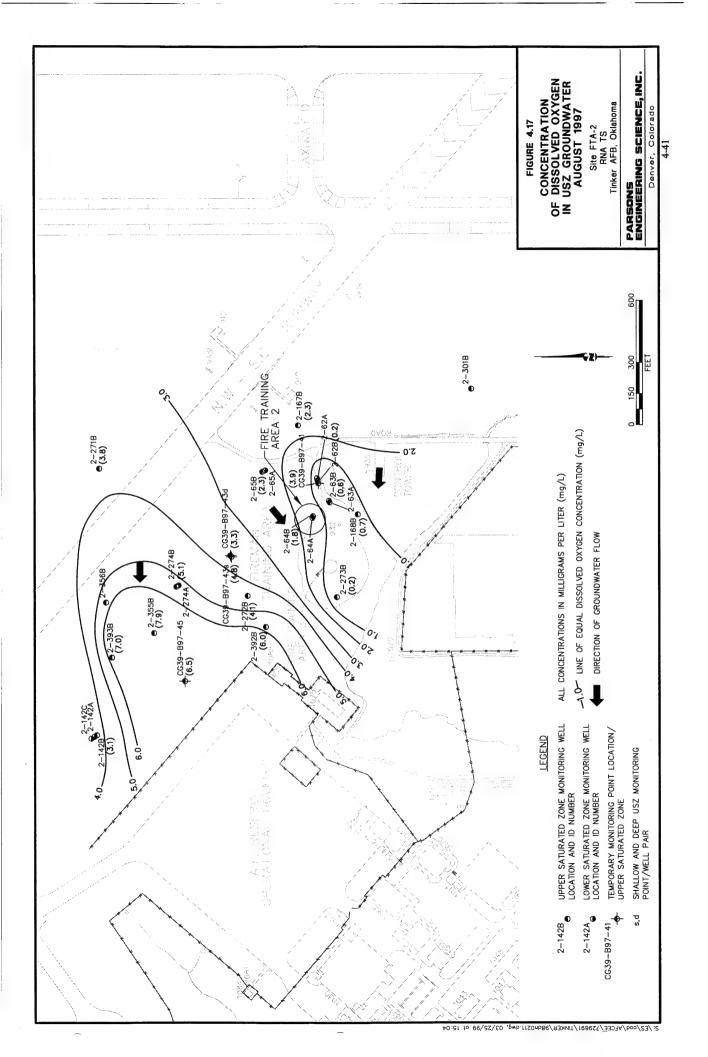
4.3.6.1 Dissolved Oxygen

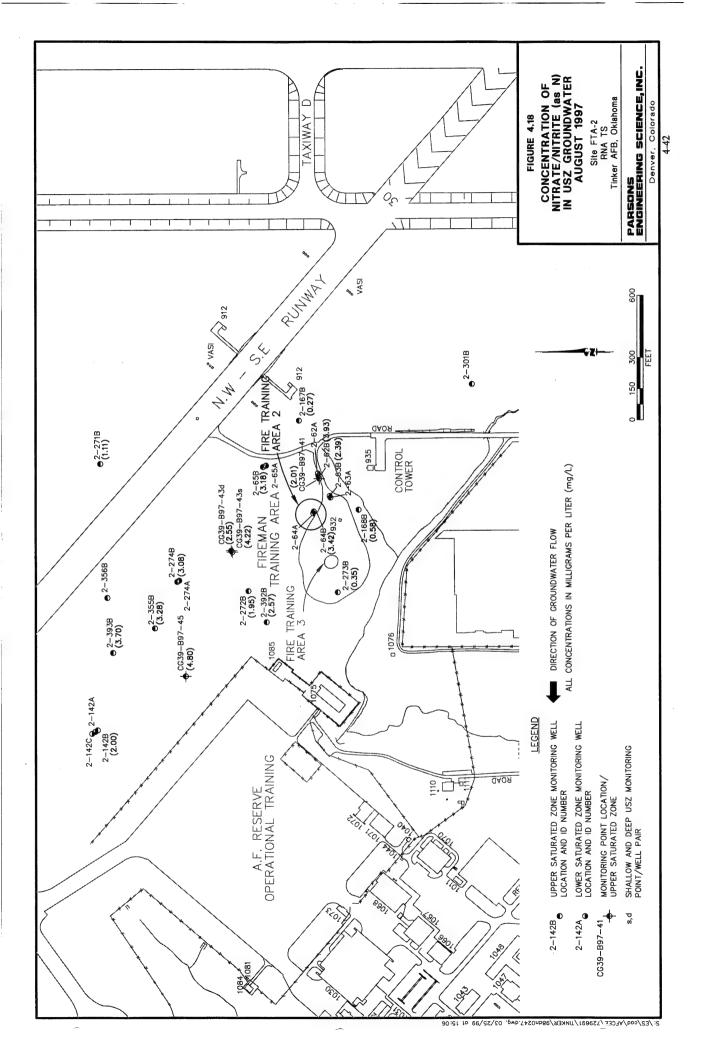
DO concentrations were measured at monitoring wells and points during the August 1997 sampling event. Concentrations ranged from 0.2 to 7.9 mg/L (Table 4.4), with the highest concentrations measured in wells/points in the furthest downgradient portion of the USZ CAH contaminant plume, or outside of the area of the contaminant plumes. In the downgradient portion of the contaminant plume, DO concentrations in excess of 6.0 mg/L may be a result of extensive well development pumping in newly installed monitoring wells. Actual background DO concentrations may be on the order of 3 to 4 mg/L (2-271B and 2-148B). DO concentrations in the LSZ range from 1.2 to 7.4 mg/L. Within and immediately hydraulically downgradient of the source area, concentrations were generally below 1.0 mg/L (Figure 4.17). The low concentrations of oxygen in the plume source area relative to upgradient and crossgradient concentrations are a strong indication of biological activity. Depletion of DO within the CAH source area at FTA-2 indicates that microbial activity stimulated by the presence of anthropogenic hydrocarbons (e.g., fuel hydrocarbons, chlorobenzenes, VC, DCA, or DCE) is consuming DO and creating anaerobic conditions.

4.3.6.2 Nitrate/Nitrite

After DO has been depleted in the microbiological treatment zone, nitrate may be used as an electron acceptor for anaerobic biodegradation of organic carbon via denitrification. Nitrate concentrations below background in areas with high organic carbon concentrations and low DO are indicative of denitrification. The oxidation of organic carbon via the process of denitrification (using nitrate as an electron acceptor) yields a relatively large amount of free energy to microbial populations.

Concentrations of nitrate/nitrite [as nitrogen (N)] were measured in groundwater samples collected in August 1997. Measured nitrate/nitrite (as N) concentrations are summarized in Table 4.4, and the distribution of nitrate/nitrite (as N) concentrations in site groundwater is shown on Figure 4.18. Background concentrations in the USZ range from 0.27 mg/L to 4.8 mg/L. Within the plume, nitrate/nitrite concentrations are generally range from 2 to 3 mg/L. An area of reduced nitrate/nitrite concentrations exists immediately cross-gradient of the CAH plume source area (or hydraulically downgradient) at wells 2-168B and 2-273B, which also coincides with an area of low ORP and low DO concentrations. These observations may indicate that DO has been





consumed, and microbes are utilizing nitrate as an electron acceptor for degradation of organic material (such as fuel hydrocarbons or native organic carbon).

4.3.6.3 Ferrous Iron

The reduction of ferric iron [iron (III)] has been shown to be a major metabolic pathway for some microorganisms (Lovley and Phillips, 1988; Chapelle, 1993). The reduction of ferric iron results in the formation of ferrous iron [iron (II)], and elevated concentrations of iron (II) often are found in anaerobic groundwater systems. Ferrous iron concentrations once were attributed to the abiotic spontaneous and reversible reduction of ferric oxyhydroxides, which are thermodynamically unstable in the presence of organic compounds such as BTEX, TMB, and naphthalene. However, recent evidence suggests that the reduction of ferric iron cannot proceed at all without microbial mediation (Lovley and Phillips, 1988; Lovley et al., 1991; Chapelle, 1993). None of the common organic compounds found in low-temperature, neutral, reducing groundwater could reduce ferric oxyhydroxides to iron (II) under sterile laboratory conditions (Lovley et al., 1991). This means that the reduction of ferric iron requires mediation by microorganisms with the appropriate enzymatic capabilities.

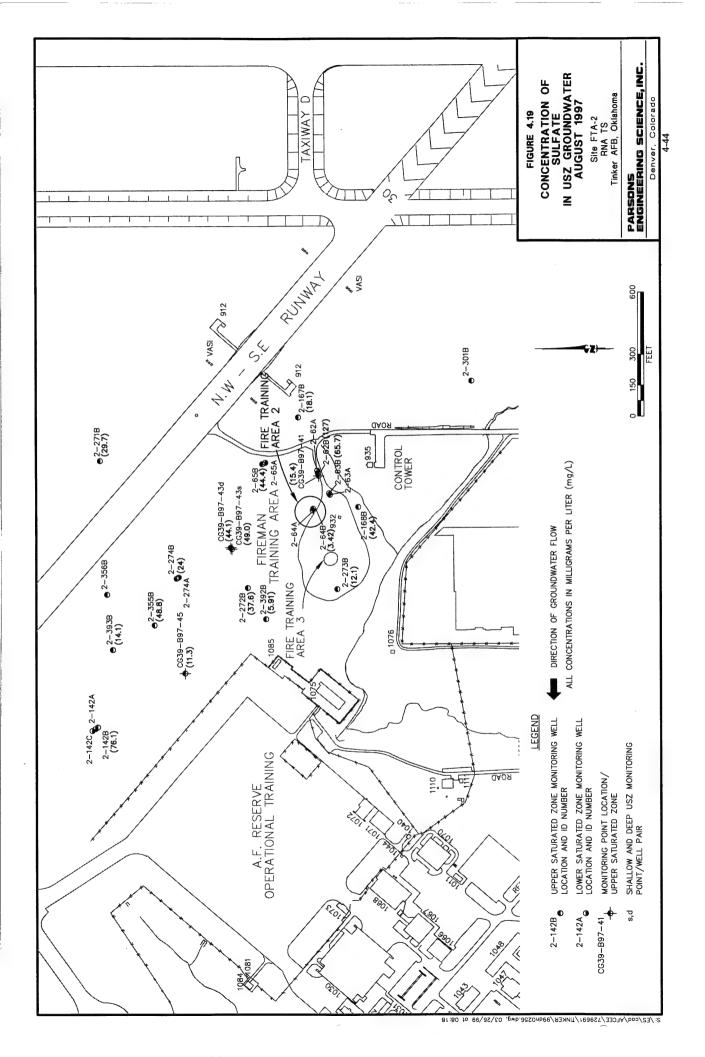
Iron (II) concentrations measured in groundwater samples collected in August 1997 are summarized in Table 4.4. Measured iron (II) concentrations in the USZ range from less than 0.01 mg/L to 0.8 mg/L, and from less than 0.01 mg/L to 1.0 mg/L in the LSZ. The highest iron (II) concentrations do not coincide with the solvent plume (2-271B, 2-142B, and 2-63A). This suggests that iron (III) hydroxide is not being reduced to iron (II) during biodegradation of native organic carbon, fuel hydrocarbons, or less-chlorinated solvents. Background levels of iron (II) in both the USZ and LSZ appear to range from below detection limits (0.01 mg/L) to about 1.0 mg/L, as measured at upgradient and cross-gradient wells.

4.3.6.4 Sulfate

Sulfate also may be used as an electron acceptor during microbial degradation of natural or anthropogenic organic carbon under anaerobic conditions (Grbic-Galic, 1990). This redox reaction is commonly called sulfate reduction. Sulfate concentrations were measured in groundwater samples collected in August 1997. Sulfate concentrations in USZ groundwater at the site ranged from 5.91 mg/L to 127 mg/L, and from 6.51 mg/L to 10.4 mg/L in the LSZ (Table 4.4). The highest sulfate concentration of 127 mg/L was detected in well 2-62B (Figure 4.19). Outside of the chlorinated solvent plume, sulfate concentrations ranged from 5.91 mg/L (2-392B) to 76.1 mg/L (2-242B). Within the plume area, nearly all sulfate concentrations ranged from 11.3 mg/L to 49 mg/L. Areas of decreased sulfate concentrations do not coincide with chlorinated solvent concentrations. It is therefore unlikely that sulfate reduction is an ongoing anaerobic biodegradation process at FTA-2.

4.3.6.5 Methane, Ethane and Ethene in Groundwater

Although anaerobic degradation may occur under nitrate- and sulfate-reducing conditions (Vogel et al., 1987; Chapelle, 1996), the most rapid biodegradation rates occur under methanogenic conditions (Bouwer, 1994). Methane, ethane, and ethene



concentrations were measured in groundwater samples collected in August 1997 (Table 4.4). Methane was detected at six locations, with concentrations ranging from less than 0.001 mg/L to 1.05 mg/L (2-62B). Ethane and ethene were not detected. The area of elevated methane concentrations coincides with the fuel hydrocarbon and chlorinated solvent source area. The presence of methane in the source area indicates highly reducing conditions favorable for oxidation of petroleum hydrocarbons and native organic matter as well as reductive dehalogenation of chlorinated solvents.

4.3.6.6 Ammonia/Ammonium

The presence of ammonia in groundwater can result from either nitrate reduction (facilitated by microbes) or fixing of atmospheric nitrogen (also a microbial process). Because fixation of atmospheric nitrogen only occurs under reducing conditions [ORP less than -500 mV (Stumm and Morgan, 1981)], ammonia production via nitrate reduction is probably more common than by fixation of atmospheric nitrogen. In either case, the presence of ammonia in groundwater is a strong indication of microbial activity. Ammonia/ammonium was detected at one location (2-271B at 0.15 mg/L) in USZ groundwater at a concentration greater than 0.05 mg/L (Table 4.4). Because nitrate is not widespread in USZ groundwater or present at any significant concentrations, it is unlikely that biodegradation by nitrate reduction is occurring at the site outside of the source area. Therefore, ammonia/ammonium is not expected at elevated concentrations.

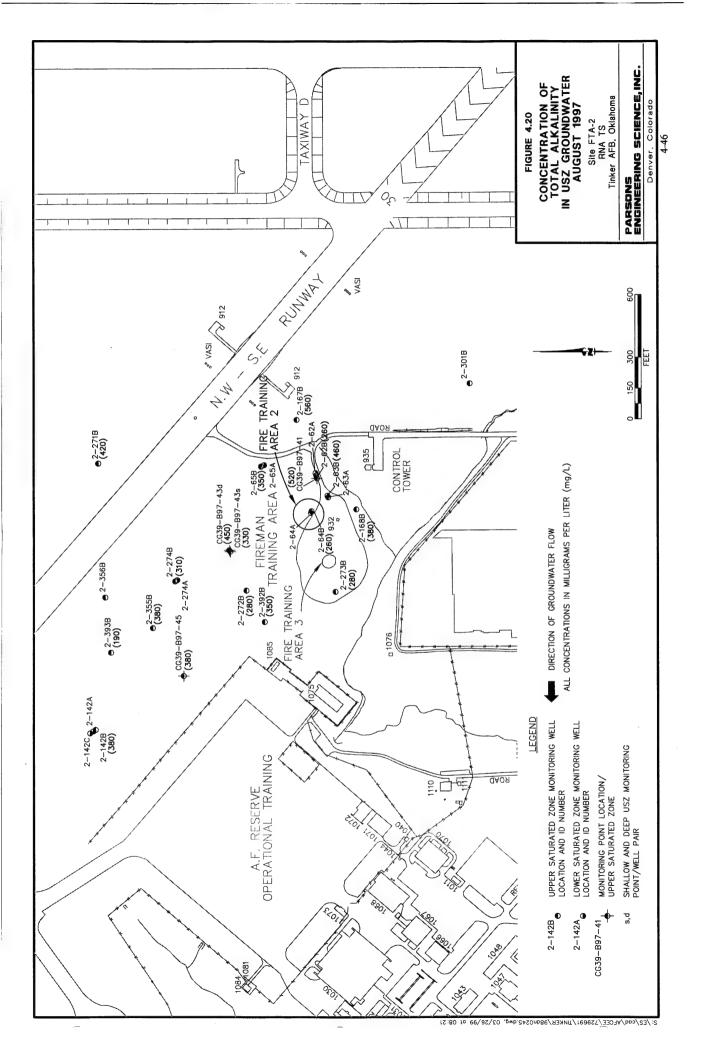
4.3.7 Additional Geochemical Indicators

Other geochemical data collected for this evaluation can be used to further interpret and support the contaminant, electron donor, electron acceptor, and byproduct data previously discussed. These parameters provide additional qualitative indications of what processes may be operating at the site.

4.3.7.1 Alkalinity and Carbon Dioxide Evolution

Carbon dioxide is produced during the biodegradation of fuel hydrocarbons and native organic carbon compounds. In aquifers that have carbonate minerals as part of the matrix, carbon dioxide forms carbonic acid, which dissolves these minerals, increasing the alkalinity of the groundwater. An increase in alkalinity (measured as calcium carbonate [CaCO₃]) in an area with fuel hydrocarbons or chlorinated solvent (particularly less-chlorinated solvents) concentrations elevated above background conditions can be used to infer that fuel hydrocarbons, less-chlorinated solvents, or native organic carbon has been destroyed through aerobic and anaerobic microbial respiration.

Total alkalinity (as CaCO₃) was measured in groundwater samples collected in August 1997 (Table 4.4, Figure 4.20). Alkalinity is a measure of the ability of groundwater to buffer changes in pH. Total alkalinity in USZ groundwater at the site varied from 190 mg/L to 560 mg/L. Alkalinity within the plume area is highly variable, but appears to be sufficient to buffer potential changes in pH caused by biologically mediated reactions. The pH data for site groundwater (Section 4.3.7.2) suggest that groundwater pH decreases slightly at some locations within the plume.



Free carbon dioxide concentrations also were measured in groundwater samples collected in August 1997 (Table 4.4). As shown on Figure 4.21, carbon dioxide concentrations within the full extent of the chlorinated solvent plume are not elevated noticeably above background concentrations. Background concentrations are about 50 to 65 mg/L, while within the plume area, concentrations are in the range of 40 to 110 mg/L. Carbon dioxide concentrations above background could result from both aerobic and anaerobic biodegradation processes as contaminants are ultimately converted to carbon dioxide and water.

4.3.7.2 pH

Groundwater pH was measured for samples collected from monitoring points and monitoring wells in August 1997 (Table 4.4). The pH of a solution is the negative logarithm of the hydrogen ion concentration [H⁺]. USZ groundwater pH measured at the site ranges from 6.80 to 7.36 standard units, and from 7.31 to 7.75 standard units in the LSZ. This range of pH is within the optimal range for BTEX- and CAH-degrading microbes of 6 to 8 standard units.

4.3.7.3 Temperature

Groundwater temperature was measured at monitoring points and monitoring wells in August 1997 (Table 4.4). Temperature can affect the types and growth rates of bacteria that can be supported in the groundwater environment, with higher temperatures typically resulting in higher growth rates. Wiedemeier *et al.* (1996a) report that biochemical processes are accelerated at groundwater temperatures greater than 20°C. Groundwater temperatures at FTA-2 for the USZ varied from 16.9°C to 22.0°C, and from 17.5°C to 21.5°C for the LSZ. These are moderately warm temperatures for groundwater (but still within the optimal range), suggesting that bacterial growth rates should be high.

4.3.8 Screening Table for CAH Degradation

Wiedemeier et al. (1996a) present a worksheet to allow an initial assessment of the prominence of the natural attenuation of CAHs at a site. The worksheet, including the point values determined for FTA-2, is included as Table 4.5. The interpretation of points awarded during the screening process outlined in Table 4.5 is shown in Table 4.6.

The score for FTA-2 computed using Table 4.5 is 9. Based on the evidence presented herein, there is limited evidence for biodegradation of CAHs at FTA-2.

4.4 APPROXIMATION OF BIODEGRADATION RATES

Biodegradation rate constants are necessary to accurately simulate the fate and transport of contaminants dissolved in groundwater. In many cases, biodegradation of contaminants can be approximated using first-order kinetics. First-order biodegradation rate constants may be calculated on the basis of field-scale data, or by using a small sample of the aquifer material and groundwater for microcosm studies. For reductive dehalogenation of chlorinated compounds, first-order rates are not necessarily an

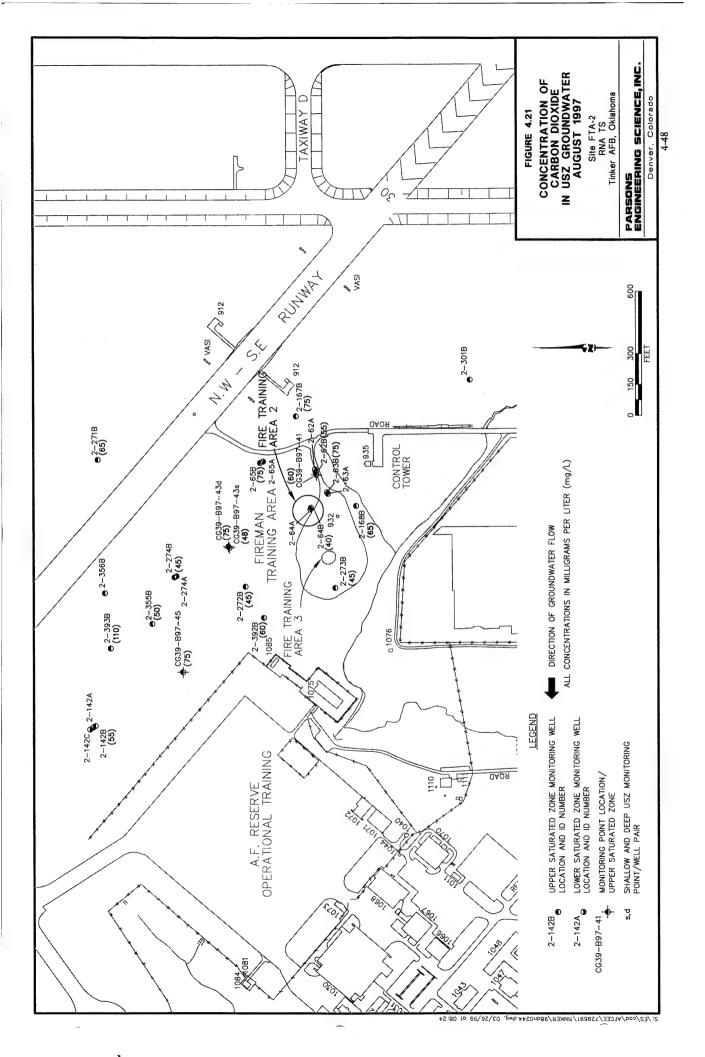


TABLE 4.5 ANALYTICAL PARAMETERS AND WEIGHTING FOR PRELIMINARY SCREENING

FTA-2 RNA TS TINKER AFB, OKLAHOMA

Analysis	Concentration in Most Contaminated Zone	Interpretation	Value	FTA-2 Score
Owner	<0.5 mg/L	Tolerated, suppresses the reductive	3	1
Oxygen	CO.5 mg/L	pathway at higher concentrations	,	1
	>1 mg/L	VC may be oxidized aerobically	-3	
Nitrate	<1 mg/L	At higher concentrations may compete	2	1
Milate	T Mg E	with reductive pathway	-	•
Iron II	>1 mg/L	Reductive pathway possible	3	0
Sulfate	<20 mg/L	At higher concentrations may compete	2	0
		with reductive pathway		
Sulfide	>1 mg/L	Reductive pathway possible	3	0
Methane	<0.5 mg/L	VC oxidizes	0	0
	>0.5 mg/L	Ultimate reductive daughter product, VC Accumulates	3	1
Oxidation	<50 millivolts (mV)	Reductive pathway possible	1	1
Reduction	<-100mV	Reductive pathway likely	2	1
Potential (ORP)				
pН	5 < pH < 9	Optimal range for reductive pathway	0	0
	5 > pH >9	Outside optimal range for reductive pathway	-2	0
TOC	> 20 mg/L	Carbon and energy source; drives dechlorination; can be natural or anthropogenic	2	0
Temperature	> 20°C	At T >20°C biochemical process is accelerated	1	1
Carbon Dioxide	>2x background	Ultimate oxidative daughter product	1	0
Alkalinity	>2x background	Results from interaction of carbon dioxide with aquifer minerals	1	0
Chloride	>2x background	Daughter product of organic chlorine	2	0
Hydrogen	>1 nM/L	Reductive pathway possible, VC may accumulate	3	
	<1 nM/L	VC oxidized	0	
Volatile Fatty Acids	> 0.1 mg/L	Intermediates resulting from biodegradation of aromatic compounds; carbon and energy source	2	0
BTEX	> 0.1 mg/L	Carbon and energy source; drives dechlorination	2	0
PCE		Material released	0	0
TCE		Material released	0	0
		Daughter product of PCE	2 a/	
1,2-DCE		Material released	0	0
		Daughter product of TCE. If cis is greater than 80% of total DCE it is likely a daughter product of TCE	2 ^{d/}	2.
VC	•••	Material released	0	0
		Daughter product of DCE	2 a/	1
Ethene/Ethane	>0.01mg/L	Daughter product of VC/ethene	2	0
	>0.1 mg/L		3	0
Chloroethane		Daughter product of VC under reducing conditions	2	

TABLE 4.5 (Continued) ANALYTICAL PARAMETERS AND WEIGHTING FOR PRELIMINARY SCREENING

FTA-2 RNA TS

TINKER AFB, OKLAHOMA

Analysis	Concentration in Most Contaminated Zone	Interpretation	Value	FTA-2 Score
1,1,1- Trichloroethane		Material released	0	••
1,2- Dichlorobenzene		Material released	0	
1,3- Dichlorobenzene		Material released	0	-
1,4- Dichlorobenzene		Material released	0	
Chlorobenzene		Material released or daughter product of dichlorobenzene	2 3/	
1,1-DCE		Daughter product of TCE or chemical reaction of 1,1,1-TCA	2ª/	
		TOTAL	45	9

Points awarded only if it can be shown that the compound is a daughter product (i.e., not a constituent of the source NAPL).

TABLE 4.6 INTERPRETATION OF POINTS AWARDED DURING NATURAL ATTENUATION SCREENING

FTA-2 RNA TS TINKER AFB, OKLAHOMA

Score	Interpretation
0 to 5	Inadequate evidence for biodegradation of chlorinated organics
6 to 14	Limited evidence for biodegradation of chlorinated organics
15 to 20	Adequate evidence for biodegradation of chlorinated organics
>20	Strong evidence for biodegradation of chlorinated organics

accurate representation because the degradation of CAHs depends on both concentration of electron donors (e.g., fuel hydrocarbons or native organic material) and the concentrations of electron acceptors (including inorganic acceptors and CAHs) (Moutoux et al., 1996). Second-order rates would therefore allow more accurate modeling, but at this time there are no readily available groundwater contaminant transport models that can utilize such rates. Therefore, first-order rates must be estimated for use in a contaminant transport model.

Microcosm studies can be used to show that the microorganisms necessary for biodegradation are present and to help quantify rates of biodegradation. Because microcosm studies are time consuming and expensive, they should be undertaken only at sites where there is considerable uncertainty concerning the biodegradation of contaminants. Biodegradation rate constants determined by microcosm studies often are much higher than those observed in the field, and it may be inappropriate to use them to generate rate constants. The preferable method of contaminant biodegradation rate-constant determination is by use of field data, so no microcosm studies were conducted at the FTA-2 site.

In order to calculate first-order biodegradation rate constants on the field scale, the apparent degradation rate must be normalized for the effects of dilution, dispersion, and sorption. Two methods for determining first-order rate constants for BTEX compounds are described by Wiedemeier *et al.* (1996a). One method involves the use of a biologically recalcitrant compound found in the dissolved contaminant plume that can be used as a conservative tracer. One organic compound that can be used at some sites is TMB. However, TMB was not detected at FTA-2. For many chlorinated solvent plumes, chloride that is released to the groundwater during dehalogenation reactions also can serve as a tracer. Chloride was not used as a tracer for FTA-2 due to the similarity between background chloride concentrations and those in the CAH plume. The other method, proposed by Buscheck and Alcantar (1995), involves interpretation of a steady-

state contaminant plume and is based on the one-dimensional, steady-state analytical solution to the advection-dispersion equation presented by Bear (1979). Decay rates computed using this method account for chemical (abiotic) decay and biological (aerobic and anaerobic) decay. For an expanding plume, this first-order approximation can be viewed as an upper bound on the biodegradation rate. Use of this method overestimates biodegradation rate, because a typical expanding plume exhibits decreasing source area concentrations, increasing downgradient concentrations, or both. Over time, these changes result in a decreasing slope on a log-linear plot, and consequently a decreasing biodegradation rate. Both methods can be adapted to estimate rate constants for CAHs.

Another method for estimating reductive dehalogenation rates of CAHs is described by Moutoux et al. (1996). This method provides a total reductive dehalogenation rate for all dehalogenation steps. All rates (including the rapid TCE to DCE rate and the slow VC to ethene) are averaged together using the Moutoux et al. (1996) method. Because abiotic reactions and aerobic reactions that involve CAH compounds in the role of an electron donor are not included in this rate (and may in fact decrease the rate estimated by this method), the rate should be considered a lower bound on the destructive attenuation rate.

It should be noted that the magnitudes of the decay rates calculated using these methods are sensitive to the estimated migration velocity of the contaminant along the selected flowpath; therefore, inaccuracies in the estimated velocity will cause the decay rates to be less accurate. Overestimation of the groundwater velocity will result in overestimation of the decay rate, and vice versa.

The method of Buscheck and Alcantar (1995) was used to estimate first-order biodegradation rate constants for TCE, cis-1,2-DCE, trans-1,2-DCE and total chlorinated ethenes at FTA-2. The method of Moutoux et al. (1996) was also used to estimate a first-order biodegradation rate constant for total chlorinated ethenes at FTA-2. The decay rate calculations are summarized in Table 4.7 and Appendix E. Additional decay rate calculations for chlorinated ethanes and chlorobenzenes are also included and summarized in Appendix E. Rate constants were not calculated for BTEX compounds because of the relatively low remaining concentrations (Table 4.2).

The first-order, steady-state, TCE decay rate computed using data from the flowpath 2-62B, CG39B97-43S, and 2-355B and the method of Buscheck and Alcantar (1995) is 3.06E-04 day⁻¹, which corresponds to a half-life of 6.20 years. This flowpath extends from the inferred primary source area at well 2-62B to near the leading edge of the dissolved TCE plume. These wells are all screened within the upper USZ sand interval, and therefore may not be representative of decay rates in groundwater in the lower USZ sand interval. However, for purposes of modeling contaminant transport at the site, the wells used above are assumed to represent biodegradation for all of the USZ. For *cis*-1,2-DCE and *trans*-1,2-DCE, the calculated first order biodegradation rates are 4.08E-4 day⁻¹ (half life of 4.65 years) and 7.58E-04 day⁻¹ (half life of 2.51 years), respectively.

Total chlorinated ethene degradation was calculated by both the methods of Buscheck and Alcantar (1995) and Moutoux, et al., (1996). The first-order, steady-

TABLE 4.7 SUMMARY OF CHLORINATED SOLVENT DECAY RATES

AUGUST 1997 FTA-2 RNA TS

TINKER AIR FORCE BASE, OKLAHOMA

Contaminant	Sample	Flow Path	Analysis	Decay Rate	Half Life
	Date		Method	(day ⁻¹)	(years)
TCE	August-97	2-62B > CG39B97-43s > 2-355B	Buscheck and Alcantar ^{a/}	3.06E-04	6.20
cis-1,2-DCE	August-97	2-62B > CG39B97-43s > 2-355B	Buscheck and Alcantar	4.08E-04	4.65
trans -1,2-DCE	August-97	2-62B > CG39B97-43s	Buscheck and Alcantar	7.58E-04	2.51
Total Chlorinated Ethenes	August-97	2-62B > CG39B97-43s > 2-355B	Buscheck and Alcantar	2.96E-04	6.42
Total Chlorinated Ethenes	August-97	2-62B > CG39B97-43s > 2-355B	Reductive Dechlorination	1.86E-07	10,210
	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1				

^{2/} Method of Buscheck and Alcantar (1995) for steady-state plumes.

^{b/} Reductive Dechlorination by method of Moutoux, et al., (1996), which gives decay rate attributable to reductive dechlorination.

state, total chlorinated ethene decay rate computed using data from the flowpath 2-62B, CG39B97-43s, and 2-355B and the method of Buscheck and Alcantar (1995) is 2.96E-04 day⁻¹, which corresponds to a half-life of 6.42 years. This rate is very similar to that calculated for TCE alone. As described above, the decay rates computed using this method can be viewed as an upper bound on the biodegradation rate if the plume is expanding. Data presented in Section 4.2 suggests that the plume is still expanding to a significant degree. Therefore, the computed decay rate should be considered an upper bound of the combined effects of both reductive dehalogenation occurring in the general source area and aerobic decay occurring downgradient.

An attempt was made to compute a lower bound dehalogenation rate along this same flowpath using the method of Moutoux et al. (1996). The corrected CAH concentration increased with distance along the flowpath between wells 2-62B and CG39B97-43s, although there was an overall decrease in corrected CAH concentration between 2-62B and 2-355B. The first-order, total reductive dehalogenation decay rate computed for total chlorinated ethenes using data from the flowpath 2-62B, CG39B97-43s, and 2-355B and the method of Moutoux (1996) is 1.86E-07 day⁻¹, which corresponds to a half-life of 10,210 years. The total reductive dehalogenation biodegradation rate calculated by the Moutoux method indicates that reductive dehalogenation along this flow path is not applicable. This is probably due to the occurrence of aerobic biodegradation of TCE and/or DCE, because the method assumes that both parent and daughter CAHs are conserved along the selected flowpath. The failure of this method indicates that other biodegradation processes besides reductive dehalogenation (e.g., aerobic processes) are occurring along this flowpath.

4.5 SUMMARY

Compounds detected in groundwater at FTA-2 include benzene, toluene, PCE, TCE, cis-1,2-DCE, trans-1,2-DCE, 1,1-DCE, VC, 1,1-DCA, 1,2-DCA, 1,2-DCB, 1,3-DCB, 1,4-DCB, chloroform, and CB. Of these compounds, benzene, PCE, TCE, 1,1-DCE, cis-1,2-DCE, VC, 1,2-DCA, CB, 1,2-DCB and 1,4-DCB were detected at concentrations exceeding USEPA MCLs. Benzene exceeded the MCL of 5 μ g/L at one location, monitoring well 2-62B at a concentration of 6.0 μ g/L. Because concentrations of benzene rapidly attenuate to levels below detection outside of the primary source area (well 2-62B), benzene is not considered for further remedial action.

PCE was detected at a maximum concentration of 52 μ g/L in well 2-65B, exceeding the MCL for PCE of 5 μ g/L. PCE was also detected at a concentration of 6.1 μ g/L in well 2-62B, and 1.9 μ g/L in well 2-63B. The occurrence of PCE is limited to the general source area and attenuates to below detection limits with groundwater migration.

TCE was detected at concentrations ranging from non-detect to a maximum of 9,440 μ g/L in well 2-62B, exceeding the MCL for TCE of 5 μ g/L. The extent of the TCE plume extends northwest in the USZ to well 2-393B at a concentration of 406 μ g/L. The distribution of TCE in USZ groundwater (Figure 4.3) suggests a vertical migration of TCE from the upper USZ sand to the lower USZ sand. The vertical migration of contaminants is the result of a discontinuous clay layer that pinches out northwest of the source area. Concentrations of TCE at source well 2-62B do not show any discernible decreasing trend (Figure 4.10). The increase of TCE concentration with time at two wells

along the contaminant flowpath (2-272B and 2-274B) suggests that the plume is expanding. TCE was also detected for the first time in the LSZ at a concentration of 10.9 µg/L in well 2-62A.

Cis-1,2-DCE was detected at a maximum concentration of 1,200 µg/L in 2-62B, exceeding the MCL for cis-1,2-DCE of 70 µg/L. The highest concentrations of cis-1,2-DCE coincide with the highest TCE detections at 2-62B and CG3-B97-43s in the upper USZ sand interval. The distribution of cis-1,2-DCE in USZ groundwater (Figure 4.4) suggests a vertical migration from the upper USZ sand interval to the lower USZ sand interval. The concentration of cis-1,2-DCE versus time at monitoring well 2-62B (Figure 4.10) does not show any discernible temporal trend. However, the relatively high ratio of cis-1,2-DCE to trans-1,2-DCE in the source area and the predominance of cis-1,2-DCE over trans-1,2-DCE throughout the plume indicate that TCE is being reductively dehalogenated in a limited source area.

1,2-DCA was detected at concentrations of 260 μ g/L and 49.2 μ g/L in 2-62B and 2-393B, respectively, both exceeding the MCL of 5 μ g/L. The two detections of 1,2-DCE were at opposite ends of the overall CAH contaminant plume, suggesting two source release areas (Figure 4.7). The distribution of total chlorinated ethanes versus time (Figure 4.11) suggests that at monitoring well 2-62B chlorinated ethanes have slightly attenuated from 1993 to 1997.

CB was detected at a concentration of 117 μ g/L in 2-62B, exceeding the MCL of 100 μ g/L. In the sample from 2-65B, the CB concentration was only 9.5 μ g/L. At every other sampling location CB was detected at 1 μ g/L (CG39-B97-43S) or not-detected. Erratic concentrations of total chlorobenzenes in source area well 2-62B (Figure 4.12) suggest that the chlorobenzene source has not significantly attenuated with time. The lack of significant quantities of CB along the contaminant plume suggests that CB is being utilized aerobically as an electron donor/substrate in the same manner as BTEX compounds.

1,2-DCB was detected at a concentration of 950 μ g/L in 2-62B, exceeding the MCL of 600 μ g/L. 1,4-DCB was detected at a concentration of 168 μ g/L in 2-62B, exceeding the MCL of 75 μ g/L. The trends in 1,2-DCB and 1,4-DCB are similar to those in CB at FTA-2.

Data have been collected from several wells at the site since 1994, and changes in compound concentrations since 1994 are shown on Figures 4.9 through 4.12. Some concentrations have decreased slightly, but contaminant concentrations in single monitoring wells generally have fluctuated over time, with no clear overall trends of increasing or decreasing concentrations (i.e., some increase, some decrease, and some are relatively constant). This is likely a function of the low concentrations, mixing in wells during sampling, slight differences in sampling procedures, variations in plume concentrations due to water level/flow direction changes, and variable rates of desorption from aquifer solids.

The additional data collected for this effort also help to suggest alternative locations for sources of groundwater contamination in the FTA-2 area. These locations include the access road area immediately upgradient of the FTA-2 fire pit in the vicinity of well 2-

62B (TCE, 1,2-DCA, and CBs) and well 2-65B (PCE), and an area north of the FTA-2 pit at 2-393B (1,2-DCA). The nature of these sources of chlorinated solvents is not clear and no mobile NAPL has been detected. The lack of information on the sources presents a problem because contaminant mass loading cannot be accurately evaluated and the possibility of multiple releases cannot be discounted. However, the high concentrations detected at monitoring well 2-62B indicate a singular source area that dominates the occurrence of the dissolved contaminant plume.

Several lines of chemical and geochemical evidence indicate that the contaminants at FTA-2 are being biodegraded, either as substrates or as electron acceptors. Geochemical data indicate that DO and nitrate are being consumed and methane is produced near and hydraulically downgradient of the suspected source area (monitoring wells 2-273B and 2-168B). This evidence suggests that microbial consumption of anthropogenic and natural organic matter (carbon compounds) is taking place at FTA-2. Anthropogenic carbon compounds are present as petroleum hydrocarbons and less-chlorinated solvents (including, but not limited to, BTEX, chlorobenzene, VC, DCE, and DCA) that are dissolved in groundwater. Native carbon compounds include solid organic matter in the aquifer matrix and dissolved organic carbon in site groundwater. These compounds are electron donors in redox reactions that are also consuming electron acceptors (e.g., DO, iron (III), carbon dioxide, and chlorinated hydrocarbons) and creating byproducts [e.g., methane, iron (II), and carbon dioxide].

Concentration plots of contaminants, electron donors, electron acceptors, and biodegradation byproducts (Figures 4.3 through 4.21) provide some qualitative evidence of biodegradation processes. DO is depleted in the vicinity of the BTEX plume area, near monitoring wells 2-62B and hydraulically downgradient of the suspected source area. Nitrate is slightly reduced in this area, indicating microbial activity, however nitrate does not show a decreasing trend along the extent of the CAH contaminant plume. The highest iron (II) concentrations do not coincide with the solvent plume (2-271B, 2-142B, and 2-63A), suggesting that iron (III) hydroxide is not being reduced to iron (II) during biodegradation of native organic carbon, fuel hydrocarbons, or less-chlorinated solvents. Sulfate concentrations vary across the site, the areas of decreased sulfate concentrations do not correlate well with chlorinated solvent concentrations. It is therefore unlikely that sulfate reduction is an ongoing anaerobic biodegradation process at FTA-2. The area of elevated methane concentrations coincides with the fuel hydrocarbon and chlorinated solvent source area. The presence of methane in the source area indicates that conditions are sufficiently reducing the petroleum hydrocarbons and native organic matter and may be used to support methanogenesis. Further, the fact that methanogenesis is ongoing indicates that conditions in the plume area are highly reducing and therefore favorable for reductive dehalogenation of chlorinated solvents. Plots of additional indicators, such as carbon dioxide and ammonia concentrations, further indicate that biodegradation reactions are not rapidly occurring at FTA-2.

Elevated chloride concentrations within the solvent plumes also may indicate dehalogenation reactions, although chloride may also come from the oxidation of less-chlorinated compounds (i.e., utilization as a substrate). Because of other reductive dehalogenation evidence, it is likely that some fraction of the excess chloride is a result of chlorinated solvent degradation. In addition, the fact that *cis*-1,2-DCE is detected instead

of *trans*-1,2-DCE in wells where TCE or VC also are present is indicative of biologically mediated reductive dehalogenation.

Given the available evidence, the dissolved CAH plume at FTA-2 exhibits characteristics of both Type 1 and Type 3 behavior. Dissolved petroleum hydrocarbons and possibly some less-chlorinated solvents appear to act as carbon sources. It is likely that the introduction of petroleum hydrocarbons and solvents due to waste disposal activities stimulated microbial activity and created groundwater conditions reducing enough to allow reductive dehalogenation of TCE, DCE, 1,2-DCA and CBs. The TCE that is not transformed to DCE and that migrates away from the source area into more aerobic environments will not be readily degraded, and may be relatively persistent. VC has been created due to dehalogenation, but it is possible that it is being oxidized (as a substrate) once it enters aerobic groundwater. Likewise, it is possible that DCE and DCA are oxidized as well. However, conclusive evidence of these oxidation processes has not been gathered. It is possible that some microbial consumption of native organic matter is taking place outside the plume and within the plume, but the DO values measured outside of the immediate source area suggest this process is minor. Type 2 behavior is not discernible from consumption of contaminants.

As discussed in Section 4.4, rates of degradation of TCE, cis-1,2-DCE, trans-1,2-DCE, and total chlorinated ethenes, estimated from data collected for this investigation, calculated by the Buscheck and Alcantar (1995) method were 3.06E-04 day-1 (6.20 years), 4.08E-04 days-1 (4.65 years), 7.58E-04 days-1 (2.51 years), and 2.96E-04 day-1 (6.42 years), respectively. As petroleum hydrocarbons and native organic matter continue to be consumed, reductive dehalogenation of the chlorinated ethenes will continue; however, once BTEX compounds are completely degraded, the rate at which dehalogenation proceeds may slow as the microbes utilize less easily degradable fuel hydrocarbons or other organic matter. It also is conceivable that the plume may eventually come to display only Type 3 behavior, but this will depend on the concentrations and degradability of the remaining (non-BTEX) fuel hydrocarbon compounds relative to the native organic matter.

As the type of organic matter used during biodegradation changes, the redox conditions also may change and alter which reactions are favored. This could also affect the rates at which CAHs are reductively dehalogenated, and could cause dehalogenation of some compounds to cease. If and how this will happen is difficult to predict; the relative "degradability" of the native organic matter (or non-BTEX fuel hydrocarbons) and the effect of a changing electron donor source is not yet well-understood. If conditions become more oxidizing (but not aerobic), the dehalogenation of DCE may also cease or slow. However, aerobic and/or anaerobic mineralization of VC and DCE may become more important (i.e., Type 3 conditions may dominate).

SECTION 5

GROUNDWATER FLOW AND CONTAMINANT TRANSPORT MODEL

Computer simulations of USZ groundwater flow and the transport of dissolved TCE are used to evaluate the future migration and natural attenuation of TCE dissolved in groundwater at Tinker AFB FTA-2. The model was developed using site-specific data and conservative assumptions about governing physical and chemical processes.

5.1 GENERAL OVERVIEW AND MODEL DESCRIPTION

The computer programs "A Modular Three-Dimensional Finite-Difference Ground-Water Flow Model" (MODFLOW) (McDonald and Harbaugh, 1988) and MT3D (Zheng, 1990) were used to evaluate dissolved TCE at the site. MODFLOW was used to compute hydraulic heads and groundwater fluxes. MT3D96 was used to compute solute transport of dissolved TCE due to advection, dispersion, adsorption, and biodegradation. The preand post-processors contained in Visual MODFLOW, Version 2.50 (Waterloo Hydrogeologic Software, 1997) were used to build a site-specific model for Tinker AFB FTA-2. The graphic user interface facilitated model development and analysis as well as presentation of model results.

MODFLOW is a three-dimensional groundwater flow simulation computer program published by the U.S. Geological Survey (McDonald and Harbaugh, 1988) and incorporated into Visual MODFLOW. MODFLOW uses implicit solution techniques to solve the transient groundwater flow equation for hydraulic head (McDonald and Harbaugh, 1988). The solution techniques are based on a one-, two-, or three-dimensional, block-centered, finite difference grid, which is superimposed on the model area. The aquifer properties can be heterogeneous and anisotropic, and aquifer layers can be simulated as confined, unconfined, or a combination of both (McDonald and Harbaugh, 1988). Results of a MODFLOW simulation include the distribution of hydraulic head within each model layer as well as groundwater fluxes through the model area.

The MT3D code incorporates advection, dispersion, sorption, and biodegradation to simulate contaminant plume migration and degradation. The MODFLOW-computed heads and fluxes are used by MT3D to compute groundwater flow velocity and thus solute transport. MT3D solution routines are based on the Method of Characteristics (MOC) solute transport model [e.g., as developed by Konikow and Bredehoeft (1978) for the US Geological Survey (USGS) two-dimensional (2-D) MOC model code]. The MOC model was modified by Zheng (1990) to allow three-dimensional (3-D) solutions, and to

allow use of a modified MOC that reduces numerical dispersion. The modified model is called MT3D and is incorporated into Visual MODFLOW.

5.2 MODEL OBJECTIVES

The modeling effort had three primary objectives: 1) predict the future extent and concentrations of the dissolved TCE contaminant plume by modeling the combined effects of advection, dispersion, sorption, and biodegradation; 2) assess the potential for downgradient receptors to be exposed to contaminants (TCE) at concentrations above regulatory levels of concern; and 3) if applicable, provide technical support for the natural attenuation remedial option at post-modeling regulatory negotiations. Specifically, the MT3D simulations for FTA-2 were constructed to address the following questions:

- 1. Does the observed distribution of TCE accurately reflect measured groundwater flow conditions or assumed source configuration?
- 2. What groundwater flow conditions (hydraulic conductivity, hydraulic anisotropy) would account for the observed distribution of TCE?
- 3. What source configuration can account for the observed distribution TCE?
- 4. To what extent is natural attenuation controlling the TCE plume at this site?
- 5. How will various remedial alternatives affect TCE plume migration and groundwater concentrations?

5.3 CONCEPTUAL MODEL DESIGN AND ASSUMPTIONS

As discussed in Sections 3 and 4, groundwater flow in USZ groundwater at FTA-2 is toward the west-southwest; however, TCE transport is apparently toward the northwest from the FTA-2 source area (location of well 2-62B). This discrepancy between the direction of maximum hydraulic gradient and the primary direction of contaminant transport is not well understood, but could be caused by either anisotropic hydraulic conductivity in the USZ or multiple source areas to the northwest of the primary FTA-2 source area. Because there are no data to support a hypothesis of multiple source areas, the model presented herein assumes that the difference between groundwater flow and apparent contaminant transport directions is due to preferential flowpaths and anisotropic hydraulic conductivity in the USZ.

The TCE plume apparently migrates northwest from the FTA-2 source area through the upper USZ sand interval where it is underlain by and intra-USZ clay layer. This intra-USZ clay layer pinches out to the northwest providing direct hydraulic connection between the upper USZ sand interval and the lower USZ sand interval in the vicinity of wells 2-274B and 2-355B. The absence of the intra-USZ clay layer northwest of the FTA-2 source area also allows dissolved TCE in the upper USZ sand interval to migrate downward into the lower USZ sand interval in this area. Chlorinated solvents are not detected in the LSZ groundwater system downgradient of the FTA-2 source area. Based on the vertical distribution of TCE, the solute transport model for FTA-2 is composed of three layers. Layer 1 represents the upper USZ sand interval, layer 2 represents the intra-

USZ clay separating the upper and lower USZ sand intervals, and layer 3 represents the lower USZ sand interval. The base of the model domain represents the regional USZ/LSZ aguitard. Model layer properties are presented in detail in Section 5.4.2.

Important assumptions made when using the MT3D code are that dispersion, sorption, and biodegradation are significant factors controlling contaminant fate and transport at the site. Dispersivity, which is a characteristic of the porous medium, is a measure of the longitudinal and lateral spreading of the contaminant plume caused by local heterogeneities that cause deviations from the average linear solute migration velocity. The magnitude of dispersion is generally believed to be scale-dependent: the longer the plume flowpath, the greater the dispersion. Given the considerable length of the plume flowpath (approximately 1,600 feet) and the documented presence of subsurface heterogeneities (Section 3), it is reasonable to assume that dispersion is an important parameter influencing solute transport in the study area. According to data presented in Sections 3 and 4, concentrations of organic carbon within the site soils may support significant sorption. Data also suggest that both anaerobic and aerobic biodegradation of TCE is occurring within the contaminant plume. Dispersion is estimated using literature values and accepted rules-of-thumb; sorption (assumed to be a linear process) is simulated using a coefficient of retardation; and biodegradation is simulated using a firstorder decay constant. Selection of values for these model-input parameters is discussed in Section 5.4.3.

5.4 INITIAL MODEL SETUP

The initial setup for this model was based on site-specific data where possible. Where site-specific data were not available (e.g., for effective porosity), reasonable assumptions were made on the basis of widely accepted literature values for materials similar to those found in the USZ at FTA-2. The following sections describe the basic model setup. Those model parameters varied during model calibration are discussed in Section 5.5.

5.4.1 Grid Design

The model area for FTA-2 includes the FTA-3 source area and the downgradient area encompassing the TCE plume. The model domain for FTA-2 is represented using three layers, with a 63-cell by 69-cell horizontal grid (Figure 5.1). The long axis of the model grid is oriented east to west, parallel to the primary direction of maximum hydraulic gradient for the USZ. The model cells were varied in size, with dimensions of 50 feet (west-to-east) by 50 feet (north-to-south) in the general source area. Grid cell dimensions were expanded to a maximum of 175 feet (west-to-east) by 200 feet (north-to-south) outside of the general source area and measured TCE plume.

5.4.2 Model Layers

The TCE plume is migrating primarily through the upper and lower intervals of the USZ hydrogeologic unit. The model area was therefore designed with three layers. The upper layer represents the upper USZ sand interval, a middle layer represents the

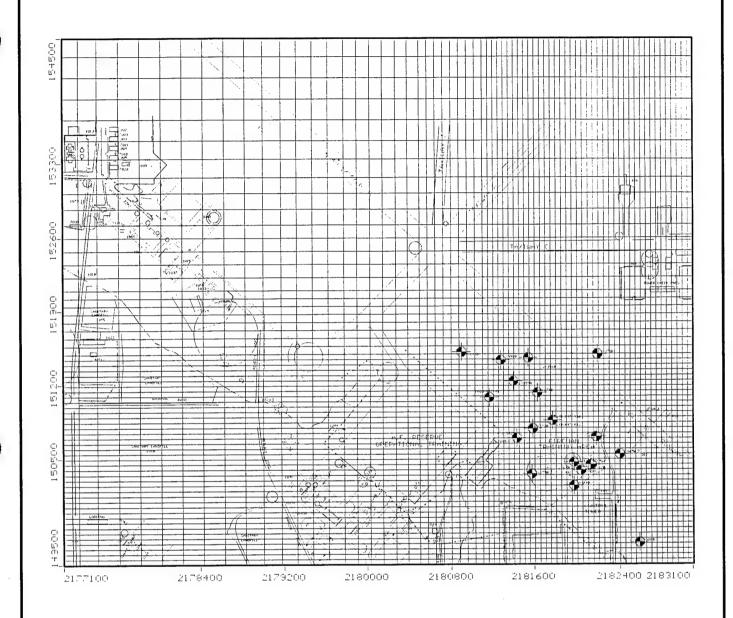




FIGURE 5.1

MODEL GRID SUPERIMPOSED ON FTA-2 SITE LOCATION

Site FTA-2 RNA TS Tinker AFB, Oklahoma

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

intra-USZ clay horizon, and the lower layer represents the lower USZ sand interval. Where the intra-USZ clay horizon is not present, the upper and lower USZ sand intervals are in direct hydraulic connection, and the middle layer is assigned properties similar to those for the upper USZ sand interval.

The base of each layer was defined using borehole data and geophysical logs. The base of layer 1 represents the top of the intra-USZ clay layer. As a result, channels. pinch-outs and other features of that surface are represented in the model. Layer 2 was defined using borehole data, and was specified as clay with a minimal thickness of one foot in areas to the north and west beyond the pinch-out of the intra-USZ clay body. This maintained the spatial relationship between layers 1 and 3 while presenting negligible resistance to vertical contaminant transport between the layers. Borehole data indicate the presence of intra-USZ clay bodies other than the near-source body occupying layer 2. These do not appear to exert an effect on contaminant transport because of location and/or limited lateral extent. These bodies were excluded from the material distribution, resulting in considerable model simplification. The base of layer 3, or the lower boundary of the model, was defined using borehole data from geophysical logs to contour the top of the USZ/LSZ Aguitard. As a result, channels and other features of that surface are represented in the model. All three layers were assumed to be hydraulically connected, with the presence of vertical hydraulic gradients largely ignored. Regional stratigraphic and hydraulic gradients are similar, at approximately 40 ft/mile (0.0076 ft/ft). Therefore the layers were sloped to the west at this gradient, although local variations in stratigraphic gradient may exist.

5.4.3 Groundwater Flow Model

This section presents the initial input parameters used for the groundwater flow simulations. Contaminant transport input parameters are discussed in Section 5.4.4.

5.4.3.1 Boundary Conditions

Boundary conditions describe the interaction between the system being modeled and its surroundings or, for transport models, the loading of contaminant mass into the system. Boundary conditions are used to include the effects of the system outside the area being modeled with the system being modeled, while at the same time allowing the isolation of the desired model domain from the larger system. In effect, the boundaries of the model tell the area immediately inside the boundaries what to expect from the outside world. The solution of any differential equation requires specification of the conditions at the periphery of the system. Model boundaries are thus mathematical statements that specify the dependent variable (head or contaminant concentration) or the flux at the model grid boundaries.

Three types of boundary conditions are generally used to describe groundwater flow and solute transport. Boundary conditions are referred to as specified-head type (Dirichlet), specified-flux type (Neumann), and head-dependent or mixed type (Cauchy). Table 5.1 summarizes boundary conditions for groundwater flow and solute transport.

TABLE 5.1 COMMON DESIGNATIONS FOR SEVERAL IMPORTANT BOUNDARY CONDITIONS FTA-2 RNA TS

TINKER AIR FORCE BASE, OKLAHOMA

		General Mathema	tical Description
Boundary Condition	Formal Name	Groundwater Flow	Contaminant Transport
Specified-Head or Specified- Concentration	Dirichlet	H = f(x, y, z, t)	C = f(x, y, z, t)
Specified-Flux	Neumann	$\frac{\partial H}{\partial n} = f(x, y, z, t)$	$\frac{\partial C}{\partial n} = f(x, y, z, t)$
Head-Dependent or Concentration-Dependent Flux	Cauchy	$\frac{\partial H}{\partial n} + cH = f(x, y, z, t)$	$\frac{\partial C}{\partial n} + cC = f(x, y, z, t)$
(mixed-boundary condition)			

(Modified from Franke et al., 1987)

In flow models, boundary conditions are ideally used to specify actual hydrogeologic boundaries to the system, such as a geologic feature that may bound a system or areas where properties (e.g., flux) are known and can be defined. When using a numerical flow model, hydrologic boundaries such as constant-head features (e.g., lakes, streams, etc.) or constant-flux features (e.g., groundwater divides, confining units, etc.,) should, when possible, coincide with the perimeter of the model. In areas that lack obvious hydrologic boundaries, constant-head or constant-flux boundaries can be specified at the numerical model perimeter as long as the perimeter is far enough removed from the contaminant plume that transport calculations would not be affected by inaccuracies in the simulated boundary conditions.

Specified-head cells were defined at the eastern and western ends of model layer 1 (Appendix E). The specified heads were estimated by projecting heads from the August 1997 groundwater flow map. Although some vertical contaminant transport is indicated by the detection of TCE in well 2-62A (screened in the LSZ), a no-flow boundary was specified at the base of layer 3. The upper model boundary is defined by the simulated water table surface, although an arbitrary top of the model domain was set at 1,250 NGVD to accommodate potential water table fluctuations during the iterative head solution procedure.

5.4.3.2 Recharge and Evapotranspiration

No recharge to the water table was included in groundwater flow simulations because annual evaporation exceeds precipitation, and because a clay layer of the Hennessey Group occurs at ground surface which may inhibit direct recharge from precipitation. Evapotranspiration was also set to zero due to the depth of the USZ water table surface (10 to 20 feet) and the fine-grained nature of the surface formation.

5.4.3.3 Aquifer Properties

Hydraulic conductivity values were obtained from the results of slug and pumping tests performed during the TS. The calculated values are summarized in Table 3.2, and discussed in Section 3.3.2.1. Average hydraulic conductivities similar to those discussed in Section 3.3.2.1 were initially defined for each portion of the model domain (e.g., for each layer unit).

As mentioned previously, the TCE plume axis does not align with the downgradient groundwater flow direction determined from field-measured water elevations. If, as assumed, the source of TCE contamination is in the vicinity of well 2-62B, there must exist preferred pathways and/or significant anisotropy in horizontal hydraulic conductivity. The flow model developed for this exercise contains both a preferred pathway (a channel of sand more permeable than surrounding materials) and horizontal hydraulic conductivity anisotropy.

The effective porosity is the percentage of a rock or sediment through which fluids can travel. An effective porosity value of 20 percent was defined for the USZ sand units within the model domain on the basis of grain-size observed in soil boreholes for the USZ aquifer sands (Spitz and Moreno, 1996). A lower effective porosity of 10 percent was assigned to the intra-USZ clay layer based on the fine-grained nature of the deposits observed in soil boreholes.

5.4.3.4 Aquifer Stresses

No groundwater supply or recovery wells are currently located in the model domain. Therefore no aquifer stresses were modeled and groundwater flow was assumed to be steady state for the 35-year model calibration simulations. Pumping wells are added to the model domain to evaluate a pump and treat remedial alternative for the site as discussed in Sections 6.3 and 6.4.

5.4.3.5 Aquifer Storage

Although steady-state groundwater flow was assumed, MT3D treats the input flow field calculated by MODFLOW as if it were transient. Aquifer storage properties must therefore be assigned for the solute transport model. In general, the model area is largely unconfined, and a specific yield of 0.20 was assigned. Specific yield is similar to effective porosity of the aquifer, because it represents the water that can drain from the aquifer material pore spaces.

5.4.4 Contaminant Transport Model

Subsequent to performing groundwater flow simulations, the model was used to simulate fate and transport of TCE. Biodegradation rates calculated in Section 4 represent both reductive dechlorination (source area) and aerobic oxidation. The TCE concentrations obtained from August 1997 laboratory analytical results for 19 monitoring wells/points were used as targets for model calibration. Table 4.3 presents dissolved TCE concentration data for August 1997, and Figure 4.3 shows the spatial distribution of dissolved TCE in August 1997.

5.4.4.1 Source

Transport models use boundary conditions to express the influence of contaminant sources such as mobile and residual NAPL, biodegradation of parent compounds, and dissolved mass entering through recharge, injection wells, surface water bodies, or leaking structures. Sources such as NAPL bodies may be represented as specified-concentration boundaries (limited by solubility constraints or observed maximum concentrations) or as specified-flux boundaries (for which the chemical loading rate must be known or estimated). However, in most cases, only the effects of the source are measured; the detailed source characteristics and history are typically unknown (Spitz and Moreno, 1996). The source must therefore be represented as a "black box" that produces appropriate contaminant concentrations or fluxes at selected points in the model. The source may be misrepresented under such a scenario, but there is often little choice in the matter. Estimating contaminant flux into groundwater from NAPL is difficult and is dependent upon several parameters, most of which cannot be measured (Abriola, 1996; Feenstra and Guiguer, 1996).

Rather than using various calculations to attempt to estimate TCE partitioning from residual NAPL into groundwater, the "black-box" source approach was used for this application. Experience modeling contaminated sites as part of the AFCEE Natural Attenuation Initiative has suggested that this is currently the best available method for reproducing observed plumes. At FTA-2, the TCE is assumed to enter groundwater in the study area through contact between groundwater and residual NAPL at or below the water table. Partitioning of TCE from this source into groundwater was simulated by specifying source area dissolved concentrations in four model cells centered around well 2-62B (Appendix E).

It is likely that contaminants were first introduced into groundwater at this site when the FTA-2 fire pit was initially used for fire training exercises in 1962 (TT, 1996). TT (1996) reported the FTA-2 pit was operational until 1966, although aerial photos indicate the pit was present as late as 1970. Fire training exercises at Air Force bases typically use contaminated or uncontaminated jet fuel. Solvents, fuels, and other liquid wastes may have been used for training exercises or cleaning, or dumped in the general fire training area. Aerial photographs show numerous dirt access roads crossing the area through the mid-1970s, indicating unknown activities continued in the area after use of the FTA-2 pit was discontinued.

For the model it was assumed that contaminants first entered the groundwater upon fire training exercises in 1962, and that residual NAPL contamination associated with the

leached waste has acted as a continuous source. It was also assumed that after fire training operations in the source area ceased, the contaminant flux into the aquifer remained constant until August 1997. This is based on the observation that the concentration of TCE detected at well 2-62B has not decreased from 1993 to 1997, and that TCE concentrations do not appear to be significantly attenuating in the source area.

For predictive modeling after August 1997, it was assumed that the source will begin to weather (lose mass) due to processes such as volatilization, dissolution, and biodegradation, and that the mass of contaminants entering groundwater will begin to decline. Past experience with the AFCEE Natural Attenuation Initiative has suggested that residual NAPL sources can weather at rates as rapid as 10 percent per year. However, to be conservative, and because the contaminant sources are poorly characterized, it was assumed that the source at FTA-2 will weather at about two percent per year. Thus, the model assumes that after 1997, the source loading rates decreased geometrically by two percent per year (each year's source strength was decreased by two percent from the previous year's strength).

5.4.4.2 Dispersivity

Published data summarized by Spitz and Moreno (1996) suggest that, as a rule of thumb, longitudinal dispersivity is approximately one-tenth the travel distance of the plume (from the source to the downgradient toe), which is equivalent to approximately 160 feet at FTA-2 (total plume length of 1,600 feet) (Figure 4.3). This was the initial value assigned to longitudinal dispersivity. The initial transverse dispersivity was estimated as one-tenth of the longitudinal dispersivity value, and vertical dispersivity was assumed to be one-hundredth of the longitudinal dispersivity (Domenico and Schwartz, 1990).

5.4.4.3 Retardation

Retardation of contaminants relative to the advective velocity of the groundwater occurs when contaminant molecules are sorbed to organic carbon, silt, or clay in the aquifer matrix. Using measured TOC concentrations near the water table at six locations across the site (see Table 4.1), an assumed bulk density of 1.65 kilograms per liter (kg/L), and published values of the TCE soil sorption coefficient (K_{oc}) (as listed in Wiedemeier et al., 1996a), the coefficient of retardation for each contaminant was calculated. The results of these calculations are summarized in Table 5.2. The lower the assumed coefficient of retardation, the faster the TCE plume will migrate downgradient. Initially, the average calculated retardation coefficient of 5.7 was assigned to the model.

5.4.4.4 Biodegradation

As discussed in Section 4.5, a first-order decay rate of approximately 3×10^{-4} day was calculated for TCE using site-specific data. This rate was used as an upper bound to define an initial value for model input. An initial estimate of 2×10^{-4} day was defined for the model domain.

TABLE 5.2
CALCULATION OF RETARDATION COEFFICIENTS
FTA-2 RNA TS
TINKER AFB, OKLAHOMA

	Retardation	Maximum Minimum Average	11.90 15.02	5.54 6.84	4.37 5.33	3.55 4.29	2.88 3.41	1.13 1.16	2.72 3.21	5.33 6.57	15.18 19.24	16.28 20.65	15.23 19.31	5.12 6.30	
Coefficient of	Retar	Maximum Mir	17.71	7.95	6.16	4.92	3.88	1.20	3.64	7.64	22.74	24.42	22.82	7.32	
	Effective	Porosity d/	0.20	0.20	0.20	0.20	0.20	0.20	0.20	0.20	0.20	0.20	0.20	0.20	
Bulk	Density	$(kg/L)^{d'}$	1.65	1.65	1.65	1.65	1.65	1.65	1.65	1.65	1.65	1.65	1.65	1.65	
ient		Average ^{c3/}	1.699	0.707	0.525	0.398	0.293	0.020	0.268	0.675	2.211	2.382	2.219	0.642	
Distribution Coefficient	K _d (L/kg)	Minimum ^{c2/}	1.321	0.550	0.408	0.310	0.228	0.015	0.209	0.525	1.719	1.852	1.725	0.499	
Distri		Maximum ^{c1/}	2.025	0.843	0.626	0.475	0.349	0.024	0.320	0.804	2.636	2.839	2.645	0.766	
Average Fraction	Organic	Carbon b'	0.00813	0.00813	0.00813	0.00813	0.00813	0.00813	0.00813	0.00813	0.00813	0.00813	0.00813	0.00813	
Minimum	Organic	Carbon b/	0.00632	0.00632	0.00632	0.00632	0.00632	0.00632	0.00632	0.00632	0.00632	0.00632	0.00632	0.00632	
Maximum Fraction	Organic		0.00969	0.00969	0.00969	0.00969	0.00969	0.00969	0.00969	0.00969	0.00969	0.00969	0.00969	0.00969	
	X,	(L/kg ^{a/})	500	87	64.6	49	36	2.45	33	83	272	293	273	62	
		Compound	PCE	TCE	1,1-DCE	cis-1,2-DCE	trans-1,2-DCE	ΛC	1,2-DCA	Chlorobenzene	1,2-DCB	1,3-DCB	1,4-DCB	Benzene	
		Location	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	USZ Sand	

^a From technical protocol (Wiedemeier et al., 1996a)

b' From laboratory analyses of site soil samples

 $^{^{\}rm cl'}$ K_{d} = Maximum Fraction Organic Carbon x K_{oc}

 $^{^{\}rm c2'}$ $\rm K_d$ = Minimum Fraction Organic Carbon x $\rm K_{oc}$

 $^{^{\}rm c3/}$ $\rm K_d$ = Average Fraction Organic Carbon x $\rm K_{oc}$

^d Estimated Value.

5.5 MODEL CALIBRATION

Model calibration is an important component in the development of any numerical groundwater model. Calibration of the flow model demonstrates that the model is capable of matching hydraulic conditions observed at the site; calibration of a contaminant transport model (using the calibrated flow field as input) helps demonstrate that contaminant loading and transport conditions are being appropriately simulated. Model input and output summaries are included in Appendix F. Due to the discrepancy noted previously between the direction of maximum hydraulic gradient and the distribution of the TCE contaminant plume, calibration of the groundwater flow model and solute transport model was a initially a simultaneous process. Final calibration of the solute transport model followed final calibration of the groundwater flow model.

5.5.1 Groundwater Flow Model

Groundwater elevation data collected in August 1997 and presented on Figure 3.10 were used to calibrate the flow model. Water level elevation data from 20 monitoring wells/points were used to compare measured and simulated heads for calibration. The selected locations are listed in Appendix E.

The numerical flow model was calibrated by altering hydraulic conductivity and constant-head elevations at the model boundaries in a trial-and-error fashion until simulated heads approximated observed field values within a prescribed accuracy. In general, hydraulic conductivities were varied within the limits of the observed data, although some values were increased slightly beyond the measured range. The flow model was initially calibrated as a steady-state model; later, it was run as a transient model to facilitate the transport solution. The steady-state heads were used as initial heads for the transient model, and the heads did not change significantly because the transient system quickly matched the steady-state system.

Figure 5.2 shows the calibrated water table for layer 1. Calibrated model horizontal hydraulic conductivities ranged between 0.01 and 36 ft/day. In general, the highest conductivity values (10 to 36 ft/day) were used for the USZ sands along the contaminant transport pathway (channels), a lower value (0.01 ft/day) was used for the intra-USZ clay layer, and intermediate values (1.5 to 2.25 ft/day) were used for the remainder of the medium. For all materials other than the layer 2 clay, the anisotropy ratio of y-direction (north-south) horizontal conductivity to x-direction horizontal conductivity was 2:1. Anistropy under the block-centered-flow option in MODFLOW also was set to 6.0 for layer 1 and 3.0 for layer 3. Hydraulic conductivity distributions for each layer are presented in Appendix E. For the majority of the medium, calibrated vertical hydraulic conductivities were in the range of 0.015 to 0.24 ft/day. To account for an apparent lack of vertical TCE mixing at some near-source locations, clay layer vertical conductivities were assigned negligible values. Similarly, to account for apparent downward TCE movement in certain downplume locations, vertical conductivities were increased (to 27 to 36 ft/day) in restricted zones for all three layers.

The root mean squared (RMS) error is commonly used to express the average difference between simulated and measured heads. RMS error is the average of the squared differences between measured and simulated heads, and can be expressed as:

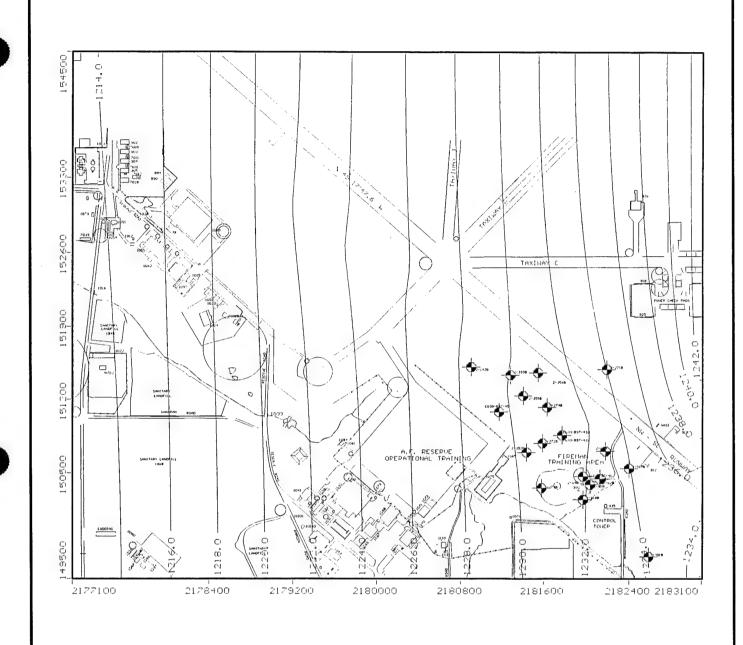


FIGURE 5.2

CALIBRATED GROUNDWATER SURFACE, LAYER 1

Site FTA-2 RNA TS Tinker AFB, Oklahoma

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

RMS =
$$\left[\frac{1}{n}\sum_{i=1}^{n}(h_{m}-h_{s})_{i}^{2}\right]^{0.5}$$

where:

n = the number of points where heads are being compared,

 h_m = measured head value, and

 $h_c = \text{simulated head value}$.

The RMS error between observed and calibrated values at 20 well locations was 0.67 feet, which corresponds to a calibration error of 11.1 percent. The calibration error is calculated using the change in relief for the area of the calibration wells only. Figure 5.3 shows the relation between measured heads and simulated heads and the resulting RMS. This plot provides a qualitative method of checking the calibrated head distribution; the points should scatter randomly about the straight line (Anderson and Woessner, 1992). In general, this is the case for the FTA-2 model.

In solving the groundwater flow equation, Visual MODFLOW establishes the water table surface and calculates an overall hydraulic balance that accounts for the numerical difference between flux into and out of the system. To the number of decimal places reported by Visual MODFLOW, the hydraulic mass balance for the steady-state calibrated flow model had a 0.0 percent discrepancy. This is more than adequate to accomplish the objectives of the modeling effort. According to Anderson and Woessner (1992), a mass balance error of around 1 percent is acceptable, while Konikow (1978) indicates an error of less than 0.1 percent is ideal.

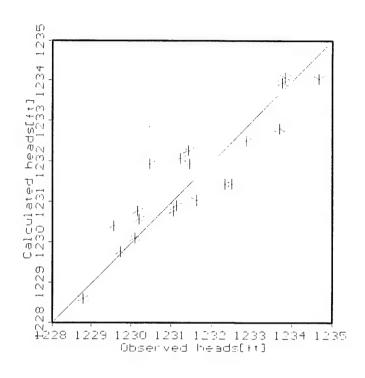
5.5.2 Plume Calibration

After calibration of the final flow model, the numerical solute transport model was calibrated by altering the contaminant specified-flux rate in the source cells and contaminant transport parameters in a trial-and-error fashion until the simulated plume approximated observed field values. For plume calibration, the model was run for a 35-year period (1962 to 1997). The transport parameters varied during plume calibration were the source-cell loading concentrations, the plume dispersivity, the TCE decay rate constant, the distribution coefficient for contaminant sorption, and effective porosity.

The dissolved TCE concentrations obtained from August 1997 laboratory analytical data from 20 monitoring wells/points were used to calibrate the contaminant transport model. Calibration sample locations and names are presented in Appendix E.

5.5.2.1 Source Concentrations

The 35-year time period was divided into five-year increments and the concentration history at each source cell was specified as a step function, with constant concentration specified for any given period. The four source cells are adjacent to each other and well 2-62B (Appendix E), where the highest TCE concentrations have been found (9,440 μ g/L in 1997). The source cells lie along a line trending north-northeast from well 2-62B. Of the four cells, the major contributor of dissolved TCE was the third cell along this line (i.e., the second farthest from well 2-62B). At this cell, specified source concentrations steadily declined from a high of 75,000 μ g/L in the first five-year



Period: 7 Step: 10

Mean error: 0.00233468 Mean abs. err: 0.558848 PMS error: 0.670091

FIGURE 5.3 **CALIBRATED VERSUS OBSERVED GROUNDWATER ELEVATIONS**

Site FTA-2 **RNA TS** Tinker AFB, Oklahoma

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

period to 20,000 μ g/L for the period 1993-1997 (years 30 through 35). The source cell farthest from well 2-62B was the next largest TCE contributor, remaining at a constant 9,500 μ g/L concentration for the 35-year calibration period. The remaining two source cells began at zero concentration for the first five-year period, built steadily to a 9,500 μ g/L concentration for years 16 through 20, and remained at that concentration for the rest of the calibration period. All specified source concentrations were at least an order of magnitude below the solubility limit for TCE in water (greater than 1,000,000 μ g/L).

5.5.2.2 Dispersivity

The initial longitudinal dispersivity of 160 feet produced a simulated plume far longer and wider than that observed in 1997. This value was gradually decreased to 20 feet. For the great majority of the model domain, this is the value used to produce the calibrated plume. Local variations were specified to obtain a better match to the calibration target. These consisted of reducing dispersivity beneath the source area and adjacent downplume areas to further inhibit any vertical dispersion of TCE. Lateral-to-longitudinal and vertical-to-longitudinal dispersivity ratios were varied by layer. For layers 1, 2, and 3, respectively, the calibrated lateral anisotropy ratios were 0.1, 0.2, and 0.22, and calibrated vertical anisotropy ratios were 0.01, 0.02, and 0.05.

5.5.2.3 TCE Decay Rates

The decay rate was varied during plume calibration, and was important in controlling plume length and concentration distributions. Two major vertical zones of constant decay rate were defined for the calibrated model: layers 2 and 3 had a rate of 0.00003 day (63-year half-life) and the majority of layer 1 had a 0.0001 day rate (19-year half-life). Layer 1 was further refined by specifying a higher 0.00034 day degradation rate (5.6-year half-life) for the near-source and plumes core area, and a surrounding zone at the intermediate rate of 0.0002 day (9.5-year half-life). The distribution of these zones improved the transport calibration and is in agreement with conclusions reached in Section 4.5. The higher rates in the near-source and plume core area are reflective of Type 1 behavior, while more aerobic downplume environments exhibit Type 3 behavior and have lower rates. It should be noted that zones of degradation actually evolve through time, whereas here they are simulated as static. Unfortunately, the process of adapting a general-purpose transport code to handle such non-linearities is highly problem-dependent and time-consuming. The rates obtained from the calibration process should be viewed as averages over the calibration period.

5.5.2.4 Sorption

During plume calibration, the retardation coefficient was varied, and the calibrated model used a value of 1.9 throughout the model domain. This value conservatively falls below the range estimated on the basis of observed TOC concentrations. However, sorption of any constituent can vary substantially among soils with similar organic content. Domenico and Schwartz (1990) and Pankow and Cherry (1996) note that sorption of organic compounds estimated on the basis of TOC alone is often underestimated because the role of sorption onto clays and other mineral surfaces is ignored. In addition, Ball and Roberts (1991) note that partitioning coefficients (used to calculate retardation coefficients in conjunction with TOC data) can be underestimated

because the laboratory studies used to derive the partitioning coefficients are performed on a short time scale (days to months). In reality, most field-scale situations (such as FTA-2) involve time scales of tens of years. While it was found that the retardation coefficients used in the model resulted in a reasonable calibration, there may be other combinations of transport parameters involving higher sorption that match measured TCE concentrations equally well.

5.5.2.5 Effective Porosity

The initial uniform effective porosity value of 0.2 was found to give reasonable calibration results, particularly for layer 1. After examination of calibration results and iterative variation, it was found that assigning a slightly lower effective porosity of 0.15 to layer 3 improved the calibration in the lower USZ sand.

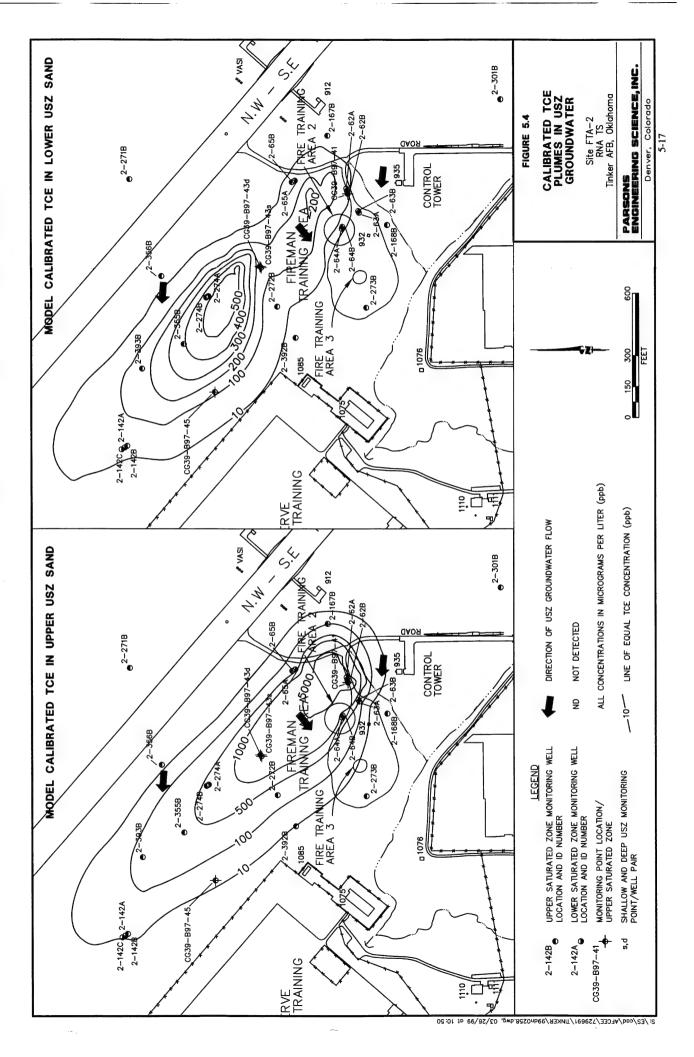
5.5.2.6 Transport Calibration Results

The calibrated layer 1 and 3 plumes calculated by the model (Figure 5.4) are similar, but not identical, to the observed August 1997 plumes (Figure 4.3). The overall shape and TCE concentration levels of the plumes are good. The RMS error of prediction over the 20 observations wells was $74.4 \mu g/L$.

In general, simulated TCE concentrations are a relatively close match to measured concentrations along the plume axis to a downplume location between wells 2-355B and 2-393B. Beyond this point, predictions were low at well 2-393B (177 μ g/L vs. 406 μ g/L measured), and high further downplume at well 2-142B (13 μ g/L vs. non-detect). Measured concentrations at well 2-393B were taken in the lower USZ sand and seem anomalously high, given the nearest upstream, downstream, and overlying measurements. It was found that no reasonable combination of transport parameters could reproduce the measured 406 μ g/L concentration at well 2-393B while maintaining an adequate calibration elsewhere. This points to the possibility of TCE sources other than those included in the model, or perhaps faster transport along preferred pathways with less attenuation than indicated by the calibrated model.

In general, compared to the calibration targets, model predictions are slightly high in layer 1 and slightly low in layer 3. Prediction error is largest at the lateral plume margins where steep concentration gradients from the plume core to outlying undetectable levels exist. In layer 1, it appears that the predicted plume is slightly wider in the cross-plume direction than the measured plume. The inability to reproduce a narrower layer 1 plume while maintaining an accurate calibration overall may be due to numerical dispersion. Numerical dispersion is particularly evident when transport is in a direction diagonal to grid coordinates, as is the case here.

Given the history of specified source area concentrations, it appears that adequate TCE mass was introduced to the system. Areas enclosed by isopleths appear to be very similar for simulated and measured TCE concentrations. This is important to note because the introduction of sufficient contaminant mass facilitates conservative predictions of downgradient receptor impacts and of plume persistence by accurately reproducing TCE concentrations in the downgradient portions of the plume. Modeling results also indicate a nonsteady-state plume.



5.6 SENSITIVITY ANALYSIS

The purpose of a sensitivity analysis is to determine the effect of varying model input parameters on model output. The sensitivity analysis was conducted by varying sorption (i.e., the distribution coefficient K_d), the first-order decay rate, dispersivity, hydraulic conductivity, and effective porosity.

To perform the sensitivity analyses, the calibrated model was adjusted by systematically changing the aforementioned parameters individually, and then comparing the new simulations to the results of the calibrated model. The models were run for a 35-year period, just as the calibrated model was, so that the independent effect of each variable could be assessed. Ten sensitivity runs of the calibrated model were made, with the following variations:

- 1. All hydraulic conductivities doubled;
- 2. All hydraulic conductivities halved;
- 3. Distribution coefficients doubled, which resulted in retardation coefficient set to 2.87;
- 4. Distribution coefficients halved, which resulted in retardation coefficient set to 1.4675;
- 5. Longitudinal dispersivities doubled;
- 6. Longitudinal dispersivities halved;
- 7. Decay rates doubled;
- 8. Decay rates halved;
- 9. Effective porosities doubled;
- 10. Effective porosities halved;

The results of the sensitivity analyses are discussed in the following subsections and summarized in Table 5.3. As described in the following paragraphs, the parameter modifications listed above generally caused substantial changes in the resulting plumes. The plume shape and distribution are most sensitive to changes in hydraulic conductivity and least sensitive to changes in longitudinal dispersivity.

5.6.1 Sensitivity to Variations in Hydraulic Conductivity

Hydraulic conductivity is an important aquifer characteristic that represents the ability of the water-bearing strata to transmit groundwater. An accurate estimate of hydraulic conductivity is important to help quantify advective groundwater flow velocities and to define the flushing potential of the aquifer. As a result, models used to estimate contaminant transport are particularly sensitive to variations in hydraulic conductivity. Lower values of hydraulic conductivity result in a slower-moving plume.

SUMMARY OF SENSITIVITY ANALYSIS RESULTS TINKER AFB, OKLAHOMA SITE FTA-2 RNA TS TABLE 5.3

	Observed	Calibrated										
Well	Concentration	Concentration	¥	В	ပ	Ω	ш	щ	Ö	Н	П	Ь
	(hg/l)	(l/gr/l)										
2-62B	9440	9500	9500	9500	9500	9500	9500	9500	9500	9500	9500	9500
CG39-B97-43s	1490	1581	2639	513	1250	1745	1559	1625	466	3178	089	2423
2-64B	914	924	2670	179	553	1196	1240	729	246	2066	270	2001
2-274B	518	470	897	60.2	290	552	536	423	127	974	131	816
2-393B	406	177	617	1.13	38.9	287	187	176	42.32	377	8.74	494
2-355B	344	417	1046	13.2	166	557	424	430	94.1	929	51.0	892
CG39-B97-43d	169	162	327	33.3	103	199	356	78.3	63.2	280	57.9	289
CG39-B97-45	153	115	1012	0.15	11.3	288	148	88.9	33.7	220	2.03	583
2-65B	8	18.1	30.2	4.78	11.8	21.7	99.4	2.61	10.1	25.1	7.50	27.1
2-63B	75	85.5	316	00.00	37.2	125	192	38.7	16.9	177	7.69	214
2-272B	28	211	1349	4.11	58.0	375	277	172	28.6	612	14.4	838
CG39-B97-41	4	8.39	30.6	1.57	3.87	13.3	45.4	1.60	6.28	98.6	2.61	18.5
2-168B	1	0.00	09.0	0.00	0.00	0.01	0.00	00.0	0.00	00.0	0.00	0.11
2-167B	QN	47.0	55.8	27.6	43.2	47.9	278	5.45	22.8	71.2	31.3	56.6
2-142B	QN	13.1	432	00.00	0.33	60.4	25.4	7.45	3.06	27.8	0.02	181
2-356B	ND	5.89	38.06	0.01	08.0	12.5	14.0	2.62	1.24	13.2	0.13	25.6
2-392B	ND ND	1.03	111	0.00	0.03	6.52	3.64	0.36	0.40	1.72	0.00	19.7
2-273B	ND	0.58	107	0.00	0.03	3.09	3.29	0.15	0.07	1.73	0.00	15.9
2-271B	ND	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
2-301B	ND	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
RMS Error (µg/l)		92	627	319	149	135	134	92	307	506	275	410

Column: A: Hydraulic Conductivity x 2.

B: Hydraulic Conductivity ÷ 2.

C: Kd x 2, Retardation = 2.87.

D: Kd \div 2, Retardation = 1.4675.

E: Dispersivity x 2.

F: Dispersivity ÷ 2.

G: Decay Rate x 2.

H: Decay Rate ÷ 2.

I: Effective Porosity x 2.

J: Effective Porosity ÷ 2.

Higher values of hydraulic conductivity result in a faster-moving plume. The effects of varying hydraulic conductivity are summarized in columns A and B of Table 5.3.

Uniformly increasing the hydraulic conductivity by a factor of two greatly increased the extent of the plume so that the 10- μ g/L isopleth reached approximately 1,100 feet further downstream than for the calibrated case, and the simulated concentrations downgradient of the source area were much higher than the calibrated model. In contrast, decreasing the hydraulic conductivity by a factor of two slowed overall plume migration, and resulted in a much shorter plume with contaminant mass occupying a much smaller area.

5.6.2 Sensitivity to Variations in the Distribution Coefficient

The effects of varying the distribution coefficient (K_d) are summarized in columns C and D of Table 5.3. Uniformly doubling K_d 's increased the retardation factor R to 2.87. This increase produced shorter and slightly narrower plumes, with concentrations outside of the source area below those in the calibrated model. This reflects the increased mass of TCE sorbed to the soil matrix. Uniformly halving K_d resulted in a retardation factor of 1.47. This significantly increased the length and width of the TCE plume and increased TCE concentrations outside of the source area.

5.6.3 Sensitivity to Variations in Dispersivity

The effects of varying dispersivity are summarized in columns E and F of Table 5.3. Longitudinal, transverse, and vertical dispersivity were all varied for this analysis, as ratios of longitudinal-to-transverse dispersivity and longitudinal-to-vertical dispersivity were kept at their calibrated values. Doubling and halving longitudinal dispersivity then resulted in doubled and halved transverse components of dispersivity.

Doubling the dispersivity components lowered peak TCE concentrations and spread more of the contaminant to the margins of the plume. The result was a plume with a slightly larger footprint, lower peak values, and slightly smaller gradients. Accordingly, concentrations at all wells increased except at well CG39-B97-43s in the heart of the plume, which decreased from 1,581 to 1,559 μ g/L.

Halving the dispersivity components had the opposite effect, resulting in a slightly smaller plume with higher peak concentrations, lower marginal concentrations, and slightly steeper gradients. Compared to the increased dispersivity case, and all other sensitivity cases, the effect of halving dispersivity produced the least amount of change from the calibrated case.

5.6.4 Sensitivity to Variations in the Decay Rate Constant

The effects of varying the first-order TCE decay rate are summarized in columns G and H of Table 5.3. Uniformly doubling decay rates resulted in rapid degradation of dissolved contaminants, producing a much smaller plume with lower concentrations everywhere outside the source. Uniformly halving the degradation rates resulted in a much larger plume with higher concentrations outside the source area than in the calibrated model.

5.6.5 Sensitivity to Variations in Effective Porosity

The results of doubling and halving effective porosity are summarized in columns I and J of Table 5.3. Doubling effective porosity halved the interstitial flow velocity and resulted in much slower transport. With more water in the system, dilution was increased and concentrations were lower than calibrated values everywhere outside the source. The opposite effect resulted from halving the effective porosity. Transport was sped up and dilution diminished, yielding a much larger plume with higher concentrations everywhere outside the source.

5.6.6 Summary of Sensitivity Analysis Results

The results of the sensitivity analysis show that, compared to the calibrated model, all parameter variations caused a significant degradation in the goodness-of-fit between predictions and measurements. This suggests that, if the underlying conceptual model is appropriate, the calibrated model discussed in Section 5.4 is a reasonable and useful approximation of site conditions.

Based on the sensitivity analysis, transport parameters can be ordered from most to least significant as follows: hydraulic conductivity, degradation rate, effective porosity, distribution coefficient, and dispersivity. More specifically, the parameter variations can be ranked according to the degree of deviation from the calibrated "fit" they cause. From the greatest deviation to the least, these are

- 1. Doubling hydraulic conductivity
- 2. Halving decay rate constant
- 3. Halving effective porosity
- 4. Halving hydraulic conductivity
- 5. Doubling decay rate constant
- 6. Doubling effective porosity
- 7. Doubling distribution coefficient
- 8. Halving distribution coefficient
- 9. Doubling dispersivity
- 10. Halving dispersivity

It should be kept in mind that typical percentage variations in field-measured values of these parameters are site-specific. For example, at a particular site effective porosity measurements may vary from the mean by less than a factor of two, while the distribution coefficient may vary over orders of magnitude. In this case, a model would be more sensitive to "typical" variations in distribution coefficient than to "typical" variations in

effective porosity, even if, for the same percentage change, the model were determined to be more sensitive to effective porosity.

5.7 DISCUSSION OF THE MODEL AND RECOMMENDATIONS TO REDUCE MODEL UNCERTAINTIES

Before discussing model predictions in Section 6, some points about the model should be emphasized. Most notably, the system that is modeled is very complex, and in many areas the relevant properties and parameters are not well defined. In addition, model computation time increases greatly as additional complexity or detail is added. Reasonable calibration results have been obtained, indicating that a proper balance was struck between model complexity, required model outputs, available input data, and numerical efficiency.

Given the underlying conceptual model of a compact TCE source near well 2-62B and downplume transport at an oblique angle to the downgradient hydraulic direction, the calibration exercise has yielded a useful tool for evaluation of future TCE transport under different remedial alternatives. However, large data uncertainties leave open the possibility that alternative conceptual models may explain the TCE plume as well or better. Among factors contributing to these uncertainties are 1) the unknown locations and history of CAH disposal at FTA-2; 2) inability to find any obvious contaminated soil "hot spots"; 3) the relatively wide range of distribution coefficients and degradation rates indicated by available field data; 4) apparent lack of subsurface structural features commonly associated with horizontal flow anisotropy; and 5) inability to discern the local change in water table gradient that would accompany preferred flow along permeable channels.

Given these facts, it is possible, that the observed TCE plume is the result of typical downgradient transport from multiple surface sources stretched along the plume length. In light of the plume's configuration and downgradient location, perhaps use of solvents to remove tire rubber from the NW-SE Runway has partially contributed to the plume. Releases responsible for the plume could have occurred in the 1960's, followed by downgradient transport with higher sorption, higher degradation rates, and/or higher effective porosities resulting in shorter migration distances in the southwest direction. Alternatively, such releases could have occurred more recently, with less time to migrate to the southwest.

Several approaches could be used to definitively answer the questions arising from data gaps at FTA-2. The first and most effective would be to conduct conservative tracer tests, injecting at well 2-62B and monitoring at wells 2-64A/B and 2-63A/B. Travel times between these wells are on the order of 100 days, and detection of significant amounts of tracer, particularly at 2-63A/B, would indicate transport to the southwest and call into question the current conceptual model. A similar tracer test could inject at CG39-B97-43s and monitor at 2-272B, although due to the distance involved a 200-day time frame would be involved. New or temporary wells could be placed to shorten the time frame and improve the spatial resolution of tracer tests.

Similarly, additional monitoring wells could be installed in hopes of detecting local gradient changes indicating the presence of preferred pathways. Soil gas surveys could

be performed, particularly upgradient of plume hot spots, in hopes of identifying previously unknown source areas. Finally, geophysical surveys could be used to attempt to identify structural features that might contribute to horizontal flow anisotropy. A combination of these approaches should provide enough evidence to clearly define the most appropriate conceptual model, and add confidence that model simulations of remedial alternatives are accurate.

SECTION 6

COMPARATIVE ANALYSIS OF REMEDIAL ALTERNATIVES

This section presents the development and comparative analysis of remedial alternatives for contaminated groundwater at Site FTA-2. The intent of this evaluation is to determine if RNA of groundwater is an appropriate and cost-effective remedial approach to consider when developing final remedial strategies for the study area, especially when combined with other remedial technologies.

Section 6.1 presents the criteria used to evaluate groundwater remedial alternatives. Section 6.2 discusses the development of remedial alternatives considered as part of this project. Section 6.3 provides a brief description of each of these remedial alternatives, and Section 6.4 provides a more detailed analysis of the remedial alternatives using the defined remedial alternative evaluation criteria. Results of this evaluation process are summarized in Section 6.5.

6.1 REMEDIAL ALTERNATIVE EVALUATION CRITERIA

The evaluation criteria used to identify appropriate remedial alternatives for USZ groundwater contamination at the site were evaluated on the basis of (1) long-term effectiveness and permanence, (2) technical and administrative implementability, and (3) relative cost. The following sections briefly describe the scope and purpose of each evaluation criterion. This report focuses on the potential use of RNA to reduce dissolved chlorinated solvent concentrations in the USZ groundwater below MCLs.

6.1.1 Long-Term Effectiveness and Permanence

Each remedial alternative (which can be a combination of remedial approaches such as RNA and institutional controls) was analyzed to determine how effectively it would minimize groundwater plume expansion so groundwater quality standards can be achieved at a downgradient sentry well. The expected technical effectiveness based on case histories from other sites with similar conditions also is evaluated. The ability to minimize potential impacts on surrounding facilities and operations is considered. Also, the ability of each remedial alternative to protect both current and potential future receptors from potential risks associated with potentially completed exposure pathways is qualitatively assessed. The evaluation criterion also include permanence and the ability to reduce contaminant mass, toxicity, and volume. Long-term reliability for providing continued protection, including an assessment of potential technology failure and potential threats resulting from such a failure are also evaluated.

6.1.2 Implementability

The technical implementability of each remedial technology or remedial alternative was evaluated in terms of technical feasibility and availability. Potential shortcomings and difficulties in construction, operations, and monitoring are presented and weighed against perceived benefits. Requirements for any post-implementation site controls such as LTM and land use restrictions are described. Details on administrative feasibility in terms of the likelihood of public acceptance and the ability to obtain necessary approvals are discussed.

6.1.3 Cost

The total cost (adjusted to present worth) of each remedial alternative was estimated for relative comparison. Estimates of capital costs and operating and post-implementation costs for site monitoring and controls are included. An annual adjustment (discount) factor of seven percent was assumed in present worth calculations. The annual adjustment factor is the difference between the rate of inflation and the cost of money (USEPA, 1993).

6.2 FACTORS INFLUENCING ALTERNATIVES DEVELOPMENT

Several factors were considered during the identification and screening of remedial technologies for addressing USZ groundwater contamination at the site. Factors considered included the objectives of the AFCEE natural attenuation demonstration program; contaminant, groundwater, and soil properties; current and future land uses; and potential receptors and exposure pathways. The following section briefly describes each of these factors and how they were used to narrow the list of potentially applicable remedial technologies.

6.2.1 Program Objectives

The intent of the RNA demonstration program sponsored by AFCEE is to develop a systematic process for scientifically investigating and documenting natural subsurface attenuation processes that can be factored into overall site remediation plans. The objective of this program and the specific demonstration at Tinker AFB is to provide solid evidence of RNA of dissolved fuel hydrocarbons and chlorinated solvents so that this information can be used to develop an effective groundwater remediation strategy. A secondary goal of this multi-site initiative is to provide a series of regional case studies, which demonstrate natural processes of contaminant degradation can often reduce contaminant concentrations in groundwater to below acceptable cleanup standards before completion of potential receptor exposure pathways.

Because the objective of this program is to study natural processes in the saturated zone rather than all contaminated media (e.g., soil, soil gas, etc.), technologies have been evaluated based primarily on their potential impact to shallow groundwater and phreatic soils. Technologies that can reduce vadose zone contamination and partitioning of contaminants into groundwater also were evaluated. Many of the source removal technologies evaluated in this section will also reduce soil and soil gas contamination, but it is important to emphasize that the remedial alternatives developed in this document are

not intended to remediate all contaminated media. Additional program objectives set forth by AFCEE include cost effectiveness and waste minimization.

6.2.2 Contaminant Properties

Site-related groundwater contaminants exceeding USEPA MCLs and considered for remediation as part of this demonstration are PCE, TCE, 1,1-DCE, cis-1,2-DCE, VC, and 1,2-DCA. Benzene, CB, 1,2-DCB, and 1,4-DCB also exceed respective MCLs, but appear to rapidly attenuate away from the source area and are not considered as part of this remedial evaluation. The primary sources of contamination at FTA-2 are likely spills or dumping of waste scattered throughout the fire training area; however, the exact location and nature of these sources is unknown. The physiochemical characteristics of the individual compounds will greatly influence the effectiveness and selection of a remedial technology.

TCE was used as the target remediation compound at FTA-2. The summary of biodegradation rates in Table 4.7 by the method of Buscheck and Alcantar (1995) indicates that the transformation rate of TCE to daughter products and the transformation of total chlorinated ethenes to daughter products are nearly identical (halve-lives of 6.20 and 6.42 years, respectively). The maximum concentration of cis-1,2-DCE coincides with the maximum concentration of TCE, although at a concentration approximately seven to eight times lower than TCE. The relatively high TCE concentrations support its use as a surrogate for all CAH compounds at this site during the remedial alternative evaluation process. Modeling of TCE fate and transport (Section 5) also allows for comparative modeling of selected remedial alternatives.

All the CAH compounds (e.g., PCE, TCE, DCE, VC, and DCA) are generally volatile, relatively soluble in water, and do not adsorb strongly to soil particles. These characteristics allow the CAHs to leach readily from NAPL and contaminated soil into groundwater and to migrate as dissolved contamination (Lyman *et al.*, 1992). Many CAHs are amenable to *in situ* degradation by both biotic and abiotic mechanisms under certain geochemical conditions.

TCE is moderately volatile, with a vapor pressure ranging from 69 millimeters of mercury (mm Hg) to 74 mm Hg at 25°C (Mackay and Shiu, 1981; Klopffer *et al.*, 1988; Howard, 1990). Henry's Law Constants reported for TCE range from 0.0086 to 0.0102 atm-m³/mol at 25°C (Ashworth *et al.*, 1988; Dewulf *et al.*, 1995). The solubility of pure TCE in water at 20°C has been reported to be 1,100 mg/L (Pearson and McConnell, 1975). TCE biodegradation products vary according to the prevailing groundwater geochemistry and are described in Section 4.1.

On the basis of these physiochemical characteristics, natural attenuation, soil vapor extraction, air sparging, and groundwater extraction and treatment could be effective options for collecting, destroying, and/or treating dissolved CAHs at Site FTA-2. In addition, *in situ* degradation of CAHs via permeable reaction walls (e.g., iron filings trenched) is a developing technology that holds promise for treatment of CAHs dissolved in groundwater. Some of these options are considered less desirable after considering site-specific conditions (Section 6.2.3) and program objectives (Section 6.2.1).

6.2.3 Site-Specific Conditions

Three general categories of site-specific characteristics were considered when identifying remedial approaches for comparative evaluation. The first category was physical characteristics such as groundwater depth, hydraulic conductivity, hydraulic gradient, flow direction, and soil type. The second category was the site geochemistry, or how the site contaminants are interacting with electron acceptors, microorganisms, and other site contaminants. Both of these categories influence the types of remedial technologies most appropriate for the site. The third category involved assumptions about future land use and potential receptor exposure pathways. Each of these site-specific characteristics has influenced the development of remedial alternatives included in the comparative evaluation.

6.2.3.1 Physical Characteristics

Site geology and hydrogeology have a profound effect on the transport of contaminants and the effectiveness and scope of required remedial technologies at a given site. Hydraulic conductivity is perhaps the most important aquifer parameter governing groundwater flow and contaminant transport in the subsurface. The velocity of the groundwater and dissolved contamination is directly related to the hydraulic conductivity of the saturated zone. The estimated average advective groundwater velocities at FTA-2 are 152 ft/yr (upper USZ sand) and 167 ft/yr (lower USZ sand) (Table 3.2).

Although the relatively high hydraulic conductivity of the study area can result in greater plume expansion and migration, this same characteristic also can enhance the effectiveness of other remedial technologies, such as groundwater extraction, air sparging, and intrinsic remediation. For example, it should be less expensive and time-consuming to capture and treat a contaminant plume using a network of extraction wells in an area of high hydraulic conductivity, because each well could envelope a larger area of influence and sustain a higher flow rate. The effectiveness of air sparging also may be increased in highly conductive and/or homogeneous aquifers because of reduced entry pressures and short-circuiting, and increased mixing of sparge air and groundwater. In addition, greater hydraulic conductivity would increase the amount of contaminant mass traveling through an air sparging network. Given a moderately high groundwater velocity, the effectiveness of natural attenuation can increase due to enhanced dilution and dispersion of the contaminant mass. The movement of contaminant mass within the subsurface away from the source area also can bring contaminants into contact with a larger mass of electron donors and acceptors, thereby increasing rates of biodegradation.

Like hydraulic conductivity, the organic carbon content of native phreatic zone soils can influence the effectiveness of remedial alternatives. Soils in the study area have a moderate organic carbon content (approximately 0.813 percent, Table 4.1), and therefore, the soils have a correspondingly moderate sorptive potential. Soil TOC values greater than 0.01 percent are typically sufficient to cause contaminant retardation due to sorption. The difference between contaminant and groundwater velocity increases the effectiveness of biodegradation in the source area because there are fresh electron acceptors flowing past the slower moving dissolved contaminant mass.

6.2.3.2 Geochemical Characteristics

To satisfy the requirements of indigenous microbial activity and RNA, the aquifer must also provide an adequate and available carbon or energy source (e.g., fuel hydrocarbon contamination or natural carbon), electron acceptors, essential nutrients, and proper ranges of pH, temperature, and redox potential. Data collected as part of this TS (Sections 3 and 4) indicate only the FTA-2 source area contains adequate and available carbon/energy sources to support reductive dehalogenation of CAH contamination by indigenous microorganisms. Dissolved oxygen, nitrate, ferric iron, sulfate, carbon dioxide and less chlorinated CAHs represent sources of significant electron acceptor capacity for the biodegradation of organic compounds at FTA-2.

The pH in USZ groundwater ranged between 6.80 and 7.36 standard units in August 1997, which is within the optimal range for biodegradation of 6 to 8 standard pH units (Wiedemeier et al., 1995). Redox potentials ranged from -105 to 138.7 mV in August 1997 (Figure 4.15), and suggest a groundwater environment that is both oxidizing and reducing. The redox potentials at the site suggest that aerobic biodegradation, nitrate reduction, manganese reduction and iron (III) reduction could reduce contaminant concentrations in groundwater. Observed methane concentrations also indicate that methanogenesis could reduce contaminant concentrations in groundwater. Groundwater data presented in Section 4 strongly support the conclusion that aerobic and anaerobic processes may have reduced fuel hydrocarbon contamination given the current geochemical conditions. These same processes also appear to be acting upon chlorinated solvents within the general source area, both as electron donors and electron acceptors.

Microbe addition was not considered a viable remedial approach for this site on the basis of observed geochemical trends that suggest that significant microbial activity is likely occurring. Hydrocarbon-degrading microorganisms are ubiquitous in the subsurface, and a large number of hydrocarbon-degrading isolates (bacteria and fungi) have been discovered in different soil environments (see Wiedemeier *et al.*, 1995 and 1996a for a partial listing). Indigenous microorganisms have a distinct advantage over microorganisms injected into the subsurface to enhance biodegradation, because indigenous microorganisms are well adapted to the physical and chemical conditions of the subsurface in which they reside (Goldstein *et al.*, 1985).

6.2.3.3 Potential Receptor Exposure Pathways

A pathway analysis identifies the potential human and ecological receptors that could come into contact with site-related contamination and the pathways through which these receptors might be exposed. To have a completed exposure pathway, there must be a source of contamination, a potential mechanism(s) of release, a pathway of transport to an exposure point, an exposure point, and a receptor. If any of these elements do not exist, the exposure pathway is considered incomplete, and receptors will not come into contact with site-related contamination. Evaluation of the potential long-term effectiveness of any remedial technology or remedial alternative as part of this demonstration project includes determining the potential for pathway completion. If a completed exposure pathway exists, potential long-term remedial options may still be sufficient to maintain exposure concentrations below regulatory action levels. Establishing site-specific, risk-based cleanup levels is beyond the scope of this TS.

Assumptions about current and future land uses at a site form the basis for identifying potential receptors, potential exposure pathways, reasonable exposure scenarios, and appropriate remediation goals. USEPA (1991) advises that the land use associated with the highest (most conservative) potential level of exposure and risk that can reasonably be expected to occur should be used to guide the identification of potential exposure pathways and to determine the level to which a site must be remediated.

Groundwater beneath FTA-2 flows generally to the west, an area of light industrial use. Because the site is near industrial facilities for a secured military base, Base workers are the most probable potential receptors that could be exposed to any site-related contamination. Other receptors include Tinker AFB water supply wells located on the western portion of the Base.

6.2.3.4 Remediation Goals for USZ Groundwater

USEPA MCLs (1996) for compounds detected in FTA-2 groundwater that exceed these standards are presented in Table 6.1. For this TS, the primary remedial objective for evaluation of remediation technologies for shallow groundwater is to reduce contaminant concentrations in groundwater to below federal regulatory criteria at a downgradient point of compliance. To accomplish this, remedial alternatives focus on limiting migration of dissolved contaminant concentrations exceeding MCLs away from Site FTA-2. It is unlikely that groundwater from Site FTA-2 would be ingested by humans, because there are no current downgradient water supplies in close proximity to the site. However, the MCL of 5 μ g/L for TCE will be used to evaluate the effectiveness, implementability, and cost of remedial alternatives in this TS.

Available data suggest there are no completed exposure pathways involving human and ecological receptors exposed to contaminated groundwater under current conditions. Site FTA-2 is part of a secured military base, so institutional controls can be incorporated to limit any future intrusive activity at the site. The required period of any groundwater and soil institutional controls associated with the selected remedial alternative is likely to expire before any anticipated future land use changes. Off-base migration of dissolved contaminants in concentrations exceeding MCLs appears unlikely, given the groundwater flow direction and distance to the western Base boundary (greater than one mile, Figure 1.2).

6.2.4 Summary of Remedial Option Screening

Several remedial options have been identified and screened for use in treating shallow groundwater at the site. Table 6.2 identifies the initial remedial technologies and approaches considered as part of this demonstration and those retained for detailed comparative analysis. Screening was conducted systematically by considering the

TABLE 6.1 GROUNDWATER QUALITY STANDARDS

FTA-2 RNA TS TINKER AFB, OKLAHOMA

Compound	USEPA MCL (µg/L) ^{a/}
Benzene	5
Tetrachloroethene	5
Trichloroethene	5
1,1-Dichloroethene	7
Cis-1,2-Dichloroethene	70
Vinyl Chloride	2
1,2-Dichloroethane	5
Chlorobenzene	100
1,2-Dichlorobenzene	600
1,4-Dichlorobenzene	75

a/ USEPA MCL = US Environmental Protection Agency (1996) Maximum Contaminant Level.

program objectives of the AFCEE RNA demonstration, physiochemical properties of groundwater contaminants detected at FTA-2, and other site-specific characteristics such as hydrogeology, geochemistry, land use assumptions, potential receptor exposure pathways, and appropriate remediation goals. All of these factors will influence the technical effectiveness, implementation, and relative cost of technologies for remediating shallow groundwater underlying and migrating from the site. The remedial options retained for development of remedial alternatives and comparative analysis include institutional controls, natural attenuation, LTM, source characterization and removal, and groundwater extraction with air stripping for treatment of extracted groundwater.

INITIAL TECHNICAL IMPLEMENTABILITY SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR GROUNDWATER REMEDIATION TABLE 6.2

FTA-2 RNA TS TINKER AFB, OKLAHOMA

Retain	Yes	Yes	Yes	No	No	Yes	No	Yes	No	No	No
Relative Cost	Low	Low	Low	Low	Moderate	Low	High	Moderate	High	High	High
Effectiveness	Necessary for all remediation strategies	Necessary	Necessary	Not required at this site	Not required at this site	Necessary	Low	Moderate	Low	Low	Moderate
Implementability	Many monitoring wells are available to confirm the progress of remediation. Sufficient space exists for additional wells.	Sufficient distance exists between the plume and point-of-compliance to locate several wells.	The plume lies within the Base boundary, and land and groundwater use are under Base jurisdiction.	No production wells are known to exist in the current or predicted plume area.	No shallow groundwater is extracted from the plume area for any use.	Base public relations and environmental management offices have many information avenues through which to communicate to workers and residents.	Limited effectiveness and high cost due to depth of groundwater, depth to USZLSZ aquitard, and size of plume. Physical controls not required at this site.	Future migration and concentration of dissolved plume is projected to be significant. Does not comply with program objectives.	Limited effectiveness and high cost due to depth of groundwater, depth to USZ/LSZ aquitard, and size of plume. Physical controls not required at this site.	Limited effectiveness and high cost due to depth of groundwater, depth to USZ/LSZ aquitard, and size of plume. Physical controls not required at this site.	Degradation of CAHS may be stimulated by allowing groundwater to flow through a nutrient-rich zone or zero-valent zone. New, unproven technologies
Process Option	Confirmation Wells	Sentry Wells	Land Use Control/Regulate Well Permits	Seal/Abandon Existing Wells	Point-of-Use Treatment	Meetings/ Newsletters	Interceptor Trench Collection	Groundwater Extraction	Slurry Walls/Grout Curtains	Sheet Piling	Biologically Active Zones/ Iron Filings Trench
Technology Tvpe	Periodic Groundwater Monitoring	0	Groundwater Use Control			Public Education	Hydraulic Controls		Physical Controls		Reactive/ Semi- Permeable Barriers
General Response Action	Long-Term Monitoring		Institutional Controls				Containment of Plume				

TECHNOLOGIES AND PROCESS OPTIONS FOR GROUNDWATER REMEDIATION TABLE 6.2 (continued) INITIAL TECHNICAL IMPLEMENTABILITY SCREENING OF FTA-2 RNA TS TINKER AFB, OKLAHOMA

General	Technology	Process Option	Implementability	Effective-	Relative	Retain
In Situ Groundwater Treatment	Biological	Oxygen and/or Nutrient Enhanced Biodegradation (Biosparging)	Nutrients and/or oxygen are injected within and/or downgradient of plume to limit plume migration by enhancing biodegradation and reducing CAH concentrations. Will inhibit anaerobic degradation of CAHs by oxygenating the groundwater. TCE not aerobically biodegradable. Would be most effective downgradient of the source area. Not proven to be more effective than natural attenuation.	Low	Low	No
	Chemical/ Physical	Natural Attenuation	A combination of natural biological, chemical, and physical removal mechanisms which occur to varying degrees on every site. Groundwater sampling at FTA-2 indicates that this is an ongoing remediation process.	High	Low	Yes
		Air Sparging (Volatilization)	Injection of air into contaminated aquifer creating a mass transfer of CAHs, specifically TCE, into air bubbles and vadose zone. Similar to biosparging in effectiveness; however, more mass is transferred rather than destroyed. Most effectively used downgradient of the source so as not to upset anaerobic conditions promoting reductive dechlorination in the source area.	Low	Low	No
Source Removal/ Groundwater Remediation	Groundwater Extraction	Vertical Pumping Wells	Source area of groundwater plume is pumped by installing submersible pumps in source area wells. Limited by contaminant mass transfer rates into the groundwater.	Moderate	Moderate	Yes
Aboveground Groundwater Treatment	Biological	Bioreactors	High flow rates require excessive retention times and large reactors. CAHs are often volatilized in these systems.	Moderate	High	No
	Chemical/ Physical	Air Stripping	Cost-effective technology for removing CAHs from groundwater at high flow rates. Potential permitting for air emissions.	High	Moderate	Yes
	•	Activated Carbon	Cost prohibitive for more concentrated Cahs. Creates a carbon disposal problem.	Moderate	High (O&M)	No
		Direct Discharge to Industrial Waste Water	Implementable option only when an IWWTP is readily available and capable of handling CAHs and hydraulic loading. IWWTP not available for this site.	High	Low	No
		Treatment Plant (IWWTP)				
		UV/Ozone Reactors	High flow rates require lengthy retention times and large, costly reactors.	Moderate	High	No No

TECHNOLOGIES AND PROCESS OPTIONS FOR GROUNDWATER REMEDIATION TABLE 6.2 (Continued) INITIAL TECHNICAL IMPLEMENTABILITY SCREENING OF FTA-2 RNA TS TINKER AFB, OKLAHOMA

Retain	No	No	No	No	Yes	Yes	No	No	No	No
Relative Cost	Low	Low	Moderate	Moderate	Low	Moderate	Moderate	High	Low	Low to Moderate
Effectiveness	High	High	Moderate	Moderate	High	Moderate to High	Moderate to High	Moderate to High	Moderate to High	High
Implementability	Implementable option only when an IWWTP is available and capable of handling hydraulic loading. IWWTP is not available.	Implementable option only when access to a sanitary sewer exists and hydraulic loading is acceptable.	Injection wells subject to clogging, high maintenance, and permitting.	Requires large trenches and can be subject to injection well permitting.	Generally requires NPDES or other discharge permit.	Some excavation may economically be feasible if it can be shown that source area is generally less than 4,000 cubic yards.	Some excavation may be feasible at this site. For reasons of risk and cost reduction, <i>in-situ</i> methods or landfilling preferable when possible.	Some excavation may be feasible at this site. For reasons of risk and cost reduction, <i>in-situ</i> methods or landfilling preferable when possible.	Air injection/extraction to increase soil oxygen levels and stimulate biodegradation of fuel residuals. Conductivity of site soils to air movement would be greater than to water movement. Might stimulate cometabolism of CAHs during the biodegradation of fuel contamination.	Air extraction to increase soil oxygen levels and extract VOCs from vadose zone. Conductivity of site soils to air movement would be greater than to water movement. May be subject to air emissions permitting.
Process Option	IWWTP	Sanitary Sewer	Vertical Injection Wells	Injection Trenches	Storm Drains or Surface Ditch	Landfilling	Biological Landfarming	Thermal Desorption	Bioventing	Soil Vapor Extraction
Technology Tvpe	Discharge to IWWTP	Discharge to Sanitary Sewer	Treated Groundwater	Reinjection	Discharge to Surface Waters	Excavation/ Treatment			In Situ	,
General Response Action	Treated Groundwater Disposal	•				Source Removal/Soil Remediation				

TABLE 6.2 (Concluded) INITIAL TECHNICAL IMPLEMENTABILITY SCREENING OF TECHNOLOGIES AND PROCESS OPTIONS FOR GROUNDWATER REMEDIATION FTA-2 RNA TS TINKER AFB, OKLAHOMA

General	Technology	Process Option	Implementability	Effectiveness	Relative	Retain
Response Action	Type	•			Cost	
Source	In-Situ	Soil Washing	Water and/or surfactant solution are forced through zones of residual	Low	High	No No
Removal/Soil	(continued)		contamination to enhance contaminant partitioning into the groundwater.			
Remediation			Most effective in homogeneous sandy soils.			
(continued)						
	Mobile NAPL	Dual-Pump	NAPL has not been detected at the site.	Moderate	High	No No
	Recovery	Systems				
		Skimmer Pumps,	NAPL has not been detected at the site.	Moderate	Low to	No No
		Bailers, Wicks			Moderate	
		Total Fluids	NAPL has not been detected at the site.	Moderate	High	No
		Pumping				
		Vacuum-	NAPL has not been detected at the site.	Moderate to	<u>e</u>	Yes
		enhanced		High	to High	
		extraction				

6.3 BRIEF DESCRIPTION OF REMEDIAL ALTERNATIVES

This section describes how remedial technologies retained from the screening process were combined into three remedial alternatives for the study area. Sufficient information on each remedial alternative is provided to facilitate a comparative analysis of effectiveness, implementability, and cost in Section 6.4.

6.3.1 Alternative 1 - RNA and Institutional Controls with Long-Term Groundwater Monitoring

RNA is achieved when natural attenuation mechanisms bring about a reduction in the total mass or concentration of a contaminant in the soil or dissolved in groundwater. RNA results from the integration of several subsurface attenuation mechanisms that are classified as either destructive or nondestructive. Destructive attenuation mechanisms include biodegradation, abiotic oxidation, and hydrolysis. Nondestructive attenuation mechanisms include sorption, dilution (caused by dispersion and infiltration), and volatilization. In some cases, RNA will reduce dissolved contaminant concentrations below numerical concentration goals intended to be protective of human health and the environment. As indicated by the evidence of RNA described in Section 4, these processes are occurring in the FTA-2 source area and will continue to reduce contaminant mass.

Implementation of Alternative 1 would require installation of additional monitoring wells, the use of institutional controls such as land use restrictions, and LTM. Land use restrictions may include placing long-term restrictions on groundwater well installations within and downgradient from the plume area. The intent of these restrictions would be to reduce potential receptor exposure to contaminants by legally restricting activities within areas affected by site-related contamination.

Long-term monitoring would be performed at a regular frequency and would consist of sampling a set of wells, including upgradient, source area, and plume extent (Section 7). Additional sentry wells may be required as the plume continues to expand. Due to uncertainty with site hydrogeology, sentry well locations should be selected after additional LTM data are available. For this site, it is assumed that 35 years of monitoring (consisting of annual monitoring for 15 years followed by biennial monitoring for 20 years) would be needed to establish plume behavior (i.e., expanding, stable, or shrinking) and that contaminant concentrations are decreasing. The recommended site-specific LTM strategy is provided in Section 7. On the basis of predictive contaminant fate and transport model results (Section 6.4), it is unlikely that contaminant concentrations exceeding groundwater MCLs would approach a potential receptor exposure point. Nevertheless, LTM is the technical mechanism used to evaluate the progress of natural attenuation processes and to ensure that remedial objectives are being met. Detection of TCE, cis-1,2-DCE, 1,2-DCA, or any other compounds listed in Table 6.1 in excess of their individual MCLs at a sentry well may require additional evaluation to assess contaminant migration, to determine the probable extent of migration, and to determine if additional corrective action is necessary. In either case, land use restrictions would require reevaluation.

Public education on the selected alternative would be developed to inform Base personnel and residents of the scientific principles underlying source reduction and RNA. This education could be accomplished through public meetings, presentations, press releases, and posting of signs where appropriate. Periodic site reviews could also be conducted annually using data collected from the LTM program. The purpose of these periodic reviews would be to evaluate the extent of contamination, assess contaminant migration and attenuation through time, document the effectiveness of source removal and/or institutional controls at the site, and reevaluate the need for additional remedial actions at the site.

6.3.2 Alternative 2 – Source Characterization and Removal, RNA, and Institutional Controls with Long-Term Groundwater Monitoring

Additional source and site characterization activities are recommended for site FTA-2 to further investigate the apparent discrepancy between the groundwater flow direction and distribution of dissolved TCE. The numerical model presented in Section 5 assumes anisotropic hydraulic parameters are responsible for the differences between groundwater flow and contaminant transport directions. However, additional sources of TCE and 1,2-DCA northwest of the primary FTA-2 source area also are suspected. Alternative 2 therefore includes additional source and site characterization activities including a soil gas survey, a geophysical survey, additional slug tests, and tracer tests, followed by source removal and LTM. As with Alternative 1, institutional controls and LTM would be required. Sentry and LTM wells would be the same as described for Alternative 1, and it is assumed that 35 years of monitoring would be needed to establish whether the plume is stable or shrinking and at contaminant concentrations are decreasing.

6.3.3 Alternative 3 - Groundwater Extraction and Treatment, Source Characterization and Removal, RNA, and Institutional Controls with Long-Term Groundwater Monitoring

Alternative 3 for FTA-2 includes groundwater extraction and treatment for plume containment in addition to the site characterization and source removal initiatives proposed for Alternative 2. Five 4-inch groundwater extraction wells would be completed in the upper USZ sand interval and located in the area with the highest TCE concentrations near well 2-62. Two additional 4-inch groundwater extraction wells would be completed in the lower USZ sand interval and located in the area near well 2-274B. A pumping rate of five gallons per minute (gpm) was assigned to each extraction well. The groundwater extraction system will prevent higher concentrations of contaminated groundwater from migrating downgradient and reduce the potential for completion of any potential receptor pathways. Given the relatively high natural hydraulic conductivity at the site, pumping should rapidly capture a significant portion of the water within the contaminant plume. It is estimated that the pumping system would operate for approximately ten years, a balance between the effectiveness of the system and the expected system lifespan.

Because groundwater extraction is not proposed for source reduction, it may not appreciably reduce the predicted length of time required for RNA to complete groundwater remediation upgradient of the extraction system. If contaminant concentrations drop rapidly, the system could be shut off, but monitoring would have to

continue to see if concentrations rebound after pumping ceases. In addition, because groundwater extraction potentially generates a large volume of waste requiring treatment (via air stripping) and disposal, the alternative does not comply with the objectives of this AFCEE initiative.

As with Alternatives 1 and 2, institutional controls and LTM would be required. LTM wells would be the same as described for Alternative 1. LTM should continue for an additional 25 years after pumping ceases (a total of 35 years) to verify that concentrations do not rebound above MCLs and to establish plume behavior patterns created by pumping.

6.4 EVALUATION OF ALTERNATIVES

This section provides a comparative analysis of each of the remedial alternatives based on the effectiveness, implementability, and cost criteria. A summary of this evaluation is presented in Section 6.5.

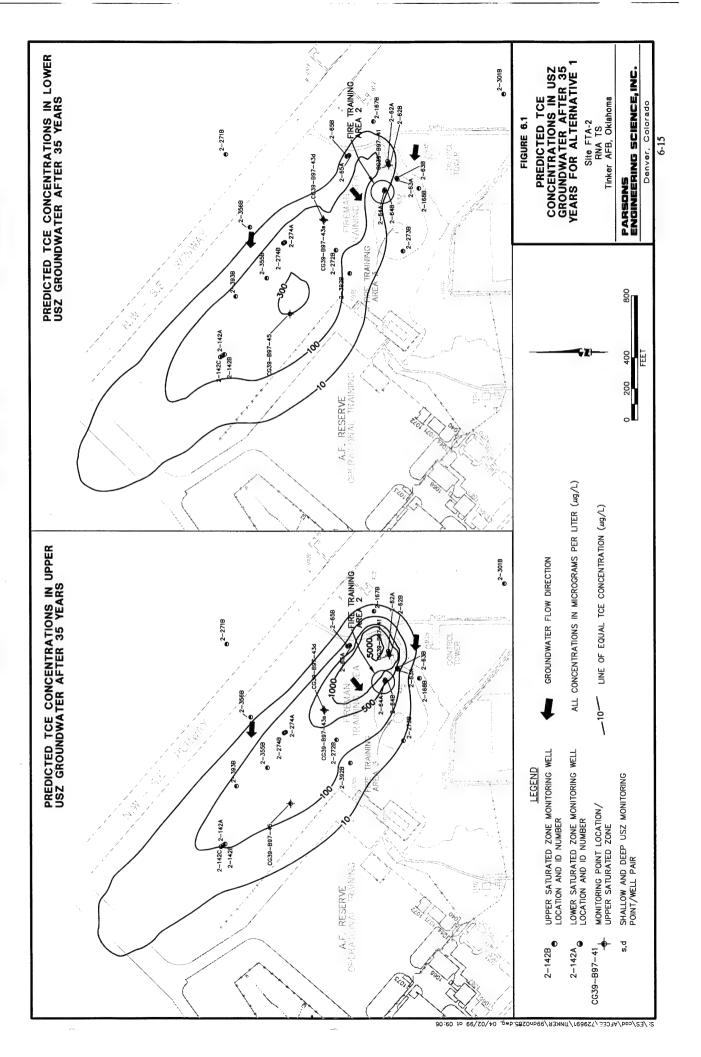
6.4.1 Alternative 1 - RNA and Institutional Controls with Long-Term Groundwater Monitoring

6.4.1.1 Effectiveness

Alternative 1 is based on the effectiveness of natural processes that minimize contaminant migration and reduce contaminant mass or concentrations over time as well as the effectiveness of institutional controls. To evaluate Alternative 1, the MODFLOW and MT3D numerical contaminant transport model presented in Section 5 was used to predict the future extent of TCE groundwater contamination. The Alternative 1 simulation assumes the TCE source from the calibrated model continues to contribute TCE to USZ groundwater for another 100 years (total simulation time of 135 years from initial source release). The TCE source is modeled with a conservative source weathering rate of two percent per year, due to processes such as volatilization, dissolution, and biodegradation (Section 5.4.4.1). Thus, the model assumes that after 1997, the source loading rates decreased geometrically by two percent per year (each year's source strength was decreased by two percent from the previous year's strength).

The predicted extent of TCE in the upper and lower USZ sand intervals for Alternative 1 after 35 years (year 2033) is shown on Figure 6.1. A 35-year period was assumed for the TCE plume calibration in Section 5, so a 35-year time period is also used to predict plume behavior in the future. The predicted extent of TCE contamination after 35 years is also used as the basis for the long-term monitoring network and sentry well locations discussed in Section 7.

The model predicts maximum TCE concentrations of up to $10,500~\mu g/l$ in the upper USZ sand interval in the vicinity of well 2-62 after 35 years, and it predicts the plume will extend northwest of well 2-142 approximately 1,000 feet. For the lower USZ sand interval, the maximum predicted TCE concentration is approximately 350 $\mu g/l$ in the vicinity of well 2-274B. The predicted plume shape and extent toward the northwest are similar for the upper and lower USZ sand intervals. Site data suggest that natural attenuation mechanisms are limited to the primary FTA-2 source area where anaerobic conditions are present.



TCE concentrations should exceed MCLs at the recommended LTM wells (Figure 7.1, Section 7). Speculative model predictions after 100 years indicate that the TCE plume may migrate approximately 1,400 feet northwest of well 2-142, with maximum TCE concentrations persisting in excess of 5,600 µg/L. Selection of additional sentry well locations should be made at a later date, when more data on plume behavior are available. Groundwater monitoring at the LTM wells will allow for continued evaluation of contaminant migration and ensure the safety of this alternative. While this alternative would not cease to be protective if the contaminant concentrations exceeding MCLs were detected in the LTM wells, such an instance would indicate that site conditions should be reevaluated.

The effectiveness of this remedial alternative requires that only properly protected site workers conduct future intrusive site activities or construction activities within the plume area. Reasonable land use assumptions for the plume area indicate that exposure is unlikely unless excavation or drilling activities bring groundwater or saturated soil to the surface. Long-term land use restrictions would be required to ensure that shallow groundwater is not pumped or removed for potable use within about 500 feet from the margins of the existing contaminant plume. Existing health and safety plans should be enforced to reduce worker exposures during additional excavation or installing and monitoring additional wells.

Compliance with program goals is one component of the long-term effectiveness evaluation criterion. Alternative 1 would satisfy program objectives designed to promote RNA as a component of site remediation and to scientifically document natural processes. In addition, the alternative satisfies waste minimization goals, as only limited drill cuttings would be generated during construction of new monitoring wells. This alternative also satisfies the program goal for cost effectiveness.

Apart from the administrative concerns associated with the enforcement of long-term land use restrictions and long-term groundwater monitoring programs, this remedial alternative should provide reliable, continued protection. It is assumed that dissolved contaminant concentrations will exceed state and federal criteria throughout the plume for at least 35 years under Alternative 1. Furthermore, it is assumed that sampling will be performed every year for 15 years, and biennially for the remaining 20 years, to demonstrate that RNA is reducing dissolved contaminant concentrations and limiting plume migration.

6.4.1.2 Implementability

Alternative 1 is not technically difficult to implement. Installation of any additional LTM wells and monitoring of groundwater are standard procedures. Long-term management efforts would be required to ensure proper sampling procedures are followed. Periodic site reviews should be conducted to confirm the adequacy and completeness of LTM data and verify the effectiveness of this remediation approach. There also may be administrative concerns associated with long-term enforcement of groundwater use restrictions. Leaving contaminated soil and groundwater in place may impact future land use within the source area. However, with the exception of any subsurface work at the site, the risk for Base personnel of exposure to contaminants will be limited. If required, the public and the regulators would have to be informed of

the benefits and limitations of the RNA option. Educational programs are not difficult to implement. Where the effectiveness of this option has been supported, the initial regulatory reaction to this alternative has been positive.

6.4.1.3 Cost

The cost of Alternative 1 is summarized in Table 6.3, and cost calculations are included in Appendix D. Capital costs are limited to the construction of 8 new LTM wells (Section 7). Included in the \$521,200 total present worth cost estimate for Alternative 1 are the costs of maintaining institutional controls and long-term monitoring at 30 LTM groundwater wells and 2 surface water stations for a total of 35 years. If the dissolved contaminant concentrations at the site decrease rapidly or drop below MCLs for consecutive sampling events, then monitoring may be reduced or eliminated. Conversely, significant increases for consecutive sampling events or a significant increase in plume extent could warrant an increase in sampling frequency or implementation of another remedial alternative.

TABLE 6.3 ALTERNATIVE 1 - COST ESTIMATE SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA

Capital Costs	Present Worth Cost
Design/Construct 8 LTM Wells	\$38,840
Monitoring Costs (per Sampling Event)	Present Worth Cost
Conduct Annual Monitoring of 30 Groundwater Wells and 2 Surface Water Stations (1999 to 2013)	\$336,200
Conduct Biennial Monitoring of 30 Groundwater Wells and 2 Surface Water Stations (2013 to 2033)	\$68,470
Site Management (Maintain Institutional Controls/Public Education) and Reporting (35 years)	\$77,690
Total Present Worth of Alternative 1 a/	\$521,200

^{a/} Based on an annual adjustment (discount) factor of 7 percent (USEPA, 1993).

6.4.2 Alternative 2 – Source Characterization and Removal, RNA, and Institutional Controls with Long-Term Groundwater Monitoring

6.4.2.1 Effectiveness

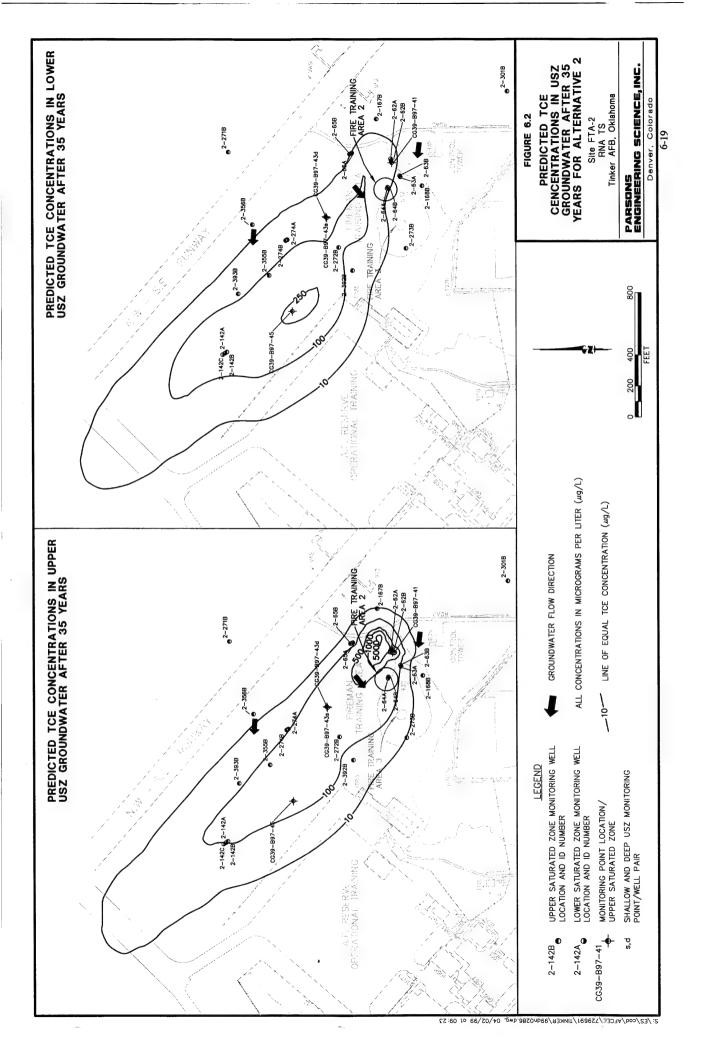
Alternative 2 includes additional source and site characterization activities including a soil gas survey, a geophysical survey, slug tests, and tracer tests, followed by source removal and LTM. Alternative 2 is based on locating and removing the source of TCE contamination as well as natural processes that minimize contaminant migration and reduce contaminant mass over time. To evaluate Alternative 2, the MODFLOW and MT3D numerical contaminant transport model presented in Section 5 was used to predict the future extent of TCE groundwater contamination after source removal. The Alternative 2 simulation assumes the TCE source from the calibrated model is reduced by 50 percent for 2 years, then is reduced by 2 percent per year as in Alternative 1.

The predicted extent of TCE in the upper and lower USZ sand intervals for Alternative 2 after 35 years (year 2033) is shown on Figure 6.2. The model predicts maximum TCE concentrations of up to 2,500 $\mu g/l$ in the upper USZ sand interval in the vicinity of well 2-62 after 35 years, and it predicts the plume will extend northwest of well 2-142 approximately 1,000 feet. For the lower USZ sand interval, the maximum predicted TCE concentration is approximately 250 $\mu g/l$ in the vicinity of well 2-274B. The predicted plume shape and extent toward the northwest are similar for the upper and lower USZ sand intervals. The predicted extent of TCE contamination after 35 years for Alternative 2 is also similar to the extent of TCE contamination after 35 years for Alternative 1. However, the maximum TCE concentrations in upper and lower USZ groundwater are significantly less for Alternative 2, particularly in the upper USZ sand where maximum TCE concentrations after 35 years are reduced approximately four-fold from Alternative 1.

Further delineation of TCE sources and hot spots at FTA-2 can be accomplished by implementing a soil gas survey of the upper USZ sand interval. Alternative 2 includes an initial soil gas survey on a relatively broad grid spacing of 50 feet, followed by a more focused investigation of areas identified as hot-spots on a tighter grid spacing of 20 ft. Results of the soil gas survey will be used to identify elevated concentrations of TCE in soil and upper USZ groundwater. Source removal by excavation will then be conducted based on results of the soil gas survey.

Two investigative techniques are recommended to further delineate anisotropic hydraulic conditions at FTA-2: an electromagnetic geophysical survey and a divergent radial flow, natural gradient tracer test. Electromagnetic (frequency domain) geophysical methods are recommended to further delineate paleochannels and/or zones of high hydraulic conductivity in the upper USZ sand interval. Geophysical methods can also be used to delineate the intra-USZ clay layer which separates the upper and lower USZ sand intervals and controls TCE groundwater transport from the upper to the lower USZ.

A conservative tracer test is proposed to evaluate contaminant transport rates and directions. For a radial flow, natural gradient tracer test, reagent grade lithium



bromide is instantaneously injected into the source area well. With time, the bromide should be detected in downgradient wells at varying concentrations. The relative concentrations of bromide in downgradient wells can be used to define groundwater velocities, dispersivity, and contaminant transport rates and directions. Three new monitoring wells should be installed in conjunction with the tracer test for the determination of downgradient bromide concentrations.

Alternative 2 should provide reliable, continuous protection. This alternative complies with AFCEE program goals, because RNA remains the predominant remediation method for contaminants dissolved in groundwater at the site. This remedial alternative, however, will result in the generation of additional contaminated soil that may require treatment and/or disposal. The effectiveness of RNA and institutional controls with LTM is discussed for Alternative 1 in Section 6.4.1.1.

6.4.2.2 Implementability

Alternative 2 is not technically difficult to implement. Recommended site characterization activities as well as installation of additional LTM wells and monitoring of groundwater are standard procedures. The technical and administrative implementability concerns associated with the natural attenuation and LTM components of this remedial alternative are similar to those discussed for Alternative 1.

6.4.2.3 Cost

The estimated capital and operating costs of Alternative 2 are shown in Table 6.4. The total present worth cost of Alternative 2 is \$1,058,000. The cost of Alternative 2 is increased from the costs of Alternative 1 by the addition of the soil gas survey, geophysical surveys, slug tests, tracer test, and source excavation. LTM is assumed to occur every year for 15 years, and biennially for an additional 20 years, to ensure that natural attenuation is reducing contaminant concentrations. The capital expense and annual costs for LTM and institutional controls are assumed to be the same as for Alternative 1.

6.4.3 Alternative 3 - Groundwater Extraction and Treatment, RNA, and Institutional Controls with Long-Term Groundwater Monitoring

6.4.3.1 Effectiveness

Alternative 3 includes groundwater extraction in addition to the source characterization and removal activities described for Alternative 2 and the long-term monitoring described for Alternative 1. Groundwater extraction is an established technology for controlling plume migration. The extraction of contaminated groundwater will prevent plume migration away from FTA-2, thereby preventing any completion of potential exposure pathways. A groundwater extraction system will also lower TCE concentrations in upper and lower USZ groundwater at FTA-2.

TABLE 6.4 ALTERNATIVE 2 - COST ESTIMATE SITE FTA-2 RNA TS

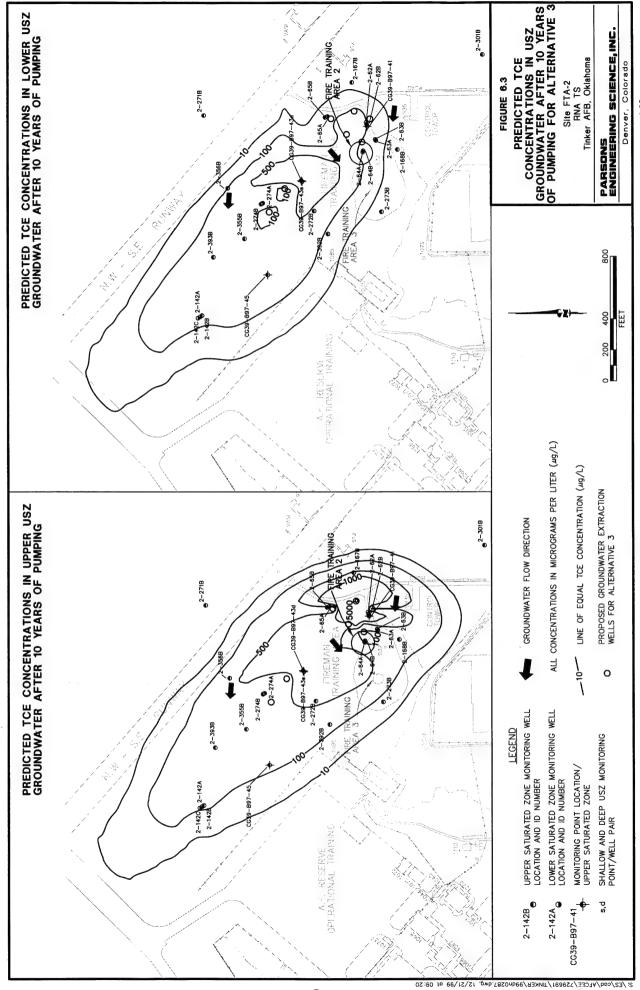
TINKER AFB, OKLAHOMA

Capital Costs	Present Worth Cost
Perform Initial Soil Gas Investigation	\$129,900
Perform Hot Spot Soil Gas Investigation	\$78,000
Perform Geophysical Survey	\$15,400
Perform Slug Tests	\$1,980
Perform Tracer Test	\$58,240
Soil Excavation and Disposal	\$253,100
Present Worth of Alternative 1	\$521,200
Subtotal for Present Worth of Alternative 2	\$536,600
Total Present Worth of Alternative 2 a/	\$1,058,000

^a/ Based on an annual adjustment (discount) factor of 7 percent (USEPA, 1993).

The MODFLOW and MT3D numerical contaminant transport model presented in Section 5 was used to estimate the number and locations of pumping wells in addition to the pumping rates needed to effect plume capture. The Alternative 3 simulations assume the TCE source is removed as in Alternative 2.

Containment of dissolved TCE concentrations greater than 1,000 μ g/l in the upper USZ is obtained after ten years of pumping five gpm from each of seven groundwater extraction wells (Figure 6.3). Five extraction wells would be completed in the upper USZ sand interval near source area monitoring well 2-62B, and two extraction wells would be completed in the lower USZ sand interval near monitoring well 2-274B. In the lower USZ, containment of dissolved TCE concentrations greater than 500 μ g/l is obtained after ten years of pumping (Figure 6.3).



Alternative 3 should provide reliable, continuous protection for downgradient receptors. This alternative, however, does not comply well with all of the AFCEE program goals because of the generation of soil and water waste requiring treatment and disposal. In addition, contaminants are not destroyed, but are instead transferred to another phase that may require further treatment. As with Alternatives 1 and 2, this alternative would rely on RNA with LTM and institutional controls to remediate the contaminated groundwater once the extraction system is shut down.

6.4.3.2 Implementability

Groundwater extraction would be labor intensive and expensive to implement. The option would require additional site investigation, design and engineering, installation, and a weekly commitment to operation and maintenance of the extraction and air stripper systems. An air emissions permit may need to be obtained for the air stripper prior to system start-up. A permit also may be needed for disposal of treated groundwater. The technical and administrative implementability concerns associated with the RNA, LTM, and institutional controls of this remedial alternative are similar to those discussed for Alternatives 1 and 2.

6.4.3.3 Cost

The estimated capital and operating costs of Alternative 3 are shown in Table 6.5. The total present-worth cost of Alternative 3 is \$2,156,000. Despite the shorter operating time frame, the cost of Alternative 3 is comparable to the costs of Alternative 2 because it is more costly to design, install, operate, and maintain the groundwater extraction system. However, the cost could be significantly higher if off-gas treatment for the air stripper is required, if the system has to run for more than 10 years, or if treated groundwater cannot be easily disposed (e.g., discharged to a storm sewer). The annual costs for LTM and institutional controls are assumed to be the same as for Alternatives 1 and/or 2. The resulting present-worth cost for LTM and institutional controls are the same as for Alternative 1 because it is assumed that the groundwater extraction system merely contains the core of the contaminant plume without treating the source, and therefore, will operate for the same length of time.

6.5 RECOMMENDED REMEDIAL APPROACH

Three remedial alternatives have been evaluated for remediation of USZ groundwater at FTA-2. Components of the alternatives evaluated include RNA with LTM of groundwater, institutional controls, source characterization and removal, and groundwater extraction with ex-situ treatment. Table 6.6 summarizes the results of the evaluation based upon effectiveness, implementability, and cost criteria. Due to the risk of vertical contaminant migration and to the high cost and risk of engineered Alternative 3, the Air Force recommends Alternative 2 as a balanced and most cost-effective option for risk reduction at the study area.

All three alternatives make use of natural attenuation mechanisms to reduce plume migration and toxicity, although Alternative 3 relies more on capture of the plume with the extraction system. Alternatives 2 and 3 would use active *in situ* and *ex situ* techniques to limit contaminant migration. Implementation of Alternatives 2 and 3

TABLE 6.5 ALTERNATIVE 3 - COST ESTIMATE SITE FTA-2 RNA TS TINKER AFB. OKLAHOMA

Capital Costs	Present Worth Cost
Install Groundwater Extraction System with Air Stripper	\$708,000
Operation, Maintenance, and Monitoring Costs	Present Worth Cost
Operation and Maintenance for the Pump and Treat System (10 years)	\$389,950
Subtotal Present Worth of Alternative 1	\$521,200
Subtotal Present Worth of Alternative 2	\$536,600
Subtotal Present Worth of Alternative 3	\$1,098,000
Total Present Worth of Alternative 3 a/	\$2,156,000

^{a/} Based on an annual adjustment (discount) factor of 7 percent (USEPA, 1993).

would not substantially decrease the time frame for remediation, but both alternatives would require greater capital expenditures. Alternatives 2 and 3 are considered less favorable because in part they simply transfer contamination to a different medium rather than reducing contaminants to innocuous byproducts.

All three remedial alternatives are implementable and effectively reduce potential contaminant migration and toxicity in the groundwater. Under all three remedial alternatives, the plume is predicted to expand, but without impact to potential receptors (Figures 6.1, 6.2, and 6.3). All three alternatives should be acceptable to the public and regulatory agencies because they are protective of human health and the environment and reduce groundwater contamination; however, only Alternative 2 addresses soil (source) contamination. Implementation of any of the three alternatives will require land and groundwater use controls to be enforced. Groundwater monitoring would be required for the respective projected cleanup periods.

The 35-year remediation time for Alternative 1 is not considered conservative because the plume is likely expanding. Uncertainties about the nature and location of contaminant sources mean that site conditions could change during the LTM period and that additional contaminant mass could unexpectedly enter groundwater. However, this potential drawback would apply to all alternatives and would require reevaluation in all

TABLE 6.6 SUMMARY OF REMEDIAL ALTERNATIVES EVALUATION FTA-2 RNA TS TINKER AFB, OKLAHOMA

Present Worth Cost Estimate	\$521,200	\$1,058,000	\$2,156,000
Implementability	Readily implementable. Groundwater water quality monitoring required for an estimated minimum of 35 years, and potentially much longer. Institutional controls, including land and groundwater use controls, may be required for more than 100 years due to the persistance of elevated dissolved CAH concentrations. Minimal exposure to potential receptors if institutional controls are implemented. If significant plume migration occurs and completion of exposure pathways is possible, then additional remedial work may be required.	Source reduction simulated by the numerical model should not pose significant implementability concerns. However, longterm site management, groundwater use controls, and monitoring would be required as elevated TCE concentrations may persist for more than 100 years. If significant plume migration occurs and completion of exposure pathways is possible, then additional remedial work may be required.	Installation/operation of the pump and treat system simulated by the numerical model should not pose significant implementability concerns. However, long-term site management, groundwater use controls, and monitoring would be required as elevated CAH concentrations may persist for more than 50 years. If significant plume migration occurs and completion of exposure pathways is possible, then additional remedial work may be required.
Effectiveness	Contaminant mass, volume, and toxicity will be slowly reduced over time, but the dissolved plume may persist for well over 100 years, necessitating a potentially lengthy period of LTM and institutional controls.	Similar to Alternative 1, with the addition of source characterization, source removal by excavation, and additional site characterization by tracer test and geophysical survey. Contaminant mass, volume, and toxicity will be reduced more rapidly than in Alternative 1. However, even if the source is substantially reduced, significantly elevated dissolved TCE concentrations may still persist for more than 50 years, and downgradient plume migration may exceed 750 feet, depending on the actual biodegradation rate.	Similar to Alternatives 1 and 2, with the addition of a multi-well pump and treat system to reduce the highest dissolved TCE concentrations. Contaminant mass, volume, and toxicity will be reduced more rapidly than in Alternative 1 and 2. However, even if the source and dissolved TCE concentrations are substantially reduced, significantly elevated dissolved TCE concentrations may still persist for more than 50 years, and downgradient plume migration may exceed 500 feet, depending on the actual biodegradation rate.
Remedial Alternative	Alternative 1 - Natural Attenuation - Long-Term Monitoring - Institutional Controls	Alternative 2 - Source Characterization - Source Removal - Tracer Test - Geophysical Survey - Natural Attenuation - Long-Term Monitoring - Institutional Controls	Alternative 3 - Groundwater Pump and Treat - Source Characterization - Source Removal - Tracer Test - Geophysical Survey - Natural Attenuation - Long-Term Monitoring - Institutional Controls

cases. The proposed LTM period is not consistent with federal recommendations that proposed or implemented remedial activities at a site should not exceed 30 years in duration (USEPA, 1988), and even if site conditions require a change in the remedial strategy it is unlikely that this time limit would be met. The final evaluation criterion used to compare each of the remedial alternatives was cost; the additional costs of Alternative 3 does not justify the slightly reduced risk resulting from plume containment.

Alternative 2 will reduce the level of contamination and maintain the necessary degree of protection to potential receptors at or downgradient from the site, and is the recommended remedial alternative for USZ groundwater at FTA-2. Alternative 2 combines natural attenuation with source reduction, and provides for additional site characterization necessary to accurately predict the future behavior of dissolved contaminants at FTA-2. Future exposure to potential receptors at the site will be minimized by land use restrictions, because access to the Base (and hence the site) is controlled by Base security. A LTM plan for groundwater, including a generic SAP, is provided in Section 7.

SECTION 7

LONG-TERM MONITORING PLAN

7.1 OVERVIEW

In keeping with the requirements of the preferred remedial alternative for Tinker AFB Site FTA-2, a LTM plan was developed. The long-term groundwater and surface water monitoring plan for FTA-2 presented in this section describes a monitoring scheme for the next 35 years for planning and budgeting purposes. It should be noted that this plan was devised on the basis of currently available data; the LTM program may be progressively revised as new data are obtained during this 35-year period. The LTM plan consists of identifying the locations of LTM wells and surface water stations, and developing a groundwater and surface water sampling and analysis strategy to accomplish the following objectives:

- Monitor changes in site conditions, including plume magnitude and extent, over time;
- Assess the effectiveness of engineered remedial actions and naturally occurring processes at reducing contaminant mass and minimizing contaminant migration;
- Assess the degree to which site-specific remediation goals (Section 6.1) are being attained and facilitate the evaluation of the need for additional remediation; and
- Verify the predictions of the numerical contaminant fate and transport model.

The strategy described in this section is designed to monitor plume migration over time and to verify that natural and engineered remedial processes are adequately protecting potential receptors. Based on the presence of an upward vertical groundwater gradient at well location 2-62B, surface water sampling in the ditch to the south of well 2-62B (Figure 4.3) also is recommended to determine if the TCE plume is discharging to surface water. In the event that data collected under this LTM program indicate planned combination of naturally occurring processes and engineered remedial actions is insufficient to protect human health and the environment, contingency controls to more aggressively remediate the dissolved TCE plume would be necessary.

7.2 GROUNDWATER AND SURFACE WATER MONITORING NETWORK

LTM of a minimum of 18 existing USZ, 4 existing LSZ, and 8 proposed monitoring wells located upgradient from, within, and on the periphery of the primary TCE plume is recommended (30 total wells). Existing USZ wells include CG39-B97-41, CG39-B97-43s, CG39-B97-43d, CG39-B97-45, 2-62B, 2-63B, 2-64B, 2-65B, 2-142B, 2-167B, 2-

168B, 2-272B, 2-273B, 2-274B, 2-355B, 2-356B, 2-392B, and 2-393B (Figure 7.1). TCE was detected in LSZ monitoring well 2-62A during the August 1997 sampling event. Therefore, wells 2-62A, 2-63A, 2-64A, and 2-274 are recommended for monitoring in the LSZ. These wells are located beneath the USZ TCE plume axis (2-62A, 2-64A, and 2-274A), and downgradient of 2-62A (2-63A). In addition, eight proposed monitoring wells and two surface water stations also would be sampled as a part of the LTM for Site FTA 2.

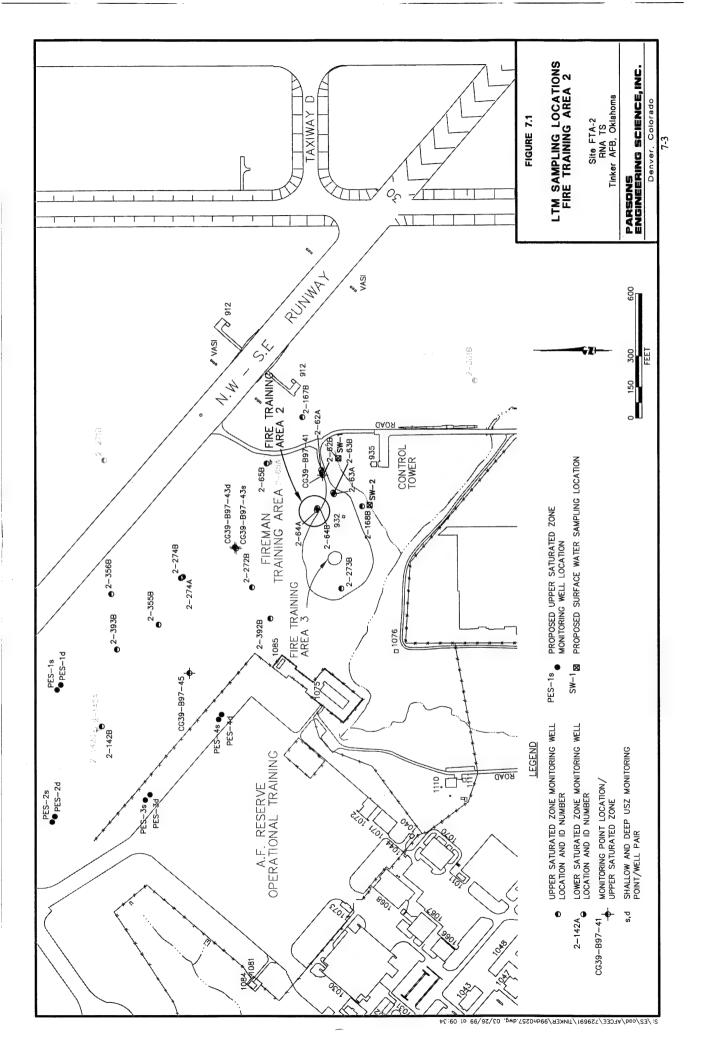
On the basis of the contaminant transport modeling results presented in Section 6, the TCE plume may migrate more than 1,000 feet past the current downgradient plume boundary over the next 35 years. Because the groundwater flow direction and subsurface properties in the area downgradient from the current plume boundary have not been investigated, installation of additional LTM wells approximately 300-500 feet downgradient of the current plume boundary is appropriate. Progressive tracking of the plume in the downgradient direction over time is recommended. In this way, the downgradient migration of the plume can be monitored. It is anticipated that additional sentry wells will be necessary to monitor the maximum plume extent before the 35-year monitoring period expires. The number and locations of these sentry wells cannot be accurately predicted at this time based on the limited hydrogeologic data for this site. Therefore these wells are not included or costed for this LTM plan.

The TCE plume in both the upper and lower USZ sands is not well defined to the north and west (Figure 4.3). The TCE plume in the upper USZ sand is only defined by non-detection in wells 2-356B (upgradient), 2-142B (downgradient), and 2-273B (downgradient). The TCE plume in the lower USZ sand is only defined by non-detection in wells 2-271B (upgradient), 2-142B (downgradient), and 2-392B (downgradient). Therefore, four well pairs screened in the USZ are recommended to monitor future plume expansion. The USZ well locations were selected to monitor USZ groundwater quality both horizontally and vertically in directions in front of the primary contaminant plume axis, and hydraulically downgradient of plume. Suggested locations for these wells are shown on Figure 7.1. If CAH contamination is detected in these wells in the future, then additional wells could be installed further downgradient. Estimates of the groundwater/CAH migration rate and direction should be progressively refined as new wells are installed to aid in locating additional LTM wells.

Prior to installation of the four LTM well pairs north and west of the current plume boundary, the stratigraphy at the well locations should be investigated to ensure that the wells are appropriately screened to intercept transmissive zones that may act as preferred contaminant migration pathways. This information could be obtained using a conventional hollow-stem auger (HSA) drilling rig and running borehole geophysical logs.

7.3 ANALYTICAL PROTOCOL

All LTM wells should be sampled and analyzed to determine compliance with chemical-specific remediation goals and to verify the effectiveness of remediation at the site. Groundwater level measurements should be made during each sampling event. Because natural attenuation is a component of the remedial strategy, appropriate geochemical parameters should also be analyzed in addition to target analytes.



Groundwater samples from LTM wells should be analyzed for the parameters listed in Table 7.1.

7.4 SAMPLING FREQUENCY

Under the current remedial strategy, dissolved TCE concentrations in excess of the 5 μ g/L standard will be present for more than 35 years. Estimated LTM costs for a 35-year period beginning in 1999 are presented in Section 7.6. The 30 LTM wells (18 existing USZ, 4 existing LSZ, and 8 proposed) and 2 surface water stations would be sampled annually for 15 years beginning in 1999 (the duration of the engineered remedial actions evaluated), and every second year for the remaining 20 years of the 35-year period.

The LTM plan should be periodically reviewed and revised as appropriate on the basis of available groundwater quality data and the collection of additional source and hydrogeologic characterization data. For example, if the data collected during this time period indicate the plume has stabilized or is receding, and that CAH concentrations are diminishing, then the sampling frequency can be reduced. If sampling results indicate that geochemical conditions in the plume area are stable over time (e.g., nitrate, sulfate, and ferrous iron concentrations), then the sampling frequency for these parameters could be reduced. If the data collected at any time during the monitoring period indicate the need for additional remedial activities at the site, sampling frequency should be adjusted accordingly.

7.5 LTM COST ESTIMATE

The estimated present worth cost for the LTM program described in the preceding sections is shown in Table 7.2. Included in the total present worth cost of \$502,900 are the estimated costs for installing four additional LTM wells, performing the recommended groundwater and surface water monitoring, maintaining institutional controls, public education, project management, and reporting.

TABLE 7.1 LONG-TERM MONITORING ANALYTICAL PROTOCOL FOR GROUNDWATER FTA-2 RNA TS TINKER AFB, OKLAHOMA

				Kecommended	Sample Volume, Sample	Field or
				Frequency of	Container, Sample Preservation	Fixed-Base
Analyte	Method/Reference	Comments	Data Use	Analysis		Laboratory
Temperature	E170.1, direct-reading	Field only, measure	Metabolism rates for	Each sampling	Measure at well-head using a	Field
	meter	at well-head.	microorganisms depend on temperature.	event	flow-through cell.	
Dissolved	Dissolved oxygen	Measure at well-	Concentrations less than	Each sampling	Collect water during purging in a	Field
Oxygen	meter	head, Refer to	I mg/L generally indicate an	event	flow-through cell or plastic	
		Method A4500	anaerobic pathway.		container; analyze immediately.	
		for a comparable laboratory procedure				
Hd	E150.1/SW9040, direct	Measure at well-	Aerobic and anaerobic	Each sampling	Collect 100-250 mL of water in a	Field
	reading meter	head. Protocols/	processes are pH-sensitive.	event	glass or plastic container; analyze	
		Handbook methods2			immediately, or measure at well-	
					head using a flow-through cell.	
Conductivity	E120.1/SW9050, direct	Protocols/Handbook	General water quality	Each sampling	Collect 100-250 mL of water in a	Field
	reading meter	methods	parameter used as a marker to	event	glass or plastic container and	
			verify that site samples are		analyze immediately, or measure	
			obtained from the same		at well-head using a flow-through	
			groundwater system.		cell.	
Redox potential	A2580 B, direct	Measurements	The redox potential of	Each sampling	Collect 100-250 mL of water in a	Field
	reading meter.	are made with	groundwater influences and is	event	glass container, filling container	
		electrodes using a	influenced by biologically		from bottom; analyze immediately	
		meter; samples	mediated reactions; the			
		should be protected	groundwater redox potential			
		from exposure to	may range from 200 mV to			
		atmospheric oxygen	less than -400 mV.			
Ferrous Iron	Colorimetric	Field only, filter if	Elevated ferrous iron	Each sampling	Collect 100 mL of water in a glass	Field
(Fe ²⁺)	A3500-Fe D or	turbid.	concentrations may be	event	container; acidify with	
	Colorimetric		indicative of the anaerobic		hydrochloric acid per method	
	Hach [®] 25140-25		biodegradation process of iron			
			reduction.			

TABLE 7.1 (Concluded) LONG-TERM MONITORING ANALYTICAL PROTOCOL FOR GROUNDWATER TINKER AFB, OKLAHOMA FTA-2 RNA TS

				Recommended Frequency of	Sample Volume, Sample	Field or Fixed-Base
Analyte	Method/Reference	Comments	Data Use	Analysis	Preservation	Laboratory
Nitrate	IC method E300 or method SW9056;	Method E300 is a Handbook method.	Substrate for microbial respiration if oxygen is	Each sampling event	Collect up to 40 mL of water in a glass or plastic container, cool	Fixed-base
	colorimetric method E353.2.	Method SW9056 is an equivalent procedure.	depleted.		10 4°C	
Sulfate (SO ₄ ²⁻)	IC method E300 or method SW9056 or	Method E300 is a photometric	Substrate for anaerobic microbial respiration	Each sampling event	Collect up to 40 mL of water in a glass or plastic container; cool	Fixed-base or field (for
	Hach [®] SulfaVer 4 method	Handbook method; method SW9056 is			to 4°C	Hach® method)
		an equivalent procedure.				
Methane, Ethane and	RSKSOP-114 modified to analyze water	Method published	The presence of methane indicates the presence of	Each sampling event	Collect water samples in 40 mL volatile organic analysis (VOA)	Fixed-base
Ethene	samples for methane by headspace sampling	USEPA Robert S. Kerr Laboratory	sufficiently reducing conditons for reductive dehalogenation to		vials with butyl gray/Teflon- lined caps (zero headspace);	
	with dual thermal conductivity and flame ionization detection.		occur		cool to 4°C	
Volatile	GC/MS method 8260B	Handbook method	Measured for regulatory	Each sampling	Collect water samples in a 40	Fixed-base
Organics			compliance	event	mL VOA vial; cool to 4°C; add hydrochloric acid to pH < 2	

a/ Protocol analytical methods are those presented by Wiedemeier et al. (1995). Handbook refers to "AFCEE Handbook to Support the Installation Restoration Program (IRP) Remedial Investigation/Feasibility Study (RI/FS).

TABLE 7.2 ESTIMATED LTM COSTS FTA-2 RNA TS TINKER AFB, OKLAHOMA

Capital Costs	Present Worth Cost
Design/construct eight LTM wells in 1999	\$38,840
Monitoring Costs	Present Worth Cost
Conduct Annual Groundwater Monitoring of 30 Wells and Surface Water Monitoring of 2 Stations from 1999 to 2013.	\$321,000
Conduct Biennial Groundwater Monitoring of 30 Wells and Surface Water Monitoring of 2 Stations from 2013 to 2033.	\$65,370
Site Management (Maintain Institutional Controls/Public Education) and Reporting (35 years)	\$77,690
Total Present Worth of LTM Program a/	\$502,900

a/ Based on an annual inflation (discount) factor of 7 percent (USEPA, 1993).

Note: Cost assume that well installation and LTM are performed by local (Oklahoma City area) personnel.

SECTION 8

CONCLUSIONS AND RECOMMENDATIONS

This report presents the results of a TS conducted to evaluate the use of natural attenuation for remediation of CAH-contaminated groundwater at FTA-2 at Tinker AFB, Oklahoma. The numerical models MODFLOW and MT3D were used in conjunction with site-specific geologic, hydrologic, and laboratory analytical data to simulate the migration and biodegradation of CAHs dissolved in groundwater. To obtain the data necessary for the RNA demonstration, soil and groundwater samples were collected from the site and analyzed. Physical and chemical data collected under this program were supplemented (where necessary) with data collected during previous site characterization events.

It is likely that contaminants were first introduced into groundwater at this site when the FTA-2 fire pit was initially used for fire training exercises in 1962. TT (1996) reported the FTA-2 pit was operational until 1966, although aerial photos indicate the pit was present as late as 1970. Fire training exercises at Air Force bases typically use contaminated or uncontaminated jet fuel. Solvents, fuels, and other liquid wastes may have been used for training exercises or cleaning, or dumped in the general fire training area. Aerial photographs show numerous dirt access roads crossing the area through the mid-1970s, indicating unknown activities continued in the area after use of the FTA-2 pit was discontinued.

Analytical data indicate that the highest concentrations of CAHs occur in the vicinity of well 2-62B, located approximately 100 to 120 feet upgradient from the former fire pit along an access road. The presence of a secondary source, located further upgradient from well 2-62B and the former fire pit along another access road near the flightline (well 2-65B), also is suggested by increasing concentrations of CAHs at that location (Appendix H).

Contaminants observed in groundwater at concentrations greater than USEPA MCLs at FTA-2 include benzene, PCE, TCE, 1,1-DCE, cis-1,2-DCE, VC, 1,2-DCA, CB, 1,2-DCB, and 1,4-DCB. Temporal data (Appendix H) for concentrations of TCE and cis-1,2-DCE from monitoring wells/points at FTA-2 indicates that concentrations in the potential source areas are stable or gradually increasing. Given the available data, it appears that the plume is expanding downgradient from the suspected primary FTA-2 source and also at an upgradient location, based on increased concentration data for wells 2-63B, 2-64B, and 2-65B. An increase in concentrations within the center of the CAH plume at wells 2-72B and 2-274B further indicate the presence of secondary source(s) and/or expansion of the CAH plume in that area. Conversely, decreasing CAH concentration data for wells

2-355B and 2-393B along the plume axis at the most northern portion of the plume indicate the plume may be either decreasing or approaching equilibrium in this area.

Comparison of CAH, BTEX, TOC, electron acceptor, and byproduct data provides evidence that CAHs dissolved in groundwater at FTA-2 are being naturally degraded in limited areas. The presence of daughter products cis-1,2-DCE and VC provides strong evidence that parent solvents (TCE and/or PCE) are being reductively dehalogenated in an area limited to the general source area(s). However, a relative lack of VC and ethene indicates that the process does not significantly proceed past the initial step of TCE to DCE. TCE, DCE and DCA are reduced to a significantly lesser extent downgradient of the source area(s). Anaerobic conditions in the source area may be driven by low concentrations of anthropogenic fuel hydrocarbons. Downgradient of the source area, reductive dehalogenation ceases, and compounds such as DCE, VC, and chlorobenzenes likely are aerobically degraded or diluted. While VC and chlorobenzenes are reduced to non-detectable levels downgradient of the source area, DCE persists with TCE to the leading edge of both the upper and lower USZ plumes. 1,2-DCA also is present in elevated concentrations downgradient of the source area. It is not apparent whether elevated concentrations of 1.2-DCA have migrated downgradient from the FTA-2 source area, or whether an additional source of 1,2-DCA exists in that area.

Molar fractions (Appendix H) of chlorinated ethenes along the apparent CAH plume axis indicates little, if any, degradation of TCE to DCE outside the immediate source area(s). At well 2-65B, an overall decrease in the molar fraction of TCE, with an increase in molar fraction of DCE and the presence of VC indicates that TCE is degrading to DCE, and DCE to VC at that location. Geochemical data indicate that biodegradation of native or anthropogenic carbon is occurring via aerobic respiration and methanogenesis in suspected source areas, but that nitrate, ferric iron, and sulfate reduction are not significant anaerobic degradation processes at the site. Furthermore, there are an abundance of alternate electron acceptors that may inhibit use of CAHs as electron acceptors. Away from the suspected source areas, DO concentrations and ORPs indicate that environmental conditions are not sufficiently reducing for dehalogenation of CAH compounds.

Site-specific hydrogeologic and laboratory analytical data were used in the numerical groundwater model to simulate the effects of advection, dispersion, sorption, and biodegradation on the fate and transport of the TCE plume. TCE is used as an indicator compound in this report due to its relatively high concentration, low MCL, and persistence downgradient of the source area. During model calibration, model parameters that could not be obtained from existing site data were estimated using common literature values for aquifer materials similar to those found at the site. Model predictions were found to be very sensitive to hydraulic conductivity, anisotropy, dispersivity, and the magnitude of the first-order TCE decay rate assigned to the upper and lower USZ sands at FTA-2.

Remedial Alternative 1 incorporates natural attenuation, but does not include engineered remediation. The model used to simulate this alternative incorporates the assumption that dissolution from residual NAPL in the source area into groundwater will naturally decrease at a geometric rate of 2 percent per year (each concentration was decreased by a factor equal to 2 percent of the previous year's concentration) during the

35-year predictive period. The results of the model indicate that TCE will continue to migrate over the next 35 years an additional 1,000 feet from the present plume toe in both the upper and lower USZ sand intervals. Plume migration is too speculative to predict beyond 35 years based on limited hydrogeologic characterization of the site. Furthermore, TCE is predicted to persist in the source area at concentrations greater than 10,500 µg/L over the same 35 year period.

The model used to simulate remedial Alternative 2 assumes that partial source removal via application of engineered remedial technologies such as such as soil vapor source characterization and excavation would reduce TCE source mass. This model incorporates the assumption that the remediation will reduce the TCE source into groundwater by 50 percent per year over a 2 year period, followed by a 2 percent per year reduction due to natural weathering for the remainder of the 35 year model simulation. Results of these models indicate that source reduction at the assumed rates would result in decreases in dissolved TCE concentrations within the plume relative to Alternative 1. The Alternative 2 model predicts that maximum TCE concentrations would be reduced approximately four-fold in the upper USZ sand interval, and by approximately two-fold in the lower USZ sand interval over the 35-year model simulation. However, the overall plume extent would not be reduced, and TCE concentrations would continue to persist above USEPA MCLs.

A third groundwater model was used to simulate the potential effects of a groundwater extraction and treatment system. The Alternative 3 model was identical to Alternative 2, with the addition of seven pumping wells placed in the currently delineated TCE plume at areas of maximum TCE concentrations. The results of this model indicate that groundwater extraction and treatment would lower TCE concentrations in upper and lower USZ groundwater. On the basis of model simulations, the extraction system will contain groundwater with dissolved TCE concentrations in excess of 1,000 μ g/L in the upper USZ after 10 years of pumping. In the lower USZ, the Alternative 3 model predicts containment of dissolved TCE concentrations in excess of 500 μ g/L after 10 years of pumping.

It should be noted that the numerical model constructed for this TS is reasonably conservative; therefore, plume migration and persistence may be less than that predicted by the model. Factors that could cause plume migration and persistence to differ from model predictions include the following:

- The calibrated model is not necessarily unique, and different combinations of input parameters could potentially have been used to achieve an acceptable calibration;
- The calibrated model is limited in accuracy by inadequate characterization of the source, preferential flowpaths, and aquifer anisotropy;
- The model does not account for the potential occurrence of abiotic degradation and volatilization of TCE; and
- Hydrogeologic and/or geochemical conditions that could slow or halt plume migration may exist downgradient from the investigated area.

This report presents a conceptual model where the primary source of TCE was near the FTA-2 fire pit and well 2-62B, and the TCE plume migrated toward the northwest due to a preferential flow path and anisotropy of the aquifer matrix. This conceptual model was primarily based on the plume configuration in 1997, which suggests linear contaminant transport in a northwest direction. Fate and transport modeling results indicate that dissolved TCE contamination present in groundwater north and west of the FTA-2 fire pit has the potential to migrate significantly. However, concentrations for wells located at the northern toe of the TCE plume (2-355B in the upper USZ sand interval and 2-393B in the lower USZ sand interval) decreased substantially from 1997 to 1999 (Appendix H).

Alternately, the TCE plume may be migrating in the direction of maximum hydraulic gradient (west-southwest) from multiple sources. The presence of multiple TCE "hotspots" does not support a conceptual model of a single source; rather, the "hotspots" suggest multiple sources. Increases in the concentration of TCE in 1999 at wells located downgradient of suspected TCE sources or "hotspots" (wells 2-64B and 2-272B) suggest a more westerly migration in the direction of maximum hydraulic conductivity. Therefore, the numerical model presented in this TS should be considered as only one of multiple possible interpretations.

Alternative 2 is the recommended remedial alternative for USZ groundwater at FTA-2 because it will reduce the level of contamination while maintaining the necessary degree of protection to potential receptors at or downgradient from the site. Alternative 2 combines natural attenuation with source reduction, and provides for additional site characterization necessary to accurately predict the future behavior of dissolved contaminants at FTA-2. Source charactrization activities may include soil gas surveys, geophysical surveys, additional aquifer testing, and tracer tests. Future exposure to potential receptors at the site can be minimized by land use restrictions, because access to the Base (and hence the site) is controlled by Base security. The beneficial affects of natural attenuation are included in Alternative 2. The degree to which RNA will continue to control significant downgradient migration of the plume is not known with certainty, but will become more apparent as successive LTM sampling events are performed.

To assess the effectiveness of both naturally-occurring processes and source removal at reducing contaminant mass and minimizing contaminant migration, groundwater from a minimum of 22 existing monitoring wells and 8 proposed monitoring wells should be sampled annually for 15 years to build a historical groundwater quality database for the site. Two surface water stations also should be sampled annually during this 15-year period. Following completion of the 15-year annual LTM period, the LTM program should be reassessed, and sampling frequency may be reduced to biennial events. The number and location of additional sentry wells and surface water stations selected for continued monitoring should be determined on the basis of results from the initial 15-year LTM period. Likewise, the model can be adjusted to reflect additional source and hydrogeologic characterization.

The LTM plan should be periodically reviewed and revised as appropriate on the basis of available groundwater quality data. For example, if sampling results demonstrate that the plume is in equilibrium (not migrating downgradient) or is receding toward the source area, then the sampling frequency could be reduced appropriately. If the data collected at any time during the monitoring period indicate the need for additional remedial activities

at the site, sampling frequency should be adjusted accordingly and the appropriate remedial actions (e.g., groundwater extraction and treatment as described in Section 6) should be evaluated and implemented.

SECTION 9

REFERENCES

- Abriola, L.M., 1996, Organic liquid contaminant entrapment and persistence in the subsurface: Interphase mass transfer limitation and implications for remediation. 1996 Darcy Lecture, National Ground Water Association, presented at Colorado School of Mines, October 25.
- Alvarez-Cohen, L.M. and McCarty, P.L., 1991a, Effects of toxicity, aeration, and reductant supply on trichloroethylene transformation by a mixed methanotrophic culture: Appl. Environ. Microbiol., vol. 57, no. 1, p. 228-235.
- Alvarez-Cohen, L.M., and McCarty, P.L., 1991b, Product toxicity and cometabolic competitive inhibition modeling of chloroform and trichloroethylene transformation by methanotrophic resting cells: Appl. Environ. Microbiol., vol. 57, no. 4, p. 1031-1037.
- Anderson, M.P., and W.W. Woessner. 1992. Applied Groundwater Modeling Simulation of Flow and Advective Transport: Academic Press, New York, 381 p.
- Arciero, D., Vannelli, T., Logan, M., and Hooper, A.B., 1989, Degradation of trichloroethylene by the ammonia-oxidizing bacterium *Nitrosomonas europaea*: Biochem. Biophys. Res. Commun., vol. 159, p. 640-643.
- Ashworth, R.A., G.B. Howe, M.E. Mullins, and T.N. Rogers. 1988. Air-water partitioning coefficients of organics in dilute aqueous solutions. *Journal of Hazardous Materials*, 18:25-36.
- Ball, W.P., and Roberts, P.V., 1991, Long-term Sorption of Halogenated Organic Chemicals by Aquifer Material: 1. Equilibrium: Environmental Science and Technology, vol. 25, p. 1223-1236.
- Bear, J., 1979, Hydraulics of Groundwater: McGraw-Hill, Inc., New York, New York, 569.
- Bingham, R.H. and R.L. Moore, 1975, Reconnaissance of the Water Resources of the Oklahoma City Quadrangle, Central Oklahoma, Oklahoma Geological Survey, Hydrologic Atlas 4.
- Bouwer, E.J., Rittman, B.E., and McCarty, P.L., 1981, Anaerobic degradation of halogenated 1- and 2-carbon organic compounds: Environ. Sci. Technol., vol. 15, no. 5, p. 596-599.

- Bouwer, E.J. and Wright, J.P., 1988, Transformations of trace halogenated aliphatics in anoxic biofilm columns: J. Contam. Hydrol., vol. 2, p. 155-169.
- Bouwer, E.J., 1992, Bioremediation of Subsurface Contaminants, In: Mitchell, R., ed.: Environmental Microbiology. Wiley-Liss, New York, New York.
- Bouwer, E.J., 1994, Bioremediation of chlorinated solvents using alternate electron acceptors. In: Handbook of Bioremediation. CRC Press, Boca Raton, FL.
- Bouwer, H. 1989. The Bouwer and Rice slug test an update. *Ground Water*, 27(3):304-309.
- Bouwer, H. and R.C. Rice. 1976. A slug test for determining hydraulic conductivity of unconfined aquifers with completely or partially penetrating wells. *Water Resources Research*, 12(3):423-428.
- Bradley, P.M., and Chapelle, F.H., 1996, Anaerobic mineralization of vinyl chloride in Fe(III)-reducing aquifer sediments: Accepted for publication in Environmental Science and Technology, 1996.
- Butler, B.J., and J.F. Barker. 1996. Chemical and microbiological transformation and degradation of chlorinated solvent compounds, <u>In</u>, Pankow, J.F., and Cherry, J.A., eds., Dense Chlorinated Solvents and Other DNAPLs in Groundwater: History, Behavior, and Remediation: Waterloo Press, Waterloo, Ontario, p. 267-312.
- Buscheck, T.E. and Alcantar, C.M., 1995, "Regression Techniques and Analytical Solutions to Demonstrate Intrinsic Bioremediation." In, Proceedings of the 1995 Battelle International Conference on In-Situ and On Site Bioreclamation, April 1995
- Chapelle, F.H., 1993, Ground-Water Microbiology and Geochemistry: John Wily & Sons, Inc., New York, 424 p.
- Chapelle, F.H., P.B. McMahon, N.M. Dubrovsky, R.F. Fujii, E.T. Oaksford, and D.A. Vroblesky. 1995. Deducing the distribution of terminal electron-accepting processes in hydrologically diverse groundwater systems. Water Resources Research, 31:359-371.
- Chapelle, F.H., 1996, Identifying redox conditions that favor the natural attenuation of chlorinated ethenes in contaminated ground-water systems, In Proceedings of the Symposium on Natural Attenuation of Chlorinated Organics in Ground Water, Dallas, TX: EPA/540/R-96/509, September 1996.
- Cline, P.V., and Delfino, J.J., 1989, Transformation kinetics of 1,1,1-trichloroethane to the stable product 1,1-dichloroethene, In: Biohazards of Drinking Water Treatment. Lewis Publishers, Inc. Chelsea, MI. p. 47-56.
- Davis, J.W., Klier, N.J., and Carpenter, 1994, Natural biological attenuation of benzene in ground water beneath a manufacturing facility: Ground Water, vol. 32, no. 2, p. 215-226.

- de Bont, J.A.M., M.J.W. Vorage, S. Hartmans, and W.J.J. van den Tweel. 1986. Microbial degradation of 1,3-dichlorobenzene: Applied and Environmental Microbiology, v. 52, p. 677-680.
- De Bruin, W.P., M.J.J. Kotterman, M.A. Posthumus, G. Schraa, and A.J.B. Zehnder. 1992. Complete biological reductive transformation of tetrachloroethene to ethane: Applied and Environmental Microbiology, v. 58, no. 6, p. 1966 2000.
- DeStefano, T.D., Gossett, J.M., and Zinder, S.H., 1991, Reductive dehalogenation of high concentrations of tetrachloroethene to ethene by an anaerobic enrichment culture in the absence of methanogenesis: Appl. Environ. Microbiol., vol. 57, no. 8, p. 2287-2292.
- Dewulf, J., D. Drijvers, and H. Van Langenhove. 1995. Measurement of Henry's law constant as function of temperature and salinity for the low temperature range. *Atmos. Environ.*, 29(3):323-331.
- Domenico, P.A., and F.W. Schwartz. 1990. Physical and Chemical Hydrogeology. John Wiley and Sons, New York, New York, 824p.
- Engineering-Science, Inc. (ES), 1982, Installation Restoration Program, Phase I Records Search, Tinker AFB, Oklahoma.
- Feenstra, S. and N. Guiguer. 1996. Dissolution of dense non-aqueous phase liquids in the subsurface. In: Pankow, J.F., and Cherry, J.A. (eds.), Dense Chlorinated Solvents And Other DNAPLS In Groundwater. Waterloo Press, Portland, OR. Folsom, B.R., Chapman, P.J., and Pritchard, P.H., 1990, Phenol and trichloroethylene degradation by Pseudomonsa cepacia G4: Kinetics and interactions between substrates: Appl. Environ. Microbiol., vol. 56, no. 5, p. 1279-1285.
- Folsom, B.R., Chapman, P.J., and Pritchard, P.H., 1990, Phenol and trichloroethylene degradation by *Pseudomonsa cepacia* G4: Kinetics and interactions between substrates: Appl. Environ. Microbiol., vol. 56, no. 5, p. 1279-1285.
- Franke, O.L., T.E. Reilly, and G.D. Bennett. 1987. Definition of boundary and initial conditions in the analysis of saturated ground-water flow systems an introduction. In: United States Geological Survey Techniques of Water-Resources Investigations, Book 3, Chapter B5.
- Freedman, D.L., and Gossett, J.M., 1989, Biological reductive dehalogenation of tetrachloroethylene and trichloroethylene to ethylene under methanogenic conditions: Appl. Environ. Microbiol., vol. 55, no. 4, p. 1009-1014.
- Geraghty & Miller Modeling Group. 1994. AQTESOLV Aquifer Test Solver, Version 2.0. Millersville, Maryland. October.
- Godsey, E.M. 1994. Microbiological and Geochemical Degradation Processes, In: Symposium on Intrinsic Bioremediation in Ground Water. Denver, Colorado. August 30 September 1. p.35-40.

- Goldstein, R.M., L.M. Mallory, and M. Alexander. 1985. Reasons for Possible Failure of Inoculation to Enhance Biodegradation: Applied Environmental Microbiology, vol. 50, no. 4, p. 977-983.
- Grbic'-Galic', D., 1990, Anaerobic microbial transformation of nonoxygenated aromatic and alicyclic compounds in soil, subsurface, and freshwater sediments, In: Bollag, J.M., and Stotzky, G., eds.: Soil Biochemistry: Marcel Dekker, Inc., New York, NY. p. 117-189.
- Haigler, B.E., Nishino, S.F., and Spain, J.C., 1988, Degradation of 1,2-dichlorobenzene by a *Pseudomonas sp.*: Applied and Environmental Microbiology, v. 54, p. 294-301.
- Harker, A.R., and Kim, Y., 1990, Trichloroethylene degradation by two independent aromatic-degrading pathways in *Alcaligenes eutrophus* JMP134: Appl. Environ. Microbiol., vol. 56, no. 4, p. 1179-1181.
- Hartmans, S., and de Bont, J.A.M., 1992, Aerobic vinyl chloride metabolism in *Mycobacterium aurum Li*: Appl. Environ. Microbiol., vol. 58, no. 4, p. 1220-1226.
- Henry, S.M., 1991, Transformation of Trichloroethylene by Methanotrophs from a Groundwater Aquifer. Ph.D. Thesis. Stanford University. Palo Alto, California.
- Howard, P.H. 1990. Handbook of Environmental Fate and Exposure Data for Organic Chemicals, Vol. II: Solvents: Lewis Publishers, Inc., Chelsea, Michigna. 546 p.
- IT Corporation, 1994, Final Report Phase I RCRA Facility Investigation for Appendix I Sites, Tinker AFB, Oklahoma, September 1994.
- Jeffers, P.M., L.M. Ward, L.M. Woytowitch, and N.L. Wolfe. 1989. Homogeneous hydrolysis rate constants for selected chlorinated methanes, ethanes, ethenes, and propanes: Environmental Science and Technology, v. 23, p. 965-969.
- Klier, N.J., West, R.J., and Donberg, P.A., 1996, Aerobic Biodegradation of Dichloroethylenes in Surface and Subsurface Soils: Accepted for Publication in Chemosphere, December 1996.
- Klöpffer, W., F. Haag, E.G. Kohl, and R. Frank. 1988. "Testing of the Abiotic Degradation of Chemicals in the Atmosphere: The Smog Chamber Approach," *Ecotoxicol. Environ. Saf.* 15(3);298-319.
- Konikow, L.F., and J.D. Bredehoeft. 1978. Computer model of two-dimensional solute transport and dispersion in groundwater. *United States Geological Survey, Techniques of Water Resources Investigations of the United States Geological Survey*, Book 7, Chapter C2, 90 p.
- Little, C.D., Palumbo, A.V., Herbes, S.E., Lidstrom, M.E., Tyndall, R.L., and Gilmer, P.J., 1988, Trichloroethylene biodegradation by a methane-oxidizing bacterium: Appl. Environ. Microbiol., vol. 54, no. 4, p. 951-956.

- Lovely, D.R., and Phillips, E.J.P., 1988. Novel Mode of Microbial Energy Metabolism: Organic Carbon Oxidation Coupled to Dissimilatory Reduction of Iron or Maganese: Applied and Environmental Microbiology, v. 54, no. 6, p. 1472-1480.
- Lovely, D.R., Phillips, E.J.P., and Lonergan, D.J., 1991, Enzymatic versus nonenzymatic mechanisms for Fe(III) reduction in aquatic sediments: Environmental Science and Technology, v. 26, no. 6, p. 1062-1067.
- Lyman, W.J., P.J. Reidy, and B. Levy. 1992. Mobility and Degradation of Organic Contaminants in Subsurface Environments. C.K. Smoley, Inc., Chelsea, MI.
- Mackay, D. and W.-Y. Shiu. 1981. "A Critical Review of Henry's Law Constant for Chemicals of Environmental Interest," J. Phys. Chem. Ref. Data, 10(4):1175-1199.
- Mayer, K.P., Grbic-Galic, D., Semprini, L., and McCarty, P.L., 1988, Degradation of trichloroethylene by methanotrophic bacteria in a laboratory column of saturated aquifer material: Wat. Sci. Tech. (Great Britain), vol. 20, no. 11/12, p. 175-178.
- McCarty, P.L., Roberts, P.V., Reinhard, M., and Hopkins, G., 1992, Movement and transformations of halogenated aliphatic compounds in natural systems, In: Fate of Pesticides and Chemicals in the Environment. Ed., J.L. Schnoor. John Wiley & Sons, Inc. New York, New York. p. 191-209.
- McCarty, P.L., and Semprini, L., 1994, Ground-Water Treatment for Chlorinated Solvents, In: Handbook of Bioremediation. Lewis Publishers, Boca Raton, FL. 1994.
- McDonald, G. and A.W. Harbaugh. 1988. A modular three-dimensional finite-difference groundwater flow model. *US Geological Survey Techniques of Water Resources Investigations*, Book 6, Chapter A1.
- Miller, R.E., and Guengerich, F.P., 1982, Oxidation of trichloroethylene by liver microsomal cytochrome P-450: Evidence for chlorine migration in a transition state not involving trichloroethylene oxide: Biochemistry, vol. 21, p. 1090-1097.
- Miser, Hugh D., et al., 1954, Geologic Map of the State of Oklahoma, Oklahoma Geological Survey.
- Moutoux, D.E., Benson, L.A., Swanson, T.H., Wiedemeier, T.H., Lenhart, J., Wilson, J.T., and Hansen J.E., 1996, Estimating the Changing Rate of Anaerobic Reductive Dechlorination of Chlorinated Aliphatic Hydrocarbons in the Presence of Petroleum Hydrocarbons. Proceedings of the 1996 API/NGWA Conference on Petroleum Hydrocarbons and Organic Chemicals in Ground Water, November, 1996, Houston, Texas.
- Murray, W.D. and Richardson, M., 1993, Progress toward the biological treatment of C₁ and C₂ halogenated hydrocarbons: Critical Reviews in Environmental Science and Technology, v. 23, no. 3, pp. 195-217.
- Nelson, M.J.K., Montgomery, S.O., O'Neille, E.J., and Pritchard, P.H., 1986, Aerobic metabolism of trichloroethylene by a bacterial isolate: Appl. Environ. Microbiol., vol. 52, no. 2, p. 949-954.

- Nishino, S.F., Spain, J.C., and Pettigrew, C.A., 1994, Biodegradation of chlorobenzene by indigeneous bacteria: Environ. Toxicol. Chem., v. 13, p. 871-877.
- Pankow, J.F. and M.E. Rosen. 1988. "Determination of Volatile Compounds in Water by Purging Directly to a Capillary Column with Whole Column Cryotrapping," *Environm. Sci. Technol.*, 22(4):398-405.
- Parkhurst, D.L., Christenson, S.C., and Breit, G.N., 1993, Groundwater-Water-Quality Assessment of the Central Oklahoma Aquifer, Oklahoma: Geochemical and Geohydrologic Investigations: U.S. Geological Survey Open-File Report 02-642, 216p.
- Parsons ES, 1997. Work Plan for a Demonstration of Remediation by Natural Attenuation for Groundwater at Sites FTA 2 and Area A. Tinker Air Force Base, Oklahoma City, Oklahoma. March.
- Pearson, C.R. and G. McConnell. 1975. "Chlorinated C1 and C2 Hydrocarbons in the marine Environment," *Proc. R. Soc. London*, B189(1096):305-332.
- Reineke, W., and Knackmuss, H.-J., 1984, Microbial metabolism of haloaromatics: Isolation and properties of a chlorobenzene-degrading bacterium: European Journal of Applied Microbiology and Biotechnology, v. 47, p. 395-402.
- Reinhard, M., Goodman, N.L., and Barker, J.F., 1984, Occurrence and distribution of organic chemicals in two landfill leachate plumes: Environ. Sci. Technol., vol. 18, no. 12, p. 953-961.
- Sander, P., R.-M. Wittaich, P. Fortnagel, H. Wilkes, and W. Francke. 1991. Degradation of 1,2,4-trichloro- and 1,2,4,5-tetrachlorobenzene by *Pseudomonas* strains: Applied and Environmental Microbiology, v. 57, p. 1430-1440.
- Spain, J.C. 1996. Future vision: Compounds with potential for natural attenuation, In Proceedings of the Symposium on Natural Attenuation of Chlorinated Organics in Ground Water: EPA /540/R-96/509, Dallas TX, September 11-13.
- Spain, J.C., and S.F. Nishino. 1987. Degradation of 1,4-dichlorobenzene by a *Pseudomonas sp.*: Applied and Environmental Microbiology, v. 53, p. 1010-1019.
- Spitz, K. and J. Moreno. 1996. A Practical Guide to Groundwater and Solute Transport Modeling. John Wiley & Sons, Inc., New York, 461 p.
- Stumm, W. and Morgan, J.J. 1981. Aquatic Chemistry. John Wiley & Sons, New York, NY.
- Suflita, J.M., and G.T. Townsend. 1995. The microbial ecology and physiology of aryl dehalogenation reactions and implications for bioremediation, In Young, L.Y., and Cerniglia, C.E., eds., Microbial Transformation and Degradation of Toxic Organic Chemicals: Wiley-Liss, New York, 654 p.
- Tetra Tech, Inc., 1996, Draft RCRA Facility Investigation for Selected Appendix I Sites, Tinker AFB, Oklahoma, RFI Draft, October 1996.

- U.S. Army Corps of Engineers (USACE), 1988, Tinker Air Force Base, Installation and Restoration Program, Fire Training Area No. 2, Tulsa District, December 1988.
- USEPA. 1991. Water Quality Criteria Summary Concentrations, Office of Science and Technology, Health and Ecological Criteria Division, Ecological Risk Assessment Branch (WH-585), Human Risk Assessment Branch (WH-550D), Washington, D.C.
- USEPA, 1993. Internal Memorandum Discussing Revisions to OMB Circular A-94 on Guidelines and Discount Rates for Benefit-Cost Analysis. June 25.
- USEPA. 1996. National Primary Drinking Water Regulations.
- van der Meer, J.R., W. Roelofsen, G. Schraa, and A.J.B. Zehnder. 1987. Degradation of low concentrations of dichlorobenzenes and 1,2,4-trichlorobenzene by *Pseudomonas* sp. strain P51 in nonsterile soil columns: FEMS Microbiology Letters, v. 45, p. 333-341.
- Vogel, T.M. 1994. Natural bioremediation of chlorinated solvents. In: *Handbook of Bioremediation*. Lewis Publishers, Boca Raton, FL.
- Vogel, T.M., and McCarty, P.L., 1985, Biotransformation of tetrachloroethylene to trichloroethylene, dichloroethylene, vinyl chloride, and carbon dioxide under methanogenic conditions: Applied Environmental Microbiology, v. 49, no. 5, pp. 1080-1083.
- Vogel, T.M., Criddle, C.S., and McCarty, P. L., 1987, Transformations of halogenated aliphatic compounds: Environmental Science and Technology, vol. 21, no. 8, p. 722-736.
- Wiedemeier, T.H., Wilson, J.T., Kampbell, D.H., Miller, R.N., and Hansen, J.E., 1995, Technical Protocol for Implementing Intrinsic Remediation with Long-term Monitoring for Natural Attenuation of Fuel Contamination Dissolved in Ground Water. Prepared by the Air Force Center for Environmental Excellence.
- Wiedemeier, T.H., Hansen, J.E., Haas, P., Wilson, J.T., Kampbell, D.H., and Chapelle, F.H., 1996a, *Technical Protocol for Natural Attenuation of Chlorinated Aliphatic Hydrocarbons in Groundwater*. Prepared by the Air Force Center for Environmental Excellence (Draft).
- Wiedemeier, T.H., Benson, L.A., Wilson, J.T., Kampbell, D.H., Hansen, J.E., and Miknis, R., 1996b, Patterns of natural attenuation of chlorinated aliphatic hydrocarbons at Plattsburgh Air Force Base, New York: In: Conference on Intrinsic Remediation of Chlorinated Solvents. Salt Lake City, UT. April 2, 1996.
- Wilson, J.T., and Wilson, B.H., 1985, Biotransformation of trichloroethylene in soil: Appl. Environ. Microbiol., vol. 49, no. 1, p. 242-243.
- Wilson, J.T., 1996, Personal communication between John Wilson of the UEPA NRMRL and Todd Wiedemeier of Parsons ES regarding USEPA definition of the term "natural attenuation", citing USEPA guidance undergoing internal review.

- Wood, P.R., and Burton, L.C., 1968, Groundwater-Resources, Cleveland and Oklahoma Counties, Oklahoma Geological Survey, Circular 71.
- Zheng, C. 1990. MT3D A Modular Three-Dimensional Transport Model for Simulation of Advection, Dispersion, and Chemical Reactions of Contaminants in Groundwater Systems; Prepared for the US Environmental Protection Agency Robert S. Kerr Environmental Research Laboratory, October 17, 1990.

APPENDIX A

SELECTED FIGURES AND TABLES FROM THE PHASE I RCRA RFI FOR FTA-2 (IT, 1994), AND THE DRAFT RCRA RFI FOR FTA-2 (TT, 1996)

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-62B

WELL COMPLETION DATA

Boring Location: EAST OF

SURFACE ELEV.(FT): 1245.940

FIRE TRAINING AREA TOTAL DEPTH(FT.): 33.0

DRILLING AND SAMPLING INFORMATION

Logged By: Drilled By:

K. KIRSCHENMANN Date Storted: P. GUERREIN

Date Completed:

11/9/93 11/9/93

GEOTECHNOLOGY. INC.

Drill Rig Type: CME-75

Drilling Method: 8" HOLLOW STEM AUGER

Sampling Method: 3"x5' CONTINUOUS SAMPLER

Elev-Top of Casing(ft.):

1. Riser Pipe-I.D.(in.): 2

Centralizers-Type:

2. Screen Dia.(in.):

Depth Interval(ft.):14-24 Centralizers-Type: S.Steel

3. Filter Pack Type: Silica Sand

Conc. Pad Size: 4'x4'x6"

Ref. Datum: MSL

Depth(ft.): 14 Type:S.Steel

Depths(ft.):

Type: S.Steel Millslotted

Slot Size(in.): .010

Depths(ft.):23

Depth Interval(ft.): 12-25

Notes: N 150444,517, E 2182140,558 GEOPHYSICAL IN. DRIVEN RECOVERED LOG (GAMMA) 200 SAMPLE SAMPLE NO. ĸ PPM GRAPHIC 12" Steel Concrete DEPTH Manhole nscs Flush e Cover Completion DESCRIPTION ż SILTY CLAY - nonplastic; hard; slightly maist; strong brown (7.5YR-4/6) - dark red (2.5YR-4/8); gradational contact cement/ bentonite 60 grout riser pipe (threaded) 7.0 60 - 10% very fine grained sand; very hard; dry; 1/4 inch 10 bentonite pellets 12.0 SW SAND — very fine grained; 5% silt and clay; compact; dry; pale yellow (2.5Y-7/3); cl sharp contact 14.0 SILTY CLAY — nonplastic; 10% very fine grained sand; very hard; dry; dark red (2.5YR-4/8); 15 sharp contact - switched to center bit at 13.0° (very hard drilling) well screen - alternating very hard layers (17.0-20.0') 20 filter pack sm SILTY SAND - approximate level of water bearing 20 mesh sediments stoinless centralizer 24.0 25.0 25 s.stl cap 1/4 inch bentonite pellets 30 8 inch borehole 33.0 TOTAL DEPTH = 33.0 FEET 35 40 **RPS** DWG. 409832-A30 PROJ. APPRV. DRAFT DRAFT Sheet 1 of 1 BY NO. BY 12/1/93 CHK CHK

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-63B

DRILLING AND SAMPLING INFORMATION

SURFACE ELEV.(FT): 1243.284 Boring Location: EAST OF FIRE TRAINING AREA TOTAL DEPTH(FT.): 29.0

K. KIRSCHENMANN Date Started:

11/8/93 11/8/93

Logged By: P. GUERREIN Drilled By:

Date Completed:

GEOTECHNOLOGY. INC.

CME-75

Drill Rig Type: Drilling Method: 8" HOLLOW STEM AUGER

Sampling Method: 3"x5' CONTINUOUS SAMPLER

WELL COMPLETION DATA

Elev-Top of Casing(ft.):

1. Riser Pipe-1.D.(in.): 2 Centralizers-Type:

2. Screen Dia.(in.): Depth Interval(ft.):18-23 Centralizers-Type: S.Steel

3. Filter Pack Type: Silica Sand Conc. Pad Size: 4'x4'x6"

Ref. Dotum: MSL

Depth(ft.): 18 Type:S.Steel

Depths(ft.):

Type: S.Steel Millslotted Slot Size(in.): .010 Depths(ft.): 22

Depth interval(ft.): 16-23

Notes: N 150386.113, E 2182035.358 SAMPLE NO.
IN. DRIVEN
IN. RECOVERED GEOPHYSICAL LOG (GAMMA) SAMPLE z PPM GRAPHIC 12" Steel Concrete DEPTH Manhole Flush USCS Completion ő. Cover DESCRIPTION SILTY CLAY — slightly plastic; <10% very fine grained sand; stiff; slightly maist; dark brown (7.5YR-3/4) cement/ - <5% to 10% very fine grained sand; dark red 60 bentonite (2.5YR-4/8); gradational contact grout - nonplastic; hard; dry riser pipe (threaded) 60 60 SANDY SILTY CLAY - nonplostic; 10% to 20% cl 10fine grained sand; hard; dry; 11.0 dark red (2.5YR-4/8); gradational contact 50 SAND — very fine grained; compact; dry; pale yellow (2.5YR-8/2); sharp contact sm SILTY SAND - 40% clay and silt; 50% to 60% fine to very fine grained sand; firm; dry; dark red (2.5YR-4/B); sharp contact - 60% fine to medium grained sand; 1/4 inch 60 bentonite 48 Y pellets 15 16.0 slightly moist; gradational contact alternating layers of siltstone and silty sand at 15'
— moist
— drill without sampling due to auger refusal filter pack 18.0 20 mesh well screen 20 stainless steel centralizer 23.0 s.stl cop 25 1/4 inch bentonite pellets 8 inch borehole 29.0 TOTAL DEPTH = 29.0 FEET 30 35 40 409832-A33 DWG. APPRV PROJ. RPS DRAFT DRAFT Sheet 1 of 1 NO. BY CHK CHK 12/2/93

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-64B

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTHEAST OF

SURFACE ELEV.(FT): 1245.586

FIRE TRAINING AREA TOTAL DEPTH(FT.): 33.0 K. KIRSCHENMANN Date Started:

11/9/93

Logged By: Drilled By:

J.

Drill Rig Type:

P. GUERREIN GEOTECHNOLOGY. INC.

Sampling Method: 3"x5" CONTINUOUS SAMPLER

CME-75

Drilling Method: 8" HOLLOW STEM AUGER

Date Completed:

1. Riser Pipe-I.D.(in.): 2 Centralizers-Type: 11/9/93

2. Screen Dia.(in.): Depth Interval(ft.):16-26

Elev-Top of Casing(ft.):

Centralizers-Type: S.Steel

3. Filter Pack Type: Silica Sand

Ref. Datum: MSL

Depth(ft.): 16 Type:S.Steel

Depths(ft.):

WELL COMPLETION DATA

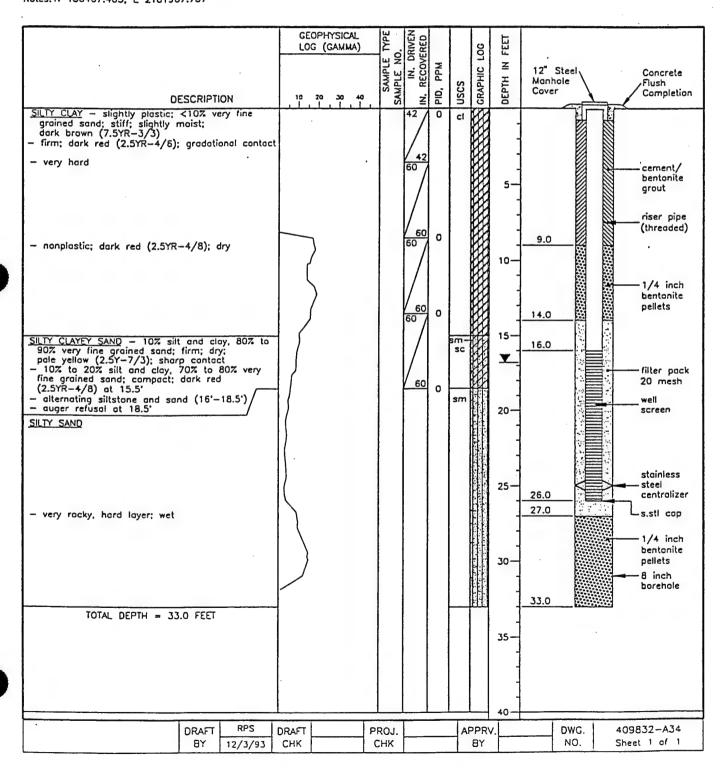
Type: S.Steel Millslotted Slot Size(in.): .010

Depths(ft.): 25

Depth Interval(ft.): 14-27

Conc. Pod Size: 4'x4'x6"

Notes: N 150467.403, E 2181967.757



Project Name: TINKER 5001

Project Number:

Project Location: TINKER AFB. OKLAHOMA 409832

MONITORING WELL 2-65B

WELL COMPLETION DATA

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTHEAST OF

SURFACE ELEV.(FT): 1250.812 FIRE TRAINING AREA TOTAL DEPTH(FT.): 49.0

11/5/93 K. KIRSCHENMANN Date Started:

Logged By: Drilled By:

P. GUERREIN

Date Completed:

11/5/93

GEOTECHNOLOGY. INC.

Drill Rig Type: CME-75

Drilling Method: 8" HOLLOW STEM AUGER

Sampling Method: 3"x5' CONTINUOUS SAMPLER

Elev-Top of Casing(ft.):

1. Riser Pipe-I.D.(in.): 2

Centrolizers-Type: S.Steel 2. Screen Dia.(in.):

Depth Interval(ft.):37-47 Centralizers-Type: S.Steel

3. Filter Pack Type: Silica Sand Conc. Pod Size: 4'x4'x6"

Ref. Datum: MSL

Type:S.Steel Depth(ft.): 37

Depths(ft.): 25

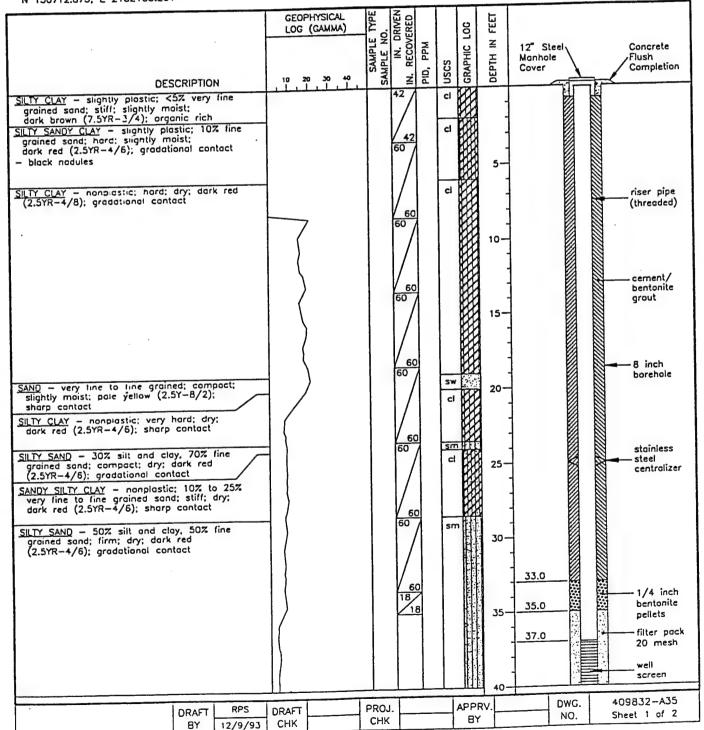
Type: S.Steel Millslotted

Slot Size(in.): .010

Depths(ft.): 46

Depth Interval(ft.):35-49

Notes: NO ANALYTICAL SAMPLES TAKEN N 150712.875, E 2182183.201



Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-65B

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTHEAST OF

SURFACE ELEV.(FT): 1250.812

FIRE TRAINING AREA TOTAL DEPTH(FT.): 49.0

Logged By: Drilled By:

P. GUERREIN

K. KIRSCHENMANN Date Started: Date Completed:

11/5/93 11/5/93

GEOTECHNOLOGY. INC.

Drill Rig Type: CME-75

Drilling Method: 8" HOLLOW STEM AUGER

Sampling Method: 3"x5' CONTINUOUS SAMPLER

WELL COMPLETION DATA

Elev-Top of Casing(ft.):

1. Riser Pipe-1.D.(in.): 2

Centralizers-Type: S.Steel 2. Screen Dia.(in.): 2

Depth Interval(ft.):37-47 Centralizers-Type: S.Stee!

3. Filter Pock Type: Silica Sand Depth Interval(ft.): 35-49 Conc. Pad Size: 4'x4'x6"

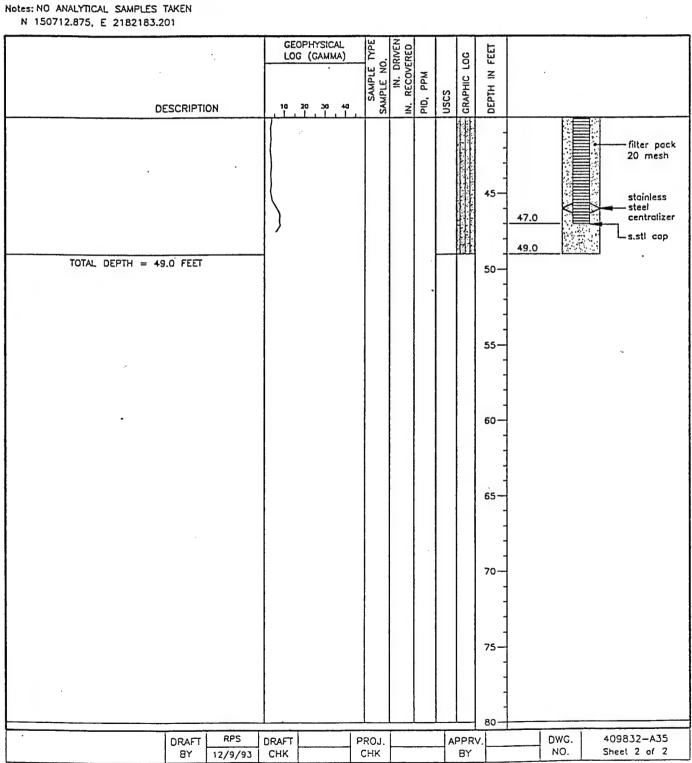
Ref. Datum: MSL

Depth(ft.): 37 Type:S.Steel

Depths(ft.): 25

Type: S.Steel Millslotted Slot Size(in.): .010

Depths(ft.): 46

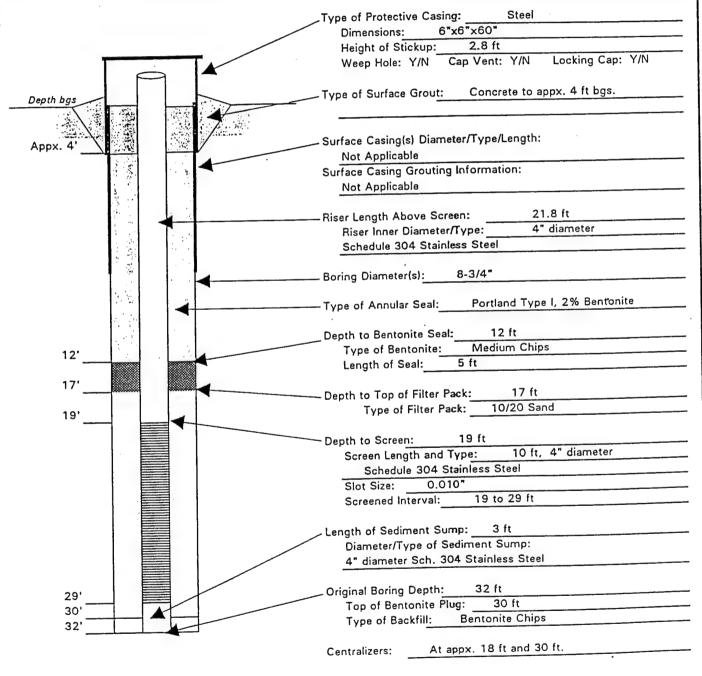


SITE FTA2 BORING NUMBER 2-167B PROJECT _____TAFB Phase II RFI App. I Sites, Fire Train. Area 2 DATE DRILLED O7 August 1995 (Drilled as MWI) DRILLING METHOD 6" Hollow-stem Augers; 5' Continuous Sampler DRILLING COMPANY Associated Environmental Industries ELEVATION 1247.70 Feet msl, Ground Surface; 1,250.74 Feet msl, TOC TOTAL DEPTH __ 32 Feet GEOLOGIST _ R. Osgood X,Y COORD. 2182408.352, 150539.8117 SAMPLE (mdd) LITHOLOGY MUNSELL eotechnica DEPTH feet GRAPHIC Analytical Interval Recovery DESCRIPTION AND REMARKS Topsoil, silty, w/vegetation and subrounded, pebble-sized clasts. Clayey silt, reddisg brown, soft, slightly moist, low plasticity, some ND black organic speckling throughout. 2.5YR4/4 At 4 ft, grades to silt, red, firm, dry, a few small subrounded ND pebbles: light grey marbling at base. 2.5YR4/6 Clayey silt, reddish brown, firm, dry, some black organic speckling, and trace of light grey marbling (mineralization?). ND 胀 2.5YR4/4 ND Clayey silt, red, firm, dry, massive (homogeneous). ND 2.5YR4/6 No Recovery Silt, light grey, stiff, moist for top I", Interval grades from silt to sandy silt, red, moist, soft, massive. At 12 ft, moderately ND cemented sandstone clasts appx. I" in diameter. 2.5YR4/6 ML ND No Recovery -20 Sandy silt (appx 75% silt), red. soft, very moist, few I" moderately. cemented sandstone fragments; lens of fine sand appx. I" thick at ND 21.5 ft. ML 2.5YR4/6 NO No Recovery SM

										SITE FTA2 BORING NUMBER 2-167B					
								PROJECT							
						•		DATE DRILLEDO7 August 1995 (Drilled as MWI) *							
										NG METHOD 6" Hollow-stem Augers: 5' Continuous Sampler					
										NG COMPANY Associated Environmental Industries					
Ì									ELEVATION 1247.70 Feet msl, Ground Surface; 1,250.74 Feet msl, TOC						
									TOTAL DEPTH 32 Feet						
ı									DEPTH TO WATER						
									GEOLOGIST						
							7	,							
		SAM	IPLE			Ê		≿	90						
	Analytical	Geotechnical	val	Recovery	DEPTH feet	PID (ppm)	MUNSELL	LITHOLOGY	GRAPHIC LOG	DESCRIPTION AND REMARKS					
	alyt	tecl	Interval	COV	H a	유	MON	ΙĔ	AP.	SESSIA TON AND REMAINS					
ĺ	A	Geo	Ξ	Re				-	69						
	\bowtie									Silty sand (60%), reddish brown, soft, wet, very fine sand laminated					
					-	ND	2.5YR4/4	SM		throughout.					
			ı		-			<u> </u>		No Recovery					
- [İ		ı		-	İ		'		in the mesoner,					
-			1		_			1							
					20										
	\times				-30	ļ			10000	Sand, red, fine grained, poorly graded, wet, homogeneous color and					
			ł		-			SP		massive texture.					
		}	l	Ц	-					Constitution and the second se					
'					-	ĺ				Completed as monitoring well 2–187 on 07 August 1995.					
										·					
١	ı									•					
١		1		Ì	-35										
		ļ		}	-										
١					_										
1			İ												
			ļ		-										
1				ŀ	-										
				-	-40										
			ļ							·					
۱															
		ı													
				ŀ	•										
1				ŀ	.										
1					-45										
1										,					
				-											
				-											
				}	.										
					-50										
					50										
1															

Tetra Tech Monitoring Well Construction Log

Project/Site Name:	TAFB Phase II RFI	, FTA2	Well No:	2-167 B	Date:	08 Aug 95		
Project No: 7	107	Drilling Method:	8-3/4"	Augers	ugers			
Geologist: R. Osgood								
Elevation of Casing: 1250.7 Ft msl		Drilling Contract			onmental Indu			
Surface Elevation: 1247.7 Ft msl		Date Well Construction Completed:			08 Aug 95	08 Aug 95		

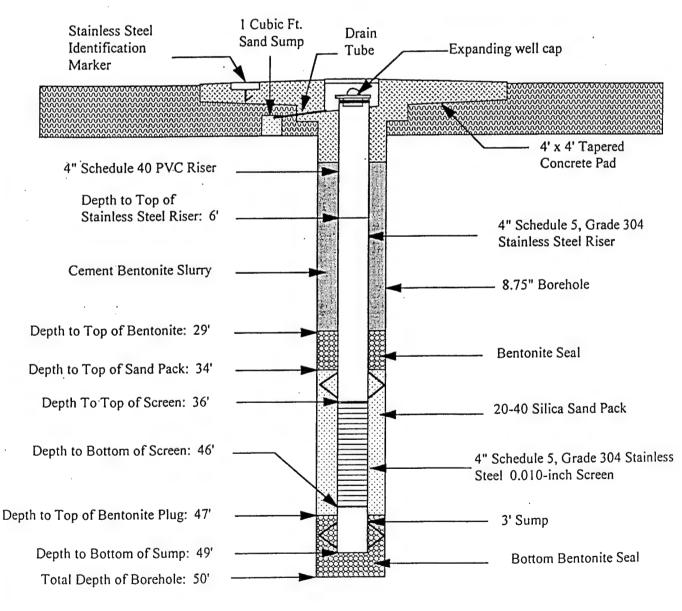


SITE FTA2 BORING NUMBER 2-168B PROJECT TAFB Phase II RFI App. I Sites, Fire Train. Area 2 DATE DRILLED <u>08 August 1995 (Drilled as MH2)</u> DRILLING METHOD 6" Hollow-stem Augers: 5' Continuous Sampler DRILLING COMPANY Associated Environmental Industries ELEVATION 1240.10 Feet msl, Ground Surface; 1,243.00 Feet msl, TOC TOTAL DEPTH _25 Feet DEPTH TO WATER ____10.6 Feet bgs GEOLOGIST _ R. Osgood X,Y COORD. 2181976.691, 150248.99500 SAMPLE (mdd) LITHOLOGY JUNSELL eotechnica DEPTH feet Analytical GRAPHIC Recovery DESCRIPTION AND REMARKS 디 Topsoil, clayey silt, dark reddish brown, soft, slightly moist, root hairs extend to 1.5 ft. Clayey silt stiffens to base of core with 2.5YR3/4 trace of organic speckling from 2 to 5 ft. 15 ND Clayey silt, reddish brown, soft to firm, slightly moist, massive, some black organic speckling and small nodules. Form 7 to 10 ft, some 2.5YR4/4 1.0 subangular pebbles and increase in the amount of black organics. ND ND Clayey silt as above to 10.8 ft. Sandy silt, red, soft, moist to saturated at 13 ft, thin bedding 2.5YR4/4 21 (laminated), black organics from 11.5 to 12 ft. ML 4/8 2.5YR4/6 No Recovery Silty sand, red, wet, laminated layers (2-5 mm). From 17 to 17.5 ft, dark black organic lens with small (<3 mm) nodules. Some black 2.5YR4/6 20 speckles of organic matter to 19 ft. 3 SM 1 No Recovery Silty sand, red, wet, laminated (appx. 2 mm), trace of black organic speckling throughout. 2.5YR4/6 No Recovery

Tetra Tech Monitoring Well Construction Log

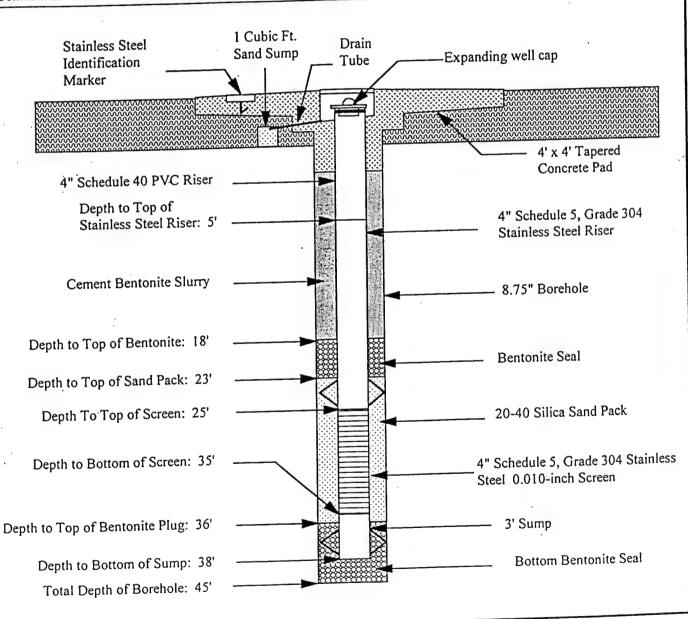
Project/Site Name:	TAFB Phase II RFI,						
Project No: 7	107	Drilling Method: 8-3/4" Hollow-stem Augers					
Geologist: R. Osg	ood	•					
Elevation of Casing:	1243.0 Ft msl	Drilling Contractor: Associated Environmental Industries					
Surface Elevation:	1240.1 Ft msl	Date Well Construction Completed: 08 Aug 95					
Surface Elevation.	1270						
		Type of Protective Casing: Steel					
		Dimensions: 6"x6"x60"					
		Height of Stickup: 2.8 ft					
7		Weep Hole: Y/N Cap Vent: Y/N Locking Cap: Y/N					
Depth bgs		Type of Surface Grout: Concrete to appx. 4 ft bgs					
Jepin ugs							
是 人名							
Appx. 4'		Surface Casing(s) Diameter/Type/Length:					
25.00		Not Applicable					
		Surface Casing Grouting Information:					
[·		Not Applicable					
\$5. \$4.		Riser Length Above Screen: 12.8 ft					
	4	Riser Inner Diameter/Type: 4" diameter					
1.		Schedule 304 Stainless Steel					
		. Ouridadio do l'elemento de la companya del companya de la companya de la companya del companya de la companya					
l.		Boring Diameter(s): 8-3/4"					
1.0		Type of Annular Seal: Portland Type 1, 2% Bentonite					
,							
_].:		Depth to Bentonite Seal: 7 ft					
7'		Type of Bentonite: Medium Chips					
		Length of Seal: 2 ft					
9'		Depth to Top of Filter Pack: 9 ft					
		Depth to Top of Filter Pack: 9 ft Type of Filter Pack: 10/20 Sand					
10'		Type of Filter Fack.					
		Depth to Screen: 10 ft					
		Screen Length and Type: 10 ft, 4" diameter					
		Schedule 304 Stainless Steel					
·		Slot Size: 0.010"					
		Screened Interval: 10 to 20 ft					
·		Length of Sediment Sump: 3 ft					
		Diameter/Type of Sediment Sump:					
		4" diameter Sch. 304 Stainless Steel					
20'		Original Boring Depth: 25 ft					
23'		Top of Bentonite Plug: 23 ft Type of Backfill: Bentonite Chips					
25'		Type of Backfill: Bentonite Chips					
		Centralizers: At appx. 9.5 ft and 21.5 ft					

Client: Tinker Air Force Base	Well ID: 2-271B
Location: TAFB, Oklahoma	Site: FT22
Contract No: F34650-94-D-0082/5017 (P&A #4)	Date 4-Inch Riser Set: 08/08/96
Contractor: Brown & Root Environmental	Ground Level Elevation (AMSL): 1,252.41'
Project Manager: David Parker	Top of Casing Elevation (AMSL): 1,252.13'
Project Geologist: Steve Kelly	Northing Coordinate: 151,494.42'
Drilling Contractor: Associated Environmental, Inc.	Easting Coordinate: 2,182,196.43'
Drilling Method: Hollow Stem Auger	Permanent Monuments Used in Survey: SE32, SE28
Dedicated Pump: 2-inch Grundfos	Legal Description: NE/4 Section 22, T11N, R2W
Comments:	



Client: Tinker Air Force Base	Well ID: 2-272B				
Location: TAFB, Oklahoma	Site: FT22				
Contract No: F34650-94-D-0082/5017 (P&A #4)	Date 4-Inch Riser Set: 08/09/96				
Contractor: Brown & Root Environmental	Ground Level Elevation (AMSL): 1,249.38'				
Project Manager: David Parker	Top of Casing Elevation (AMSL): 1,248.91'				
Project Geologist: Steve Kelly	Northing Coordinate: 150,784.70'				
Drilling Contractor: Associated Environmental, Inc.	Easting Coordinate: 2,181,581.74'				
Drilling Method: Hollow Stem Auger	Permanent Monuments Used in Survey: SE32, SE28				
	Legal Description: SE/4 Section 22, T11N, R2W				
Dedicated Pump: 2-inch Grundfos	Section plate rather than surveyor's pin.				

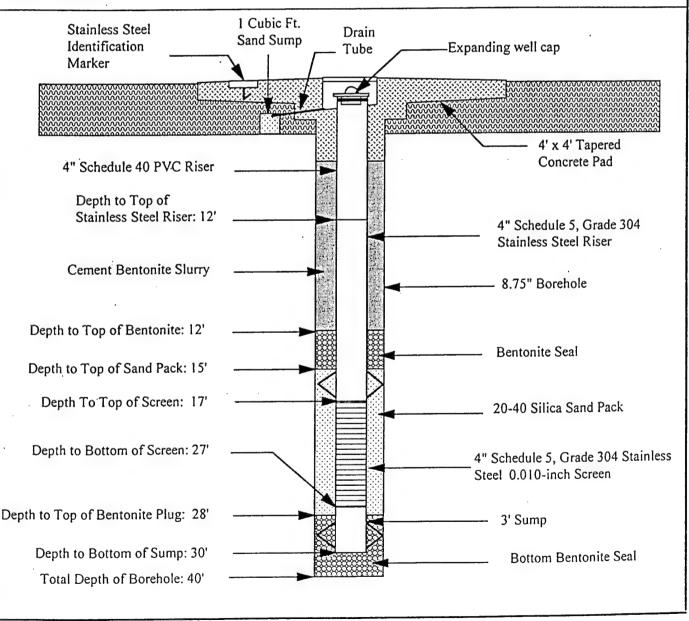
Comments: Ground Level Elevation measured on identification plate rather than surveyor's pin.



Centralizers

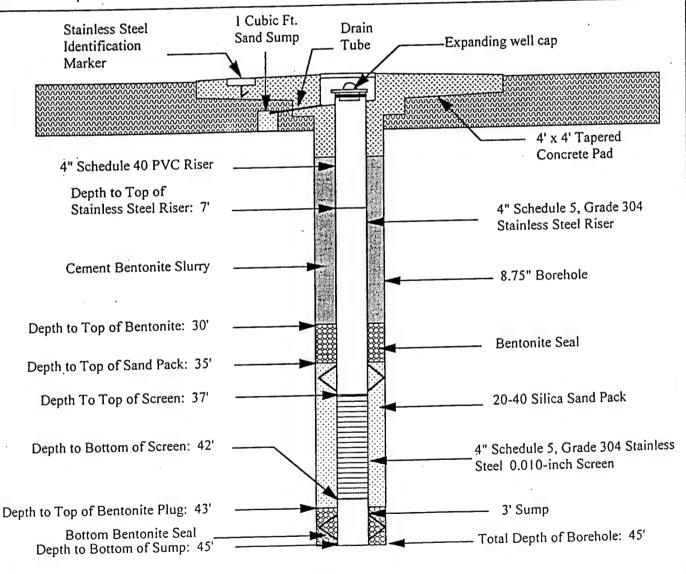
Well_ID: 2-273B
Site: FT22
Date 4-Inch Riser Set: 06/18/96
Ground Level Elevation (AMSL): 1,238.92'
Top of Casing Elevation (AMSL): 1,238.42'
Northing Coordinate: 150,351.11'
Easting Coordinate: 2,181,576.36'
Permanent Monuments Used in Survey: SE32, SE28
Legal Description: SE/4 Section 22, T11N, R2W

Comments: Ground Level Elevation measured on identification plate rather than surveyor's pin. 3' thick top bentonite seal set.



Client: Tinker Air Force Base	Well ID: 2-274B
Location: TAFB, Oklahoma	Site: FT22
Contract No: F34650-94-D-0082/5017 (P&A #4)	Date 4-Inch Riser Set: 07/25/96
Contractor: Brown & Root Environmental	Ground Level Elevation (AMSL): 1,248.77'
Project Manager: David Parker	Top of Casing Elevation (AMSL): 1,248.32'
Project Geologist: Steve Kelly	Northing Coordinate: 151,126.23'
Drilling Contractor: Associated Environmental, Inc.	Easting Coordinate: 2,181,625.92'
Drilling Method: Hollow Stem Auger	Permanent Monuments Used in Survey: SE32, SE28
Dedicated Pump: 2-inch Grundfos	Legal Description: NE/4 Section 22 T11N, R2W
	1 C 17 1 Floreties mangured on

Comments: 5' long screen set. Well set at bottom of borehole. Ground Level Elevation measured on identification plate rather than surveyor's pin.



MW1-67B

RISER SCREEN SAND PACK BENTONITE GROUT

0-34 34-44 32-44 30-32

0 - 30

MW 1-67 A

Data Brand Brand

W' SURFACE CASING-RISER SCREEN SAND PACK BENTONITE GROUT

0-46 0-75 75-85 73-85 71-73

0-71

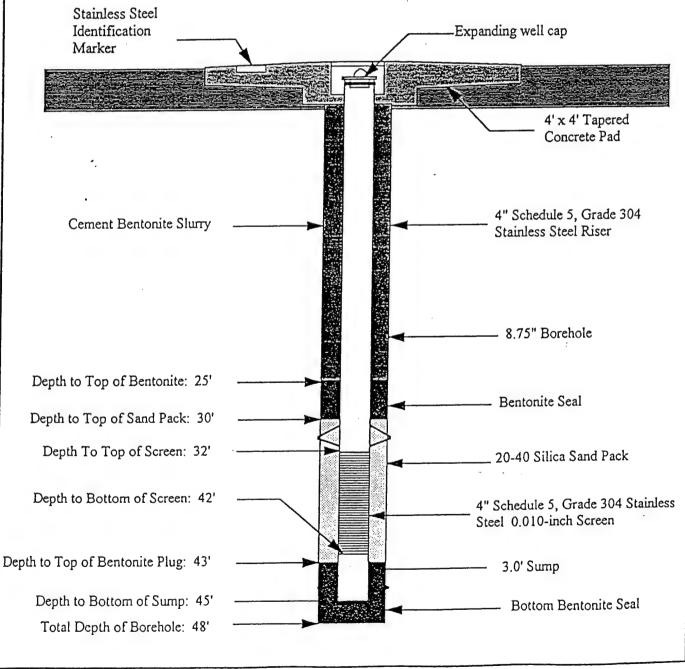
MW 1-67C

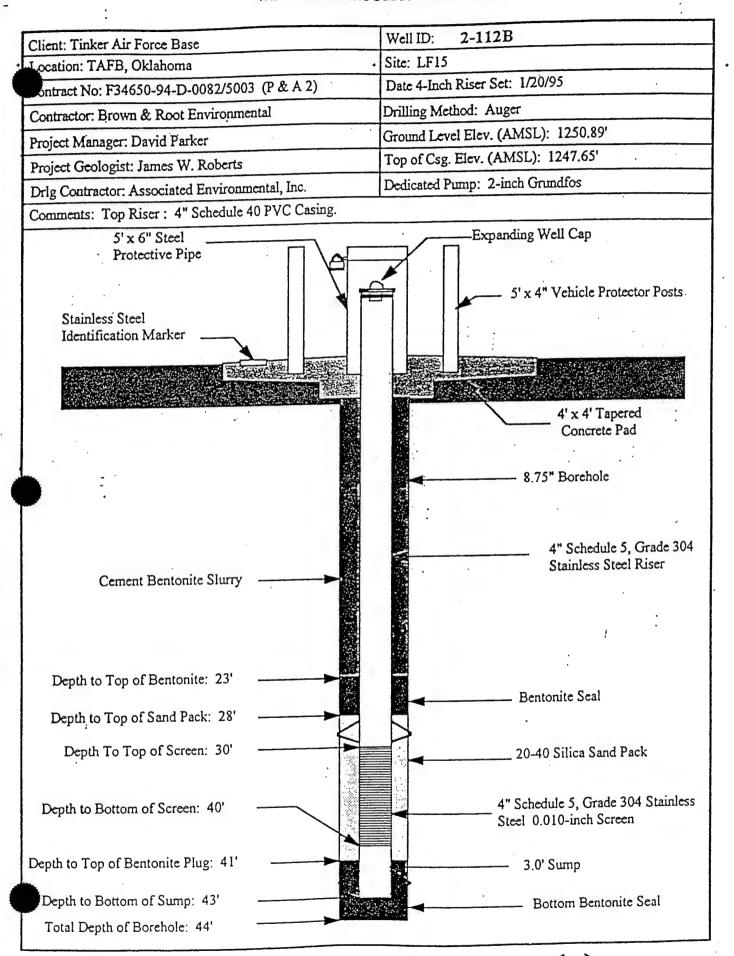
16" SURFACE CASING10" CASINGRISER
SCREEN
SCREEN
SAND PACK
BENTON ITE
GROUT

0-46 0-90 0-109 109-119 107-119 105-107

MW 1-67 WELL COMPLETION INFO.

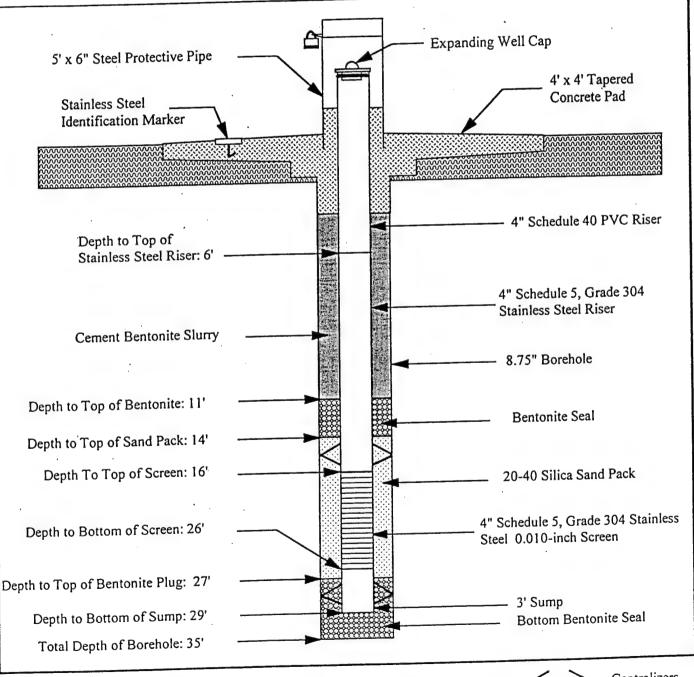
Client: Tinker Air Force Base	Well ID: 2-142B
Location: TAFB, Oklahoma	Site: BW
Contract No: F34650-94-D-0082/5004 (P & A 3)	Date 4-Inch Riser Set: 05/19/95
Contractor: Brown & Root Environmental	Drilling Method: Mud Rotary
Project Manager: David Parker	Ground Level Elev. (AMSL): 1242.45'
Project Geologist: James W. Roberts	Top of Csg. Elev. (AMSL): 1242.09'
Drlg Contractor: Associated Environmental, Inc.	Dedicated Pump: 2-inch Grundfos
Comments: Top Riser: 4" Schedule 40 PVC Casing.	



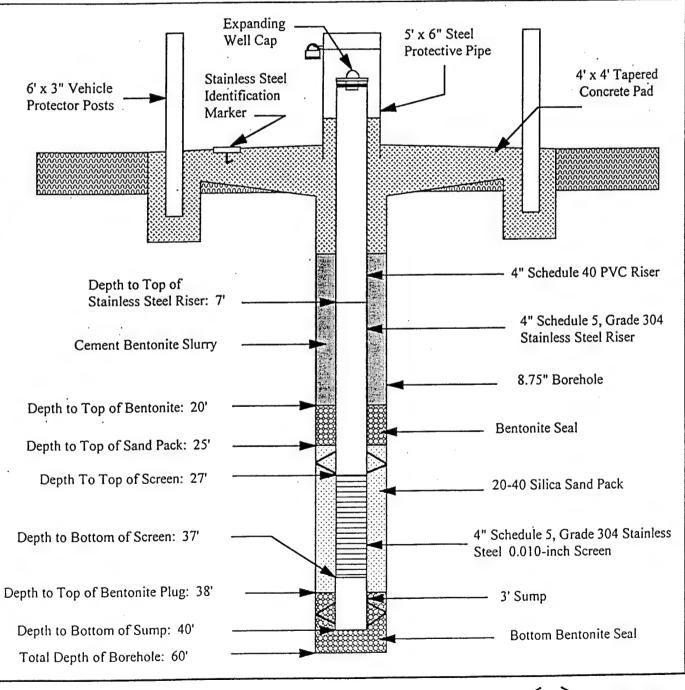


Client: Tinker Air Force Base	Well ID: 2-301B
Location: TAFB, Oklahoma	Site: LF15
Contract No: F34650-94-D-0082/5017 (P&A #4)	Date 4-Inch Riser Set: 06/17/96
Contractor: Brown & Root Environmental	Ground Level Elevation (AMSL): 1,263.19'
Project Manager: David Parker	Top of Casing Elevation (AMSL): 1,265.71'
Project Geologist: Steve Kelly	Northing Coordinate: 150,500.39'
Drilling Contractor: Associated Environmental, Inc.	Easting Coordinate: 2,176,072.65'
Drilling Method: Hollow Stem Auger	Permanent Monuments used in Survey: SE40, PR15
Dedicated Pump: 2-inch Grundfos	Legal Description: SW/4 Section 23, T11N, R2W

Comments: 3' thick top bentonite seal set.



Client: Tinker Air Force Base	Well ID: 2-302B
Location: TAFB, Oklahoma	Site: LF15
Contract No: F34650-94-D-0082/5017 (P&A #4)	Date 4-Inch Riser Set: 07/24/96
Contractor: Brown & Root Environmental	Ground Level Elevation (AMSL): 1,246.46'
Project Manager: David Parker	Top of Casing Elevation (AMSL): 1,249.15'
Project Geologist: Steve Kelly	Northing Coordinate: 149,491.80'
Drilling Contractor: Associated Environmental, Inc.	Easting Coordinate: 2,182,106.79'
Drilling Method: Hollow Stem Auger	Permanent Monuments used in Survey: SE32, SE28
Dedicated Pump: 2-inch Grundfos	Legal Description: SE/4 Section 22, T11N, R2W
Comments:	



WEED CONSTI	UCTION SCHEMATIC
Client: Tinker Air Force Base	Well ID: 2-160B
Location: TAFB, Oklahoma	Site: ST07
Contract No: F34650-94-D-0082/5004 (P & A 3)	Date 4-Inch Riser Set: 06/05/95
Contractor: Brown & Root Environmental	Drilling Method: Auger
Project Manager: David Parker	Ground Level Elev. (AMSL): 1250.99'
Project Geologist: James W. Roberts	Top of Csg. Elev. (AMSL): 1253.56'
Project Geologist: James W. Roberts	Dedicated Pump: 2-inch Grundfos
Drlg Contractor: Associated Environmental, Inc. Comments: Top Riser: 4" Schedule 40 PVC Casing. 6" the	nick top bentonite seal set. Sand set 1' above screen.
Comments: Top Riser: 4 Schedule 40176 Gustage 1	Expanding Well Cap
5' x 6" Steel Protective Pipe Stainless Steel Identification Marker	
	4' x 4' Tapered
	Concrete Pad 8.75" Borehole
Cement Bentonite Slurry	4" Schedule 5, Grade 304 Stainless Steel Riser
Depth to Top of Bentonite: 3.5' Depth to Top of Sand Pack: 4'	Bentonite Seal .
Depth To Top of Screen: 5'	20-40 Silica Sand Pack
Depth to Bottom of Screen: 15'	4" Schedule 5, Grade 304 Stainless Steel 0.010-inch Screen
Depth to Top of Bentonite Plug: 16'	3.0' Sump
Depth to Bottom of Sump: 18' Total Depth of Borehole: 45'	Bottom Bentonite Seal

Logged By:

Drilled By:

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL

DRILLING AND SAMPLING INFORMATION

Boring Location: EAST SIDE OF FIRE TRAINING AREA 2

M. WILSON D. MEYER

SURFACE ELEV.(FT): 1246.213 TOTAL DEPTH(FT.): 70

11/22/93 Date Storted: Date Completed: 11/29/93

GEOTECHNOLOGY. INC.

Drill Rig Type: CME-75

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS AND MUD ROTORY WITH 5-5/8" TRICONE ROCK BIT

Sampling Method: 3"x5" CONTINUOUS SAMPLER

Notes: N 150447.890, E 2182152.166

WELL COMPLETION DATA

Elev-Top of Casing(fl.):

1. Surf Casing-1.D.(in.):8

Centralizers-Type: S.Steel 2. Riser Pipe-I.D.(in.): 2

Centralizers-Type: S.Steel

3. Screen Dia.(in.): 2 Depth Interval(ft.):53.8-63.7 Slot Size(in.): .010 Centralizers-Type:

Conc. Pod Size: 4'x4'x6"

Ref. Datum: MSL

Depth(ft.): 25 Type: Carbon Steel

Depths(ft.): 13

Depth(ft.): 53.8 Type: S.Steel

Depths(ft.): 24, 53

Type: S.Steel Wire Wound

Depths(ft.):

4. Filter Pock Type: Silica Sand Depth Interval(ft.): 52.0-65.0

	GEOPHYSICAL LOG (GAMMA)	TYPE	E C	ي ا	FEET		
. DESCRIPTION	10 20 30 40	SAMPLE MPLE NO IN. DI	FID, PPM	USCS GRAPHIC 10G	' I	12" Steel Manhale Cover	Concrete Flush Completion
LAYEY SILT — slightly plastic; stiff; moist; red (2.57R-4/8); roots and gravel at surface; trace caliche nodules; many black carbon imprint stains — hard at 1'		1561 36 60		Ē	5-	1.0	cement/ bentonite grout
· very hard; dry		1562 5428 24 1563 36	7		10-		8 inch carbon sto cosing
- 2 inch layer of anhydrite/gypsum; light gray (5YR-7/1) ANDY SILT - 30% very fine grained sand, 70% silt; stiff; compact; becoming more sandy ILTY SAND - 30% silt, 70% very fine grained sand; compact; moist; red (2.5YR-4/8); few layers (.5 inch) of moderately well cemented		1564	2.5	E SE	15-		stoinless steel centralizer
AND — very fine to fine grained; poorly graded; compact; moist; red (2.5YR-4/8); gradational contact; sand forms thin layers		1565	2.5	sp	20—		12 inch borehole
ANDY SILT — 20% very fine grained sand, 80% silt; hard; dry; red (2.5YR-4/8); sharp contact;		1365/6		Ē	25	_25.0	stainless steel centralizer
many thin layers of light gray clay ILT SAND - 40% silt, 60% very fine grained sand; hard; compact; wet; red (2.5YR-4/8); sharp contact; thin layers of light gray clay small patches of black stains		24 24	7 6	sm	30-	-	riser pipe (threaded)
							cement/ bentonite grout
AND - poorly graded; <5% fines				sp .	35-		6 inch borehole
DRAFT RPS	DRAFT	PROJ.		APP		DWG.	409832-A57

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-62A

WELL COMPLETION DATA

DRILLING AND SAMPLING INFORMATION Boring Location: EAST SIDE OF

FIRE TRAINING AREA 2

SURFACE ELEV.(FT): 1246.213

TOTAL DEPTH(FT.): 70

11/22/93

Logged By: Drilled By:

M. WILSON D. MEYER

Date Started: Date Completed: 11/29/93

GEOTECHNOLOGY. INC.

Drill Rig Type: CME-75

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS AND MUD ROTORY WITH 5-5/8" TRICONE ROCK BIT

Sampling Method: 3"x5' CONTINUOUS SAMPLER

Notes: N 150447.890, E 2182152.166

Elev-Top of Casing(fL):

1. Surf Casing-I.D.(in.):8 Centralizers-Type: S.Steel Depths(ft.): 13

2. Riser Pipe-LD.(in.): 2

Centrolizers-Type: S.Steel

3. Screen Dia.(in.): 2 Depth Intervol(ft.):53.8-63.7 Slot Size(in.): .010

Centralizers-Type:

Conc. Pad Size: 4'x4'x6"

Ref. Datum: MSL

Type: Carbon Steel Depth(ft.): 25

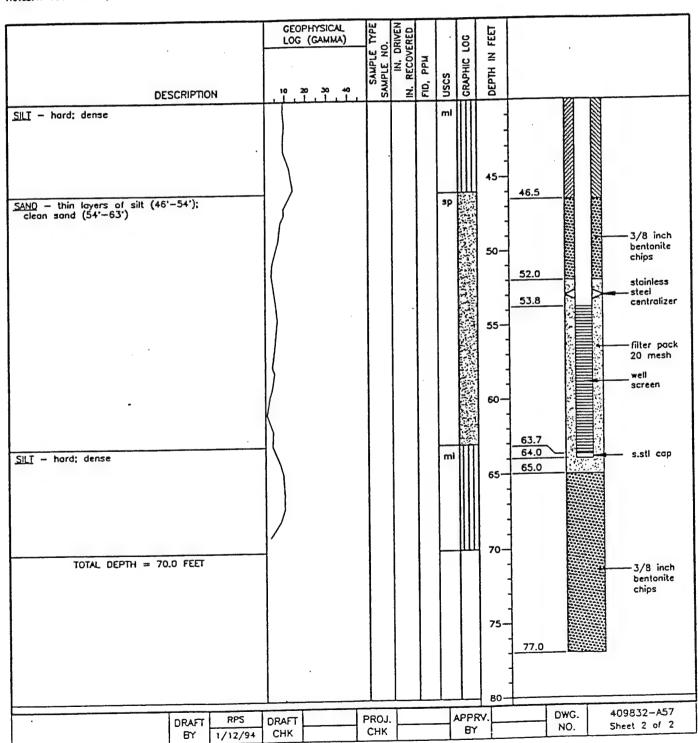
Depth(ft.): 53.8 Type: S.Steel

Depths(ft.): 24, 53

Type: S.Steel Wire Wound

Depths(ft.):

4. Filter Pack Type: Silica Sand Depth Interval(ft.): 52.0-65.0



Logged By:

Drilled By:

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-63A

DRILLING AND SAMPLING INFORMATION

Boring Location: EAST SIDE OF FIRE TRAINING AREA 2

SURFACE ELEV.(FT): 1243.387

TOTAL DEPTH(FT.): 67 Date Started:

M. WILSON D. MEYER

Date Completed:

11/19/93 11/23/93

GEOTECHNOLOGY, INC.

Drill Rig Type: CME-75

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS AND MUD ROTORY WITH 5-5/8" TRICONE ROCK BIT

Sampling Method: 3"x5' CONTINUOUS SAMPLER

Notes: N 150389.964, E 2182043.019

AND 1-1/2"x2" SPLIT SPOON

WELL COMPLETION DATA

Elev-Top of Cosing(ft.):

1. Surf Casing-I.D.(in.):8

Centrolizers-Type: S.Steel 2. Riser Pipe-1.D.(in.): 2

Centralizers-Type: S.Steel 3. Screen Dia.(in.): Depth Interval(ft.):53.0-63.0 Slot Size(in.): .010 Centralizers-Type:

4. Filter Pock Type: Silica Sand Depth Interval(ft.):51-64 Conc. Pod Size: 4'x4'x6"

Ref. Datum: MSL

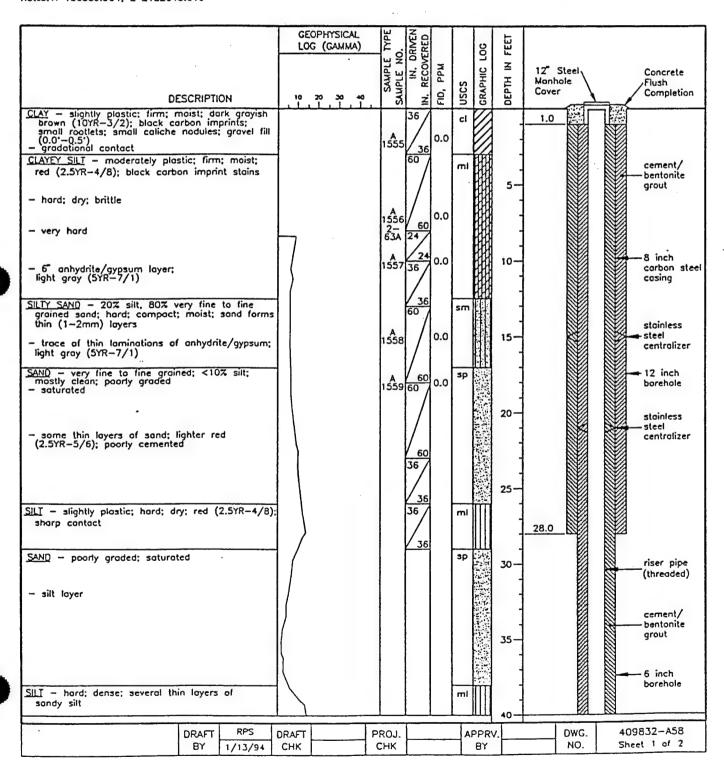
Depth(ft.): 28 Type: Corbon Steel

Depths(ft.): 15

Depth(ft.): 53 Type: S.Steel

Depths(ft.): 21, 51 Type: S.Steel Wire Wound

Depths(ft.):





Logged By:

Drilled By:

Project Name: TINKER 5001

Project Location: TINKER AFB. OKLAHOMA

Project Number: 409832

MONITORING WELL 2-63A

Boring Location: EAST SIDE OF

DRILLING AND SAMPLING INFORMATION

FIRE TRAINING AREA 2

M. WILSON D. MEYER

SURFACE ELEV.(FT): 1243.387 TOTAL DEPTH(FT.): 67

Date Started: Date Completed:

11/19/93 11/23/93

GEOTECHNOLOGY. INC.

Drill Rig Type: **CME-75**

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS AND MUD ROTORY WITH 5-5/8" TRICONE ROCK BIT

Sampling Method: 3"x5" CONTINUOUS SAMPLER

AND 1-1/2"x2" SPLIT SPOON

Notes: N 150389.964, E 2182043.019

WELL COMPLETION DATA

Elev-Top of Casing(ft.):

1. Surf Casing-I.D.(in.):8 Centralizers-Type: S.Steel

2. Riser Pipe-L.D.(in.): 2

Centralizers-Type: S.Steel
Screen Dia.(in.): 2 3. Screen Dia.(in.): Depth Interval(ft.):53.0-63.0 Slot Size(in.): .010

Conc. Pad Size: 4'x4'x6"

Centrolizers-Type:

Ref. Datum: MSL

Type: Carbon Steel Depth(ft.): 28

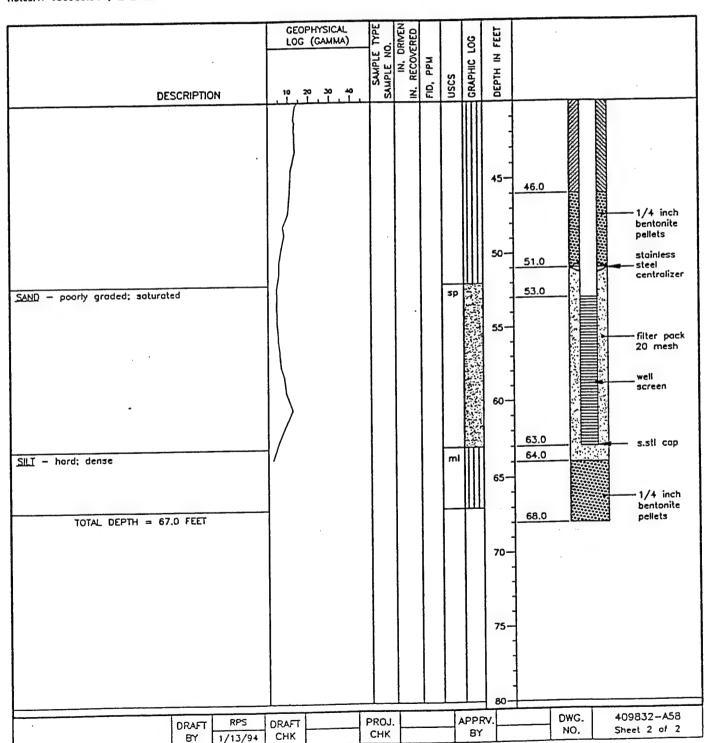
Depths(ft.): 15

Type: S.Steel Depth(ft.): 53

Depths(ft.): 21, 51 Type: S.Steel Wire Wound

Depths(ft.):

4. Filter Pack Type: Silica Sand Depth Interval(ft.): 51-64



Logged By:

Drilled By:

Project Name: TINKER 5001

SURFACE ELEV.(FT): 1246.052

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-64A

WELL COMPLETION DATA

Elev-Top of Casing(ft.):

1. Surf Casing-I.D.(in.):8 Centrolizers-Type: S.Steel

2. Riser Pipe-I.D.(in.): 2 Centralizers-Type: S.Steel

3. Screen Dia.(in.): 2 Depth Interval(ft.):56-66 Centrolizers-Type:

Conc. Pod Size: 4'x4'x6"

Ref. Datum: MSL

Depth(ft.): 30 Type: Carbon Steel

Depths(ft.): 16

Type: S.Steel Depth(ft.): 56

Depths(ft.): 30, 54

Type: S.Steel Wire Wound Slot Size(in.): ,010

Depths(ft.):

4. Filter Pock Type: Silica Sand Depth Interval(ft.): 54.5-67.0

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTH SIDE OF

FIRE TRAINING AREA 2

M. WILSON

D. MEYER

Date Started: Date Completed:

11/18/93 11/22/93

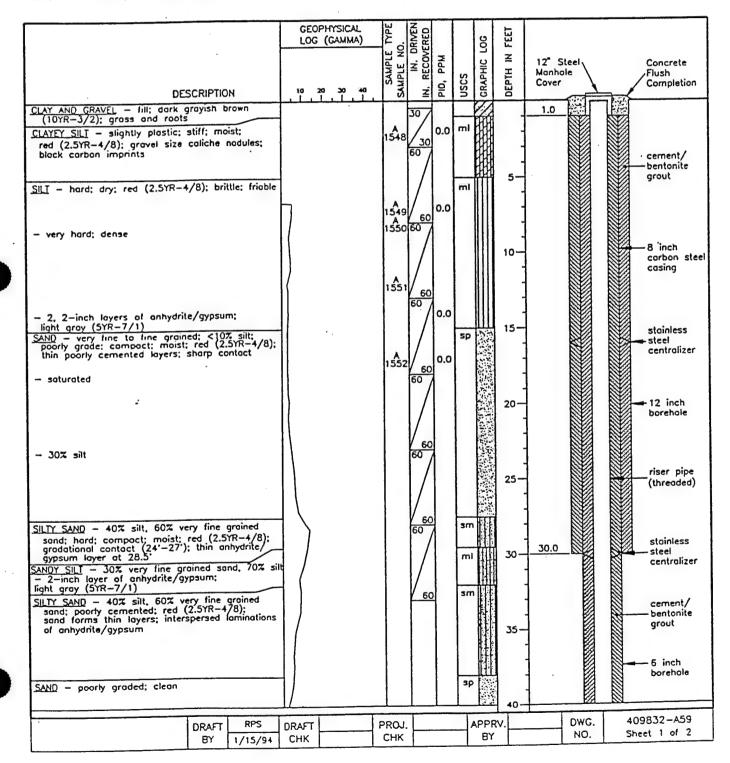
TOTAL DEPTH(FT.): 75

GEOTECHNOLOGY. INC.

Drill Rig Type: **CME-75**

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS AND MUD ROTORY WITH 5-5/8" TRICONE ROCK BIT Sampling Method: 3"x5" CONTINUOUS SAMPLER

Notes: N 150470.205, E 2181959.365





Logged By:

Project Nome: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-64A

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTH SIDE OF FIRE TRAINING AREA 2

SURFACE ELEV.(FT): 1246.052

TOTAL DEPTH(FT.): 75

Date Completed:

Date Started:

11/18/93 11/22/93

D. MEYER Drilled By: GEOTECHNOLOGY. INC.

Drill Rig Type: CME-75

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS AND MUD ROTORY WITH 5-5/8" TRICONE ROCK BIT

Sampling Method: 3"x5" CONTINUOUS SAMPLER

M. WILSON

Notes: N 150470.205, E 2181959.365

WELL COMPLETION DATA

Elev-Top of Casing(ft.):

1. Surf Casing-I.D.(in.):8 Centralizers-Type: S.Steel Depths(ft.): 16

2. Riser Pipe-I.O.(in.): 2 Centralizers-Type: S.Steel Depths(ft.): 30, 54

3. Screen Dia.(in.): 2 Depth Interval(ft.):56-65

Centralizers-Type:

Conc. Pad Size: 4'x4'x6"

Ref. Datum: MSL

Depth(ft.): 30 Type: Carbon Steel

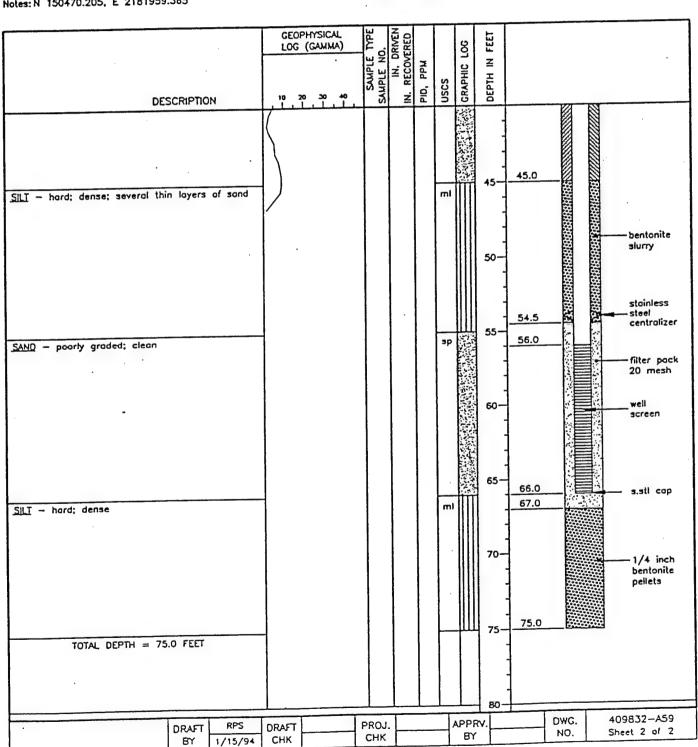
Depth(ft.): 56 Type: S.Steel

Type: S.Steel Wire Wound

Slot Size(in.): .010

Depths(ft.):

4. Filter Pack Type: Silica Sand Depth Interval(ft.): 54.5-67.0



Logged By:

Drilled By:

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA Project Number: 409832

MONITORING WELL

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTHEAST OF FIRE TRAINING AREA 2

Drill Rig Type: CME-75

SURFACE ELEV.(FT): 1250.976

TOTAL DEPTH(FT.): 79

D. MEYER GEOTECHNOLOGY. INC.

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS

M. WILSON

Date Started: Date Completed:

11/15/93 Centralizers-Type: 11/19/93 2. Riser Pipe-1.D.(in.): 2

> Centrolizers-Type: S.Steel 3. Screen Dia.(in.): Depth Interval(ft.):66-76

Elev-Top of Cosing(ft.):

1. Surf Casing-I.D.(in.):8

Conc. Pod Size: 4'x4'x6"

Centrolizers-Type:

Ref. Datum: MSL

Depth(ft.): 55 Type: Carbon Steel

Depths(ft.):

WELL COMPLETION DATA

Depth(ft.): 66 Type: S.Steel

Depths(ft.): 61

Type: S.Steel Wire Wound Slot Size(in.): .010

Depths(ft.):

4. Filter Pock Type: Silica Sand Depth Interval(ft.): 64-77

Sampling Method: 3"x5" CONTINUOUS SAMPLER AND 1-1/2"x2" SPLIT SPOON

Notes: N 150698.281, E 2182189.884

			_						
	GEOPHYSICAL LOG (GAMMA)	TYPE	IN. DRIVEN RECOVERED			0	FEET		
	200 (0.31112.)	유	문 등			100	<u>π</u>	_	
·		SAMPLE MPLE NO	≆ છે	РРМ		H		12" Steel	Concrete Flush
DESCRIPTION	10 20 30 40	SAMPLE	Z.	6	nscs	GRAPHIC	ОЕРТН	Cover	Completion
	<u> Jahalala</u>	_	± 48 /	-	_	77	-	- OF	734
CLAY — moderately plastic; stiff; moist; grayish brown (5YR-4/2); grass and roots					cl		+	1.0	
— gradational contact		1540			_		+		
GRAYELLY CLAY — slightly plastic; 20% granule size gravel; firm; moist; red (2.5YR-4/8); rootlets;		1 1	48	0.0	cl		1		cement/
block carbon imprint - hard; dry; brittle			60				. 1		bentonite
- 50% light red caliche - 10% gravel	•	1541	_ /				5-		grout
- 10% graves	_	1541	/	0.0			- 4		
			/				4		
- 30%-40% sand		1 1	60 A		- 1		- 1		
CLAYEY SILT - nonplastic; very hard; brittle; dry;			°° /	ı	ml		10-		8 inch carbon stee
red (2.5YR-4/8); shorp contact	}		-/1						casing
		1542	/	0.0			1		
- dense; compoct		l	60	0.0		础	- 1		
SANDY SILT - 25% very fine grained sand; 75%	1		60		ml		1		
SANDY SILT - 25% very fine grained sand; 75% silt; loose; dry; red (2.5YR-4/8); thin silt layers (0.1-0.5 thick)			_/	ا م			15-		
	1	1543	/	0.0]		12 inch
			/]		borehole
			36	. [- 1		
— few thin seams of very fine grained sand; gray	1	1544	30/				20		
		.54.	/30				- 1		
- hard		[
	1						1		
							1		riser pipe
SILTY SAND - 30% silt, 70% very fine grained					sm		25-		(threaded)
sand; compact; stiff; dry; red (2.5YR-4/8); hard to drill		1545	9/9]		
]		
- saturated		1546					4		
- Soldrated		1546				111	30-		
							-		
SAND dies seriests 410° Sant seets Gradets			24/		3p	Щ	+		
SAND — fine grained; <10% fines; poorly graded; compact; saturated; red (2.5YR-4/8)					-	12.0	-		cement/
SILTY SAND - 30% silt, 70% very fine grained			60		sm		1		bentonite grout
sand; compact; saturated; red (2.5YR-4/8)							35-		,
CLAYEY GRAVELLY SAND - 30% clay, 40% gravel, 30% line grained sand; loose; saturated;					3¢-		1		
red (2.5YR-4/8)			/			\mathscr{L}			
SAND — fine grained; poorty graded; compact; saturated; red (2.5YR-4/8)	1		60		зÞ		-		
Suturated; 180 (2.51N-7/0)	<u> </u>			\Box			40		
· DRAFT RPS	DRAFT P	ROJ.			A	PPR	/	DWG.	409832-A60
BY 1/15/94	CHK	CHK				BY		NO.	Sheet 1 of 2



Project Nome: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

MONITORING WELL 2-65A

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTHEAST OF FIRE TRAINING AREA 2

SURFACE ELEV.(FT): 1250.976

TOTAL DEPTH(FT.): 79 · Date Started:

M. WILSON Logged By: D. MEYER Drilled By:

Date Completed:

11/15/93 11/19/93

GEOTECHNOLOGY. INC.

CME-75 Drill Rig Type:

Drilling Method: 8" AND 12" HOLLOW STEM AUGERS

Sampling Method: 3"x5' CONTINUOUS SAMPLER AND 1-1/2" x2" SPLIT SPOON

Notes: N 150698.281, E 2182189.884

WELL COMPLETION DATA

Elev-Top of Cosing(ft.):

1. Surf Casing-LD.(in.):8 Centralizers-Type:

2. Riser Pipe-I.D.(in.): 2 Centralizers-Type: S.Steel

3. Screen Dia.(in.): Depth Interval(ft.):66-76 Centralizers-Type:

Conc. Pad Size: 4'x4'x6"

Ref. Dotum: MSL

Depth(ft.): 55 Type: Carbon Steel

Depths(ft.):

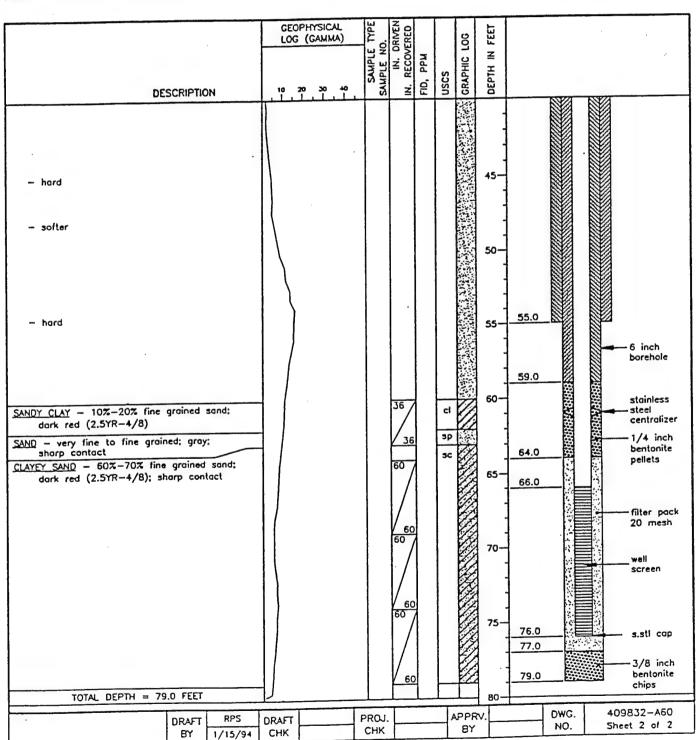
Depth(ft.): 66 Type: S.Steel

Depths(ft.): 61

Type: S.Steel Wire Wound Slot Size(in.): .010

Depths(ft.):

4. Filter Pack Type: Silica Sand Depth Interval(ft.): 64-77



Client: Tinker Air Force Base	Well ID: 2-142A
Location: TAFB, Oklahoma	Site: BW
Contract No: F34650-94-D-0082/5004 (P & A	3) Date 4-Inch Riser Set: 05/19/95
Contractor: Brown & Root Environmental	Drilling Method: Mud Rotary
Project Manager: David Parker	Ground Level Elev. (AMSL): 1242.30'
Project Geologist: James W. Roberts	Top of Csg. Elev. (AMSL): 1241.76'
Drlg Contractor: Associated Environmental, Inc.	Dedicated Pump: 2-inch Grundfos
Comments: Top Riser: 4" Schedule 40 PVC Cas	sing.
Stainless Steel Identification Marker	Expanding well cap
	4' x 4' Tapered Concrete Pad
Cement Bentonite Slurry	4" Schedule 5, Grade 304 Stainless Steel Riser
•	8.75" Borehole
Depth to Top of Bentonite: 65' Depth to Top of Sand Pack: 70'	Bentonite Seal
Depth To Top of Screen: 72'	20-40 Silica Sand Pack
Depth to Bottom of Screen: 82'	4" Schedule 5, Grade 304 Stainles Steel 0.010-inch Screen
h to Top of Bentonite Plug: 83'	3.0' Sump
Depth to Bottom of Sump: 85'	Bottom Bentonite Seal

WEDD COLLE	
	Well ID: 2-112A
Client: Tinker Air Force Base	Site: LF15
Location: TAFB, Oklahoma	Date 4-Inch Riser Set: 1/23/95
Contract No: F34650-94-D-0082/5003 (P & A 2)	Drilling Method: Auger
Contractor: Brown & Root Environmental	Ground Level Elev. (AMSL): 1248.03'
Project Manager: David Parker	Top of Csg. Elev. (AMSL): 1251.41'
Project Geologist: James W. Roberts	Dedicated Pump: 2-inch Grundfos
· · · · · · · · · · · · · · · · · · ·	een bottom. Top Riser: 4" Schedule 40 PVC Casing.

Comments: Sand pack set 1' above screen and even with screen bottom. Top Riser: 4" Schedule 40 PVC Casing. Expanding Well Cap 5' x 6" Steel Protective Pipe 5' x 4" Vehicle Protector Posts Stainless Steel Identification Marker 4' x 4' Tapered Concrete Pad 15" Borehole 10" Schedule 40 PVC Surface Casing Depth to Bottom of 4" Schedule / Type 304 Surface Casing: 60' Stainless Steel Riser Cement Bentonite Slurry 8.75" Borehole Depth to Top of Bentonite: 80' - Bentonite Seal Depth to Top of Sand Pack: 85' - 20-40 Silica Sand Pack Depth To Top of Screen: 86' 4" Schedule 5, Grade 304 Stainless Steel 0.010-inch Screen Depth to Bottom of Screen: 96' Depth to Top of Bentonite Plug: 96' 3.0' Sump Bottom Bentonite Seal Depth to Bottom of Sump: 99'

Centralizers

Total Depth of Borehole: 104'

Project Name: TINKER 5001

Project Location: TINKER AFB, OKLAHOMA

Project Number: 409832

SOIL BORING 2-65P

DRILLING AND SAMPLING INFORMATION

Boring Location: NORTH OF FIRE TRAINING SURFACE ELEV.(FT):

AREA

TOTAL DEPTH(FT.): 100.0

K. KIRSCHENMANN

Date Storted:

11/2/93

Logged By: Drilled By:

P. GUERREIN

Date Completed: 11/3/93

GEOTECHNOLOGY, INC.

Drill Rig Type: CME-75

Drilling Method: 8" HOLLOW STEM AUGER AND

MUD ROTARY WITH 3-7/8" BIT

Sampling Method:

Notes: STRATIGRAPHIC TEST USED FOR GEOPHYSICAL LOGGING

DESCRIPTION	GEOPHYSICAL LOG (GAMMA)	SAMPLE TYPE SAMPLE NO. IN. DRIVEN	PIO, PPM USCS GRAPHIC LOG DEPTH IN FEET
SILTY CLAY			cl -
– very hard			90-
TOTAL DEPTH = 100.0 FEET			100-
			110-
DRAFT	_ APPRVBY	DWG.	115— 120— 409832—A27 Sheet 3 of 3

Analytical Results for Fire Training Area 2 For USZ Groundwater Tinker Air Force Base, Oklahoma Table 5-3

Parameters A1601 A1602 A1603 A1603 A1604 A1604	Well/Boring:	2-62B	2-62B	2-62B	2-63B	2-64B	2-65B	
Depth in Feet: 0-0	Sample ID:		A1602	A1665	A1600	A1603	A1604	
Result QFR Result QFR Result QFR Result QFR Result QFR Result QFR Result QFR Result QFR Result QFR Result QFR Result QFR Result GFR Result GFR Result GFR Result GFR Result GFR Result GFR GFR Result GFR Result GFR	Depth in Feet.		0-0	0-0	0-0	0-0	0-0	
eous (mg/L) 520 330 470 380 and 35 89 34 34 229 2.7 3.8 3.8 3.8 3.8 4.3 110 230 660 660 510 450 0.38 3 3 3 2.2 11.2 150 150 150				Result OFR	Result OFR	Result OFR		
520 330 470 380 35 89 39 34 2,9 2,7 3,8 3,8 7,3 6,9 8,9 4,3 110 230 36 24 0,15 660 510 450 0,38 3 2,2 1,2 190 250 750 160	ellaneous (mg/L)							
nd 35 89 39 34 87 89 3.8 3.8 7.3 6.9 8.9 4.3 110 230 36 24 0.15 660 510 450 0.38 3 2.2 1.2 190 250 750 160	ric	520	330		470	380	200	
87 89 39 34 2.9 2.7 3.8 3.8 7.3 6.9 8.9 4.3 110 230 36 24 0.15 660 510 450 0.38 3 2.2 1.2 190 250 750 160	Semand	35	•					
2.9 2.7 3.8 3.8 7.3 6.9 8.9 4.3 110 230 36 24 0.15 660 510 450 0.38 3 2.2 1.2 190 250 750 160	rom.	87	89		39	34	57	
7.3 6.9 8.9 4.3 110 230 36 24 0.15 650 510 450 0.38 3 2.2 1.2 190 250 750 160		2.9	2.7		3.8	3.8	2.6	
110 230 36 24 0.15 630 660 510 450 0.38 3 2.2 1.2 190 250 750 160		7.3	6.9		8.9	4.3	. 11	
0.15 650 510 450 0.38 3 2.2 1.2 190 250 750 160	m.	110	230		36	₹.	37	
630 660 510 450 0.38 3 2.2 1.2 190 250 750 160		0.15						
0.38 3 2.2 1.2 3 3 3 7.50 1.60	ids	. 630	099		510	420	650	
3 3 2.2 1.2 190 250 750 160	ogen	0.38						
190 250 750 160	uox	3	e		2.2	1.2	1.6	
	olids	180	250		750	160	42	

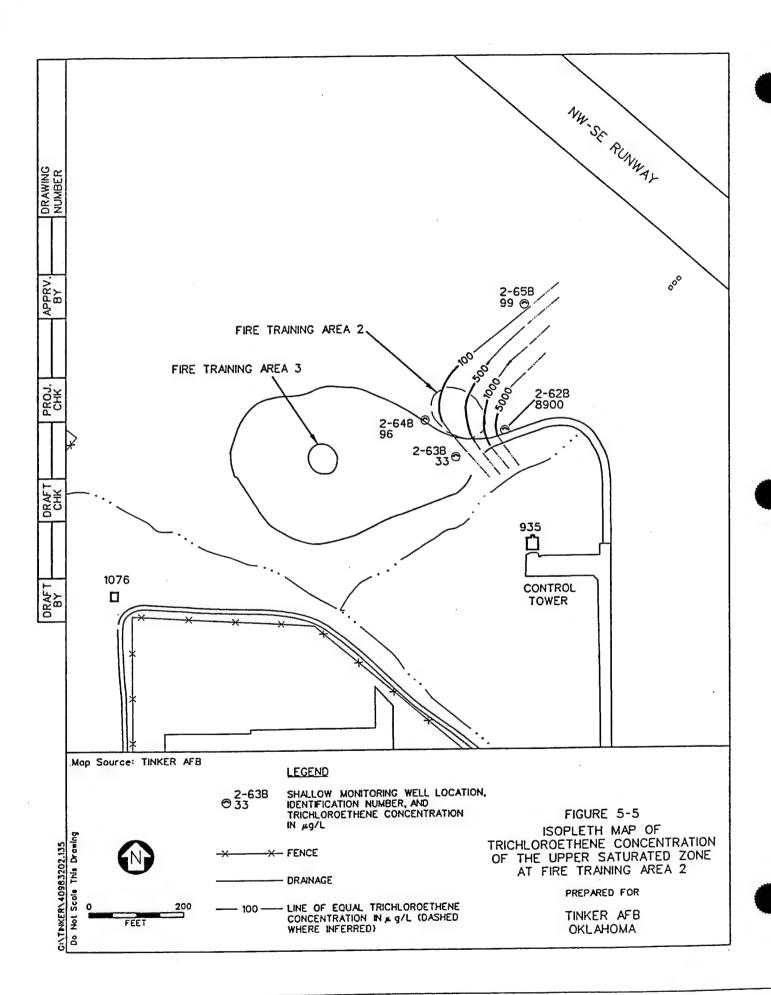
= Compound identified at a secondary dilution factor.

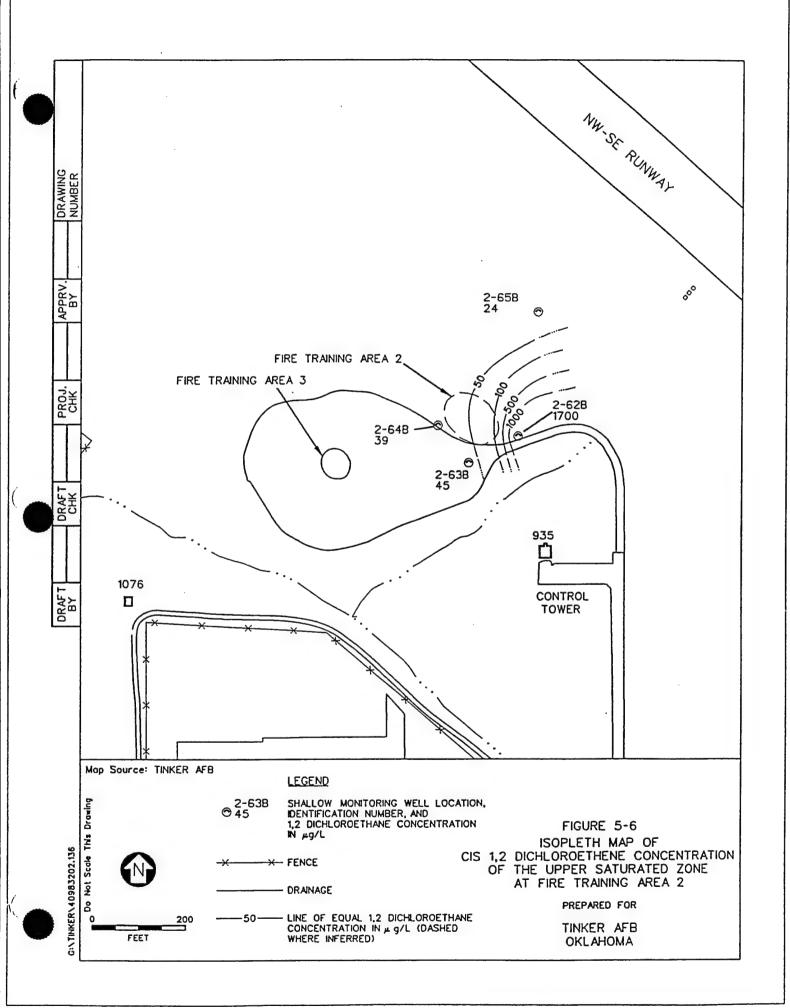
Concentration exceeds instrument calibration range for that specific analysis
 Concentration is an estimated value
 Sample is outside of Matrix Spike QC Limit
 Not detected

QFR = Qualifier Analytical data has not been validated

Analytical Results for Fire Training Area 2 For LSZ Groundwater Tinker Air Force Base, Oklahoma Table 5-4

A1608 A160 0 0 0 2.9 N 2.8 0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 38 258 200 190 26 24 1.1 0.78 8.6 11	A1608 A1609 Result O-0 0.0 0-0 0.056 61 61 36 0.075 N 0.053 4.8 4.8 4.3 0.0042 3.5 0.007 N 0.007 0.007 N 0.0028 23 22 22 23 22 22 24 11 14 N 24 3.9	A1608 A1609 Co 0	A1608 A1608	A1608 A1608 A1608 A1608 0	A1608 A1608 A1608 A1609 0 - 0	A1608 A1608
0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.0	2.9 N 2.8 0.56 4.8 0.075 N 2.8 0.056 4.8 0.007 N 0.0053 3.5 0.007 0.007 0.007 0.007 0.007 0.007 190 24 111 14 N 24 3.9	FR Result OFR Result 0-0 0-0 0-0 0-0 0-0 0-0 0-0 0-0 0-0 0-	OFR Result OFR Result 0.56 4.8 4.8 4.8 4.3 0.0042 4.0 0.007 4.0 0.008 0.007 0.007 0.008 0.007 0.008 0.008 0.009	0.0 OFR Result 2.9 N 2.8 61 36 61 36 62 4.3 63 4.3 64 2 64 3 65 0.067 67 N 0.028 88 228 88 228 60 190 60 190 60 190 61 11 0.78 61 14 N 24 63 3.9	OFR Result OFR Result 0.00 4.8 61 36 4.8 4.3 4.3 0.0042 19 4.8 4.3 0.0042 19 2.3 0.007 N 0.007 4.8 4.3 2.3 0.007 N 0.007 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.8 4.3 4.9 6.007 4.8 6.6 119 4.8 7.8 6.6 4.8 6.6 111 4.8 7.8 6.6	OFR Result OFR Result 0.56 4.8 4.8 4.3 0.0042 4.8 4.3 0.0075 N 0.0053 4.8 4.3 0.0077 N 0.0053 4.8 2.4 2.4 2.4 2.4 3.9
2.9 N 2.8 0.56 4.8 0.075 N 0.053 4.8 0.0042 35 0.012 N 0.0057 0.057 N 0.0058 38 228 0.021 N 0.028 388 200 200 200 200 201 14 N 24 111 0.78 3.9	2.9 N 2.8 0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.007 0.067 N 0.028 23 22 0.021 N 0.028 200 190 26 24 1.1 0.78 8.6 111 14 N 24	FR Result OFR Result 2.9 N 2.8 0.56 61 36 0.005 N 0.053 4.8 4.3 0.0042 35 0.007 N 0.007 0.067 N 0.028 23 23 22 0.021 N 0.028 28 28 28 28 38 28 38 38 38	FR Result OFR Result 2.9 N 2.8 0.56 61 36 0.0075 N 0.0053 4.8 4.3 0.0042 35 19 0.12 N 0.007 0.067 N 0.028 22 0.021 N 0.028 388 228 280 24 1.1 0.78 8.6 111 14 N 24	FR Result OFR Result 2.9 N 2.8 0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 35 0.007 N 0.007 0.067 N 0.028 23 22 0.021 N 0.028 388 228 389 24 111 0.78 8.6 111 14 N 24	FR Result OFR Result 2.9 N 2.8 0.56 61 36 0.0075 N 0.053 4.8 4.3 0.0042 35 0.021 N 0.007 0.067 N 0.028 23 22 0.021 N 0.028 24 1.1 24 N 24 3.9	FR Result OFR Result 2.9 N 2.8 0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 35 0.021 N 0.007 0.067 N 0.028 388 258 200 200 20 24 1.1 0.78 8.6 11 14 N 24
N 2.8 N 0.053 N 0.067 N 0.067 N 0.028 N 0.028 N 0.028 N 0.028	36 N 0.0053 N 0.0057 N 0.0067 N 0.028 N 0.028 N 0.026 N 0.028 110 110 24 0.26 124 127 128 129 124 124 124 124 125 126 127 127 127 127 127 127 127 127 127 127	2.9 N 2.8 0.56 0.56 4.8 4.3 0.0042 35 0.12 N 0.057 0.067 N 0.028 23 22 0.021 N 0.028 388 228 360 190 26 24 1.1 0.78 8.6 111 14 N 24	2.9 N 2.8 0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 35 0.12 N 0.067 0.067 N 0.028 22 0.021 N 0.028 388 228 388 228 348 111 0.78 8.6 111 14 N 224 3.9	2.9 N 2.8 0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 0.028 388 228 388 258 360 190 26 24 1.1 0.78 8.6 111 14 N 24	2.9 N 2.8 0.56 0.56 4.8 4.3 0.0042 35 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 228 380 284 1.1 0.78 8.6 111 14 N 244	2.9 N 2.8 0.56 0.56 4.8 4.3 0.0042 35 0.12 N 0.057 0.067 N 0.028 388 228 388 258 350 190 26 24 1.1 0.78 8.6 111 14 N 24
N 2.8 N 0.053 N 0.067 N 0.067 N 0.028 N 0.028 N 0.028 N 0.028	N 2.8 N 0.053 N 0.0053 N 0.0067 N 0.028 N 0.028 N 0.026 N 0.026 110 124 124 N 124 N	2.9 N 2.8 0.56 0.56 4.8 4.3 0.0042 35 0.12 N 0.057 0.067 N 0.028 23 22 0.021 N 0.028 388 228 380 288 244.3 0.067 N 0.028 380 284 1.1 0.78 8.6 111 14 N 2.4	2.9 N 2.8 0.56 0.56 4.8 4.3 0.0042 35 0.12 N 0.057 0.067 N 0.028 388 228 388 258 350 190 26 24 1.1 0.78 8.6 11 4 N 24	2.9 N 2.8 0.56 0.56 4.8 4.3 0.0042 35 0.12 N 0.057 0.067 N 0.028 388 228 200 350 190 26 24 1.1 0.78 8.6 111 14 N 24	2.9 N 2.8 0.56 0.56 4.8 4.3 0.0042 35 0.12 N 0.057 0.067 N 0.028 388 228 388 258 350 190 26 24 1.1 0.78 8.6 111 14 N 24	2.9 N 2.8 0.56 0.56 0.075 N 0.053 4.8 4.3 0.0042 35 0.012 N 0.057 0.067 N 0.028 388 228 388 258 350 190 26 24 1.1 0.78 8.6 111 14 N 24
36 N 0.0053 N 0.0067 N 0.028 N 0.028 N 0.028 N 0.028 N 0.028	36 N 0.053 N 0.067 N 0.028 N 0.028 N 0.028 N 0.028 N 0.028 110 111 N 111 N 1 N 1	0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 258 380 224 1.11 0.78 8.6 111 14 N 24	0.56 61 36 0.075 N 0.053 4.8 4.8 4.3 0.0042 35 0.12 N 0.067 0.067 N 23 23 22 0.021 N 0.028 388 228 388 228 380 110 24 1.1 14 N 24 3.9	0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 N 0.067 0.067 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 38 258 38 258 36 24 1.1 0.78 8.6 111 14 N 24	0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 N 0.067 0.067 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 38 258 38 258 36 24 1.1 0.78 8.6 111 14 N 24	0.56 61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 0.028 38 228 38 228 38 228 36 19 26 24 1.1 0.78 8.6 11 14 N 24
36 N 0.053 N 0.067 N 0.028 N 0.028 N 0.028 N 0.026 N 228 N 1190 258 N 0.26	36 N 0.053 H 0.067 N 0.028 N 0.028 N 0.028 N 228 0.26 1190 N 24 111 0.78	61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 0.028 38 228 388 228 350 190 26 24 1.1 0.78 8.6 111	61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 22 0.051 N 0.028 388 228 350 190 26 24 1.1 0.78 8.6 11	61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 22 0.051 N 0.028 388 228 380 190 26 24 1.1 0.78 8.6 111	61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 22 0.021 N 0.028 388 228 380 190 26 24 1.1 0.78 8.6 111 14 N 24	61 36 0.075 N 0.053 4.8 4.3 0.0042 19 0.12 N 0.067 0.067 N 0.028 388 228 380 190 26 26 1.1 0.78 8.6 111 14 N 24
36 N 0.053 N 0.067 N 0.028 N 0.028 N 0.028 N 0.028 N 0.26 N 2.48 N 2.48	36 N 0.0053 N 0.0067 N 0.028 N 0.028 N 0.028 N 0.026 N 0.26 111 N 24 124 N 124 N 124	4.8 4.3 4.3 4.3 6.0042 19 0.0042 19 0.0042 19 0.007 N 0.007 N 0.007 N 0.005 N	4.8 4.3 4.3 6.0042 19 0.0042 19 0.0042 19 0.0047 N 0.0067 N 0.0067 N 0.0067 N 0.028 19 0.026 200 1350 1350 14 N 24 11 1 0.78 8.6 11 14 N 24	4.8 4.3 4.3 6.0042 19 0.0042 19 0.0042 19 0.0047 N 0.0067 N 0.0067 N 0.0067 N 0.028 19 0.26 200 15 0.20 200 15 0.20 200 15 0.20 200 15 0.20 200 15 0.20 200 15 0.20 200 15 0.20 200 15 0.2	4.8 4.3 4.3 4.3 6.0042 19 0.0042 19 0.0042 19 0.0047 N 0.0067 N 0.0067 N 0.0067 N 0.0028 190 0.26 200 190 24 1.1 0.78 8.6 11 0.78 8.6 11 0.78 8.	4.8 4.3 4.3 4.3 6.0042 19 0.0042 19 0.0042 19 0.007 N 0.007 N 0.007 N 0.007 N 0.002 19 0.26 200 19 0.26 24 1.1 0.78 8.6 11 1 4 N 2.4
N 0.053 N 0.067 N 0.028 N 0.028 N 0.026 N 228 N 248 N 248	N 0.053 4.3 N 0.067 N 0.028 N 0.028 N 0.026 0.26 0.26 1190 24 111 N 124 139	4.8 4.3 6.0042 19 0.0057 N 0.0053 4.8 4.3 6.0042 19 0.0057 N 0.0057 N 0.0057 N 0.0058 228 220 200 200 200 200 200 200 200 20	4.8 4.3 6.0042 19 0.0042 19 0.0047 N 0.0067 0.0067 N 0.0067 N 0.0028 22 0.021 N 0.028 200 200 200 24 1.1 0.78 8.6 11 14 N 24 3.9	4.8 4.3 6.0042 19 0.0042 19 0.0047 N 0.0067 O.0067 N 0.0067 N 0.0067 N 0.028 388 228 0.26 200 130 24 11.1 0.78 8.6 11 14 N 24 3.9	4.8 4.3 6.0042 19 0.0057 N 0.0053 4.8 4.3 6.0042 19 0.0067 N 0.0067 N 0.0067 N 0.0028 388 228 6.26 240 1.1 0.78 8.6 11 14 N 2.4 3.9	4.8 4.3 6.0042 19 0.0033 0.0042 19 0.0042 19 0.007 N 0.0067 N 0.0067 N 0.0028 388 228 0.26 240 1.1 0.78 8.6 11 1 4 N 2.4
4.3 N 0.0067 N 0.028 N 0.028 N 0.26 190 248 N 248	4.3 N 0.067 N 0.028 N 0.028 0.26 0.26 1190 24 0.78 N 24 111 N 24 3.9	4.8 4.3 9.00042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 228 388 228 350 190 26 24 1.1 0.78 8.6 111	4.8 4.3 0.0042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 258 200 350 190 26 24 1.1 0.78 8.6 11 1 0.78	4.8 4.3 35 0.0042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 258 30 190 26 24 1.1 0.78 8.6 11 14 N 24	4.8 4.3 35 0.0042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 228 380 190 26 24 1.1 0.78 8.6 111 14 N 24	4.8 4.3 9.00042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 228 380 190 26 24 1.1 0.78 8.6 111 14 N 24
4.3 N 0.0067 N 0.028 N 0.028 N 0.258 190 248 N 248 N 248	4.3 N 0.0067 N 0.028 N 0.028 0.26 0.26 110 N 24 N 24 N 3.9	4.8 4.3 0.0042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 228 380 190 26 24 1.1 0.78 8.6 111 14 N 24	4.8 4.3 9.00042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 258 200 190 26 24 1.1 0.78 8.6 111 14 N 24	4.8 4.3 35 0.0042 36 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 258 200 190 26 24 1.1 0.78 8.6 111 14 N 24	4.8 4.3 9.0042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 258 200 350 190 26 24 1.1 0.78 8.6 111 14 N 24	4.8 4.3 0.0042 35 19 0.12 N 0.067 0.067 N 0.028 23 22 0.021 N 0.028 388 228 380 190 26 24 1.1 0.78 8.6 111 14 N 24
19 N 0.0067 N 0.028 N 0.028 190 24 24 N 24 N 24 N 24 N 24	19 N 0.067 N 0.028 N 0.028 0.26 0.26 110 N 24 111 N 24 3.9	0.0042 35 0.12 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 200 350 190 26 24 1.1 0.78 8.6 111 14 N 24	0.0042 35 0.12 N 0.067 0.067 N 0.067 23 0.021 N 0.028 388 258 200 200 350 190 24 1.1 0.78 8.6 111 14 N 24	0.0042 35 0.12 N 0.067 0.067 N 0.067 23 0.021 N 0.028 388 258 200 200 350 190 26 24 1.1 0.78 8.6 111 14 N 24	0.0042 35 0.12 N 0.067 0.067 N 0.067 23 0.021 N 0.028 388 258 200 200 350 190 26 24 1.1 0.78 8.6 111 14 N 24	0.0042 35 0.12 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24
19 N 0.067 N 0.028 N 0.028 258 0.26 0.26 190 190 190 111 N 0.067	19 N 0.067 N 0.028 N 0.028 0.26 0.26 110 110 110 110 110 110 110 11	35 19 0.12 N 0.067 0.067 N 23 22 0.021 N 0.028 388 258 200 200 350 190 26 24 1.1 0.78 8.6 111 14 N 24	35 19 0.12 N 0.067 0.067 N 23 22 0.021 N 0.028 388 258 200 350 190 24 1.1 0.78 8.6 111 14 N 24	35 19 0.12 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 350 190 24 1.1 0.78 8.6 111 14 N 24	35 19 0.12 N 0.067 0.067 N 23 22 0.021 N 0.028 388 258 200 350 190 24 1.1 0.78 8.6 111 14 N 24	35 19 0.12 N 0.067 0.067 N 23 22 0.021 N 0.028 388 258 380 250 200 350 190 24 1.1 0.78 8.6 111 14 N 24
N 0.067 N 0.028 N 0.028 258 0.26 0.26 190 190 111 N 24 N 0.78	N 0.067 N 222 N 0.028 0.26 0.26 0.26 110 111 N 24 111 N 3.9	0.067 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 190 26 24 1.1 0.78 8.6 111 14 N 24	0.067 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 190 26 24 1.1 0.78 8.6 111 14 N 24	0.067 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 190 26 24 1.1 0.78 8.6 111 14 N 24	0.067 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 190 26 24 1.1 0.78 8.6 111 14 N 24	0.067 N 0.067 0.067 N 0.067 23 22 0.021 N 0.028 388 258 200 190 26 24 1.1 0.78 8.6 111 14 N 24
N 0.028 N 0.028 258 0.26 190 190 111 N 24 3.9	22 N 0.028 N 258 0.26 0.26 110 N 24 3.9	23 22 0.021 N 0.028 388 258 200 246 200 190 26 24 1.1 0.78 8.6 11 14 N 24	23 22 0.021 N 0.028 388 258 200 246 200 190 26 24 1.1 0.78 8.6 111 14 N 24	23 22 0.021 N 0.028 388 258 200 246 350 190 26 24 1.1 0.78 8.6 111 14 N 24	23 22 0.021 N 0.028 388 258 200 246 200 190 26 24 1.1 0.78 8.6 11 14 N 24	23 22 0.021 N 0.028 388 258 200 200 350 190 26 24 1.1 0.78 8.6 111 14 N 24
22 N 0.028 258 0.26 190 190 111 N 24 3.9	22 N 0.028 258 0.26 0.26 24 0.78 111 N 24 3.9	23 22 0.021 N 0.028 388 258 200 26 200 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 200 350 190 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 200 350 190 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 200 350 190 24 1.1 0.78 8.6 11 14 N 24 3.9
22 N 0.028 258 0.26 190 24 0.78 111 N 24 3.9	22 N 0.028 258 0.26 1190 111 N 24 3.9	23 22 0.021 N 0.028 388 258 200 26 26 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 26 26 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	23 22 0.021 N 0.028 388 258 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9
N 0.028 258 0.26 190 24 0.78 111 N 24 3.9	N 0.028 258 0.26 190 24 0.78 N 24 3.9	0.021 N 0.028 388 258 200 0.26 200 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	0.021 N 0.028 388 258 200 200 350 190 24 1.1 0.78 8.6 11 14 N 24 3.9	0.021 N 0.028 388 258 200 200 350 190 24 1.1 0.78 8.6 11 14 N 24 3.9	0.021 N 0.028 388 258 200 200 350 190 24 1.1 0.78 8.6 11 14 N 24 3.9	0.021 N 0.028 388 258 200 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9
258 0.26 190 24 0.78 111 N 24 3.9	258 0.26 190 24 0.78 N 24 3.9	388 258 0.26 200 1.26 350 190 26 24 1.1 0.78 8.6 11 14 N 24	388 258 0.26 200 190 26 24 1.1 0.78 8.6 11 14 N 24	388 258 0.26 200 190 26 24 1.1 0.78 8.6 11 14 N 24	388 258 0.26 200 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	388 258 0.26 200 190 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9
258 0.26 0.26 24 0.78 111 N 24 3.9	258 0.26 0.26 190 24 0.78 N 24 3.9	388 258 200 0.26 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	388 258 200 0.26 350 190 26 24 1.1 0.78 8.6 11 14 N 24 15 24	388 258 200 0.26 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	388 258 200 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3,9	388 258 200 0.26 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9
0.26 190 190 111 111 10.24 13.9	0.26 190 24 0.78 11 N 24 3.9	0.26 350 190 26 24 1.1 8.6 11 14 N 24 3.9	0.26 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	0.26 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	0.26 200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	0.26 350 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9
190 42 0.78 0.78 111 8,0 8,0	190 24 0.78 111 N 3.9	200 350 100 24 24 1.1 8.6 11 14 N 24 3.9	200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	200 350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9
N 92.0 87.0 111 42 6.0	190 24 0.78 111 24 3.9	350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	350 190 26 24 1.1 0.78 8.6 11 14 N 24 3.9	350 180 26 24 1.1 0.78 8.6 11 14 N 24 3.9	350 180 26 24 1.1 0.78 8.6 11 14 N 24 3.9
0.78 11 42 9.6	0,78 11 42 3,9	1.1 0.78 8.6 11 14 N 24 3.9	1.1 0.78 8.6 11 14 N 24 3.9	1.1 0.78 8.6 11 14 N 24 3.9	1.1 0.78 8.6 11 14 N 24 14 N 24	1.1 0.78 8.6 11 14 N 24 14 N 24 3.9
N 11 24 3,9	11 N 24 3.9	8.6 11 14 N 24 3.9	8.6 14 N 24 3.9	8.6 14 N 24 3.9	8.6 14 N 24 3.9	8.6 14 N 24 3.9
N 24 3,9	N 24	14 N 24	14 N 24 3.9			
		3,9	3.9	3.9	3.9	3.9





APPENDIX B

GEOLOGIC BOREHOLE LOGS, GEOPHYSICAL BOREHOLE LOGS, WELL CONSTRUCTION DIAGRAMS, MONITORING WELL DEVELOPMENT AND SAMPLING RECORDS, SLUG TEST ANALYSES, AND SURVEY DATA

JOB NUMBER: BORING NUMBER: C639-897-41 RIG TYPE: TEMPERATURE (°F):

LOCATION:

COMMENTS:

729691 FAILING FID 90-45°F FTA Z-MPI DRLG MED:

CLIENT: BORING DIA.: CONTRACTOR: WEATHER:

DATE SPUD: (I " 114 | GO" 415B DATE CMPL: AEI Curp. ELEVATION: Suny, Hot Hullow Stem Auger DATUM: GEOLOGIST:

7128197 7128197 1245.61 Crinc1 Surface Brue Henry

					Laboratory	Г		
Depth	Pro-	USCS		Sample	Sample	Sample	PID	Remarks
(ft.)	file	USCS	Geologic Description	Interval	Identification	Туре	ppmv	Kemarks
1	inc	ĆL.	SILT and CLAY-with trace growe	U-5'	identification	CME	4-5'	3.5' Receivery
	<u> </u>	CL	fill, brown chy, loose	0.3		CME	0.1)	210 Mackery
	-		1-5' CLAY- reddish brown, very				00	
			firm, blocky dry to slightly moist					
5			Tirm, blocky, dry to stignity moist					
3		4.	with promouth colourous nodules, USO			_	9-10'	Cal Darrient
	_	CL	5-10' SILTY CLAY-reddish brown,	5-10		CWE	0.11	S.o' Recceny
			very firm to hard, dry to shantly	1		1	(J) (J)	
			moist, platy, with raliche nochts,					
10			10-15' SILTY CLAY. recklish brown.			1		
10		CL		10-15			14-15"	4.5' Recovery
			occaisionally methed gray, with	10.3		CWE	1.01	41.5 Melloven
			SILT Stringers, moist SILT at Base,	{			0.0	***
	~		soft to firm, NSO	1				
15				1				
15		cul	15-18'SILTY CLAY- As Above, Moist,	المحت	CU39-897-41-		الم، ح.	11.510
	. =	ML	soft to firm, increasingly silty,	'" ~	\$18-14'	CME	12.51	4.5 Recovery
			W. MSILT Stynes, NSC	{	1130		0.0	
		ML	18-20' SILT. readish brown, soft,	{	(vocs)			1 1 1
			with thin streams of moderately	ł	(554)		4	WATER AT 191
20	٠.		comented sondy SILT, wet at Brue			1	24.52	11.51.2
		Sm/	20-25' SILTY SAND - reddish brown,	20 25		CME	241	4.0' Recovery
		Mr	soft, with thin comented layers,			i	0.0	
			some clay matrix, very fire grained,					
		,	NSO	ł				
25	, , , _	sm/	25-30 SILTY CLAY AND SILTY SAND.	 25-30 ¹		l	- 6 2.5	4 = 1 0
		CL	right gray with thin stracks of	C2-20,		CME	11.51	1.5' Recovery
			recklish bram clay, silt-very fire	ł			0.0	
			grained, soft, with some firm	ł			0.0	
			and clayey, NSO	ł			i	
30		sm/	30-35' SILTY SAND- recklish brew			1		
		ML	wi clay matrix, moist to	30-35		CWE	34-52	1.4' Recovery
	* *		wet, soft firm very fregraned,	ł			9.11	
	: :		NSO, grading to CLAYER SILT	ł			0.2	
				ł	i		1	
35	· <u>-</u>		35-40 SILTY SAND- gray with					2 21 0
<u> </u>				35-40		CME		3.5' Recovery
			to firm, with a 2-311 very	ł			1.21	•
	··· .		had comented layer very	Į.			0.5	
			Most to wet, NSG	ł		1		
40				 		+	 	
	ļ		104 = 01	{				
	1	!	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	1				
	1	1	Lagged for GRISPI Resistanly	{				
	1	l	and Califfer.	{	1	1	1	
45	1		Screened 34-39'	-	1			
	4]		-	1			
	Į			-	!	1		
	1			-{				
	1			-	[
50	<u> </u>	L		<u> </u>	L	1		l
NOTE	<u>:S:</u>							

BGS - Below Ground Surface

SS - Split Spoon Sample

NSO - no stain or odor

NS - Not Sampled TOC - Top of Casing G - Grab Sample D - Drive Sample

PID - Photoionization Detector

ppmv - Parts per Million, Volume per Volume

RIG TYPE:

LOCATION:

COMMENTS:

JOB NUMBER: 738695 CLIENT:
BORING NUMBER: 639-697-430/5 BORING DIA.: FAILING FRO TEMPERATURE (°F): 85°F

FTA Z

CLIENT: CONTRACTOR: WEATHER: DRLG MED:

AFREE RUAL TINKER AFB DATE SPUD: 8314,00 , 41,4. ED DATE CMPL: AEI Corp. ELEVATION: DATUM: Cloudy, Mild, Shant Breeze Hollansten Augu GEOLOGIST:

7/29/97 7127197 1252.6 Grand Surface Bree Henry

					Laboratory			
Depth	Pro-	USCS		Sample	Sample	Sample	PID	Remarks
(ft.)	file		Geologic Description	Interval	Identification	Туре	ppmv	
1	53-		0-21 TOPSOIL - dak brown soft, dry	0-5'		CME	4-5'	4.8' Recary
	<u> </u>		Z-5' CLAY- reddish brown with				1.011.0	
	_	Cr	black 5 treates, from to very still					
5			silty in out, slightly mais! NSO				1	
		۸.	5-10' SILTY CLAY- reddish brown,	5-10		CME	9-10'	5.0' Recovery
	7-7	CL	Firm- very firm, platty to blocky			-	20120	3
	7-7		Slightly moist with scatened	í			•	
			nateavecus needles, USO	ĺ				
10	"							
		CL	10-15'SIFTY CLAY- readish bram,	10.15		CME	w.c	4.6' Recent
		"	firm to hard blocky, slightly			l	l .	,
			moist, uso				32/32	
)				
15								
	-	CL	15-20 SILTYCLAY-recklish brown,	15-20		CME		SO' Recover/
		-	firm to hard, blocky, dry to	ļ		Conte	3.015.0	/
	-		Slightly Moist trace black					
			carbanacions specs NSC					
20								
		CLI	20-25' SILT - reddish brown,	SE-52,		CWE		30' Recovery
	=	mL	Firm, clayer, clay to shoutly				2.L	·
			moist, with layers of CLAYSAA					
25								
25			25-301 SILT - rectalish brown, soft	25-301	Cc39-897-43	2		Z.8' Recovery
		MLI	to from strently maist increasingly	23-32	- 529	CME	2 412	C.C Recovery
	1	SWI	sondy at base grading to		(vccs)		3.4/5.4	SATURATED AT
	-		SILTY SAND -NSC		,			801
30			3.5.1					
	l	sm	30-35' SILTYSAND- crangish brown,	50.35'	C639-897-43-	CHE	34.35	5.2' Recury
		-	louse, very fine grained very		534	CVIL	8.1(3.7	
	1		sity trace clay layers, wet		(TOC)		0.1(3.7	
			NSO		0900			
35								
		SM	35-40' SILTY SAND- reddish brown,	35-4C		CME	39-40	2.0' Revenery
	· - ·		10000 clayey-silty, very five				5.3/	,
			grained party sorted, Few 1-2"				2.3	
			ramented layers, NSO					
40	7.5.			40.42		CME	44.4	3.2' Recovery
		SM	40-45' SILTY SAND - redclish brown,		V 1		4.71	3.2 100000
			loose, very fine grained silty to				2.2	
			clayey, wet, USO					
45			45-50' SICTY SAND - reddish	45-50'				Zio' Recovery
73		SM	brown soft to loose claver to	13-50		COME	49-50	
		2,,,	silty very fine grained 3-4" clean				4.41	
	• •		Sand at 491, 2-5" clay at base.				2.2	
	: :		The state of the s					
50	2:-							
NOTE	S:		TD at So	, 1				
BGS -	Below G	round Sur	face SS - Split Spoon Sample					

NS - Not Sampled

G - Grab Sample

TOC - Top of Casing

D - Drive Sample

PID - Photoionization Detector ppmv - Parts per Million, Volume per Volume

NSO- no stain or octor

JOB NUMBER: BORING NUMBER: RIG TYPE: TEMPERATURE (°F):

LOCATION:

COMMENTS:

729691 C639-897-45 BORING DIA.: FAILINGFID 75-85 FTAL

CLIENT: CONTRACTOR: WEATHER: DRLG MED:

TINKER AFB DATE SPUD: 8314"00/4"14"ID AEI Corp. DATE CMPL: ELEVATION: Clear, Sunny, Warm Hollow Stem Auger DATUM: GEOLOGIST:

7131197 7131197 1242.21 US. BMH

					Laboratory	1		
D	D	Hece		Cammia	Laboratory	Commis	DID.	Damanta
Depth	Pro-	USCS	God at Developing	Sample	Sample	Sample	PID	Remarks
(ft.)	file		Geologic Description	Interval	Identification	Туре	ppmv	
1			0-2' TOPSCIL-dark brown, roots	0.5'	1	CME	4-5'	1.2' Recovery
		CL	2-5' SILTY CLAY-brown at				0.91	
			too becoming recklish brown				0.8	
			firm to hard blocky black			1		
5			organic Inclusions, Moist, NSO					
		CL	5-10 SILTY CLAY - reddish	5-10		CME	C	5.0' Recovery
		CL	brown, mother gray in not.	0 .0		· •		
			blocky, dry, Kard to				0.9	
			Very hard Não				0.7	
10			vay kato . Nao					
10		4.	in the company of the contract of	!				4.1' Recovery
-		CL	10-15' SILTY CLAY- recklish	10-15		CME	13-15	7.1 Kecciery
\vdash			brown, mother light gray,				1.01	
\vdash			hard, USO			ŀ	0.7	
						ŀ		
15						Į.		
		CL	15-201SILTYCLAY- reddish brum.	15-20		CME		4.0' Recovery
			blacky, hard with firm to			-	18-50	'
			Friable SILT layers towards			1	5.51	
			the base dry to slightly moist				4.5	
20			NSO					
			20-22' SILT to SILTY CLAY-	20-25			-	3.5' Recorny
		ML		W 23	NOTE:	CME	22-25	3.3 12007
		c	soft to tirm, moist, grading		WATER		4.8140	
	• • •	SM	to SILTY SAND at 221 L rackish			1	1.0110	
25			brown, loose, clayey, very time	ł	AT 20'			
ا			grained, poorly comented, wet			ا ما ا		- 10
	3.5	SM	25-30' SILTY SAND- reddish	25-30	ì	cme.	25-30	2.0' Recovery
			brown to tan, loose to parry			į.	3.13.1	
	-		comented, very fine grained,				2	
			clay matrix ust, U NSO					
30								
	• •	Sm/	30-35' Interpreted SILTYSAND-	30-35	1	CME	3035	Zil' Recovery
			reddish brown, soft, very				211/200	
	~	CL	claver, and SANDY TO SILTY	ĺ			211)	
	-		CLAY- reddish brown, soft to	ĺ		1		
35	1		stiff wery moist NSO	i		1		
			35-40' SILTY SAND-reddish	35-40		CME		1.5' Receivery
		SM	brown, loose to soft clay	J.3 - 7U			35-40	11.5
\vdash	*		ercun, loose to sort clay				3.9/	
	• •		matrix very fire to fire				27	
40			grained, wet, No		ļ,			
40			U. Malaa	40~45		CME		1.51.0
-		5m1	40-45' SAND- arong wh brown,	{			40~15	1.0' Recovery
ļ	```	SP	fire to medium grained pockly			1	3.2/3.7	HARD DRILLIAN
			to moderately soluted, loose,	Į				AT 42-43'
	-::		to marked at base ruet uso		1		1	
45								
			TD at 45'				1	
	1			1		1		
	1			1				
50				1				
NOTE	S:							
1.011	~.							

BGS - Below Ground Surface

SS - Split Spoon Sample

NS - Not Sampled

G - Grab Sample

TOC - Top of Casing

D - Drive Sample

PID - Photoionization Detector ppmv - Parts per Million, Volume per Volume

NSO- No stain or Octor

JOB NUMBER: BORING NUMBER: RIG TYPE: TEMPERATURE (°F): LOCATION:

COMMENTS:

7 29894 CLIENT: BORING DIA .: 2-3928 Faining FIO CONTRACTOR: WEATHER: DRLG MED: FTAZ

AFLEE RNAL TINKER AFB DATE SPUD: 6314"00 14"4"ID AEI Corp. DATE CMPL: ELEVATION: DATUM: Cloudy, Bracy, Warm GEOLOGIST: Holke Stem Auger

7/29/97 7179197 1246.21 30 Bree Henry

Depth Pro USCS Geologic Description Geologic Description 1 21 CL 2-5' Siltricity - Geologic Description CL 2-5' Siltricity - Geologic Description Type ppm CL 2-5' Siltricity - Geologic Description CL 2-5' Siltricity - Geologic Description Type ppm CL 2-5' Siltricity - Geologic Description CL 2-5' Siltricity - Geologic Description Type ppm CL 2-5' Siltricity - Geologic Description CL 2-5' Siltricity - Geologic Description Type ppm CL 2-5' Siltricity - Geologic Description CL 2-
1 27 2 27 2 27
CL Z-5' SILTY CLAY - Orongish brown to real, firm to very Firm, shightly mast, with absorbed promise material The firm sendy in part day 1500 The firm sendy in part day 1500 The firm sendy in part day 1500 The firm sendy in part day 1500 The in-is silty clay - redish brown to 15 The to have group firm hard day Let mothed group silty in testally Chy NSO CL IS-20' SILTY CLAY - backing to backy Let mothed group silty in testally Chy NSO CL IS-20' SILTY CLAY - backing to 1000 The in-is and blocky day to 1000 Slightly most mothed group May 11th get 22-15, greating to 1000 Silty silty at 22-15, greating to 1000 Silty silty at 22-15, greating to 1000 Silty silty at 22-15, greating to 1000 Soft, Clay of Silty shap most, NSO Soft, Clay of Silty shap to mongeth to readen by 1000 Teachest brown, loose, with 34" Layers of party to maden by 1000 To methy the firm grained, clay Methy well the grained, clay Methy well the grained, clay Methy well the grained, clay
CL Z-S' SILTY CLAY - Orenessh brown to real, from to very firm, slightly maist, with abundant organic material to them south in part day NSO CL 10-15' SILTY CLAY - reddish brown, to 15 CL 10-15' SILTY CLAY - reddish brown, to 15 CL 10-15' SILTY CLAY - reddish brown, to 15 CL 15-20' SILTY CLAY - bolding, to 15 CL 15-20' SILTY CLAY - browned, 15-20 Silgntly most, mothed gray English to the part of the silver of the silver of the silver of the silver of party to make the silver of th
rect firm to very firm, slightly maist, with abundant organic males at 5 ML 5-7 SILT- crongly brown, soct 10 CL 7-10 SILTY CLAY- reddish from, mothed grow, firm-hard dry 10 CL 10-15' SILTY CLAY- reddish brown, firm to havel plately to backey with method grow, silty intertally, CME 14-15 1.9' Recovery 1.0' 1.0' 1.0' 1.0' 1.0' 1.0' 1.0' 1.0' 2.0' Recovery 3.1' Recovery 3.1' 3.2' Come 24-25 3.2' Come 24-35 3.2' Come 24-35 3.2' Come 24-30 3.2'
maist, with abindent organic material 5.7 SILT carongly brown, 5ctt 10 thrm Sordy in part, dry 150 CL 7-10 SILTY CLAY- reddish brown, 10-15 CL 10-15' SILTY CLAY- reddish brown, 10-15 Firm to have a platey to boday 11.0 15
The firm sendy in part, dry USO CL 7-10 Silty CLAY- recklish brown, mothed gray firm-hard dry Them to have plately to beday, usin mothed gray silty interbally, and part, MSO CL 15-20' Silty CLAY- small red CL 15-20' Silty CLAY- small red Slightly most mothed gray In part, MSO CL 15-20' Silty CLAY- small red Slightly most, mothed gray In part, MSO CME 19-20' Recovery CME 19-20' Recovery CME 19-20' Recovery CME 19-20' Recovery The part, MSO CME 19-20' Recovery CME 19-20' Recovery Sightly most, mothed gray In part, MSO CME 24-25' Yell' Recovery Soft, Clay et , Slightly most, MSO Soft, Clay et , Slightly most, MSO CME 24-25' Yell' Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft, Clay et , Slightly most, MSO CME 29-30 Silv Recovery Soft Recovery CME 29-30 Silv Recovery Soft Recovery CME 29-30 Silv Recovery CME 29
to firm sendy in part, dry USO 7-10 SILTY CLAY- recklish brown, Mothed gray, firm-hard dry CL 10-15' SILTY CLAY- recklish brown, firm to havel plately to beday, Luth method gray silty interballs, CME 14-15 1.8' Recovery List was suffered gray Slightly most mothed gray Apart, NSO CL1 20-25' SILTY CLAY- SAA becoming, 20-25 ML Dearly silty at 22-23', gracing bo Silt - crangesh brown Prim to Soft, Clayer, slightly most, NSO 25 SM 25-30 SILTY SAND - crangesh to recklish brown, loose, with 34'' layers at party to maderately Camented, new fire grained, clay Mothed fire grained, clay Mathematical services grained, clay Mathematical services grained, clay Mathematical services grained, clay Mathematical services grained, clay Mathematical services grained, clay Mothed services and services grained, clay Mathematical services grained, clay Mothed services and services grained, clay Mothed services and services grained, clay Mothed services and
CL 7-10 SILTY CLAY- recklish brown, 10-15 CL 10-15' SILTY CLAY- recklish brown, 10-15 Firm to havel plately to backly 11-3 Lind 11-15 1-3' Recovery 11-3' Receivery 11-3' R
mothed gray firm-hard clay 10
10 CL 10-15' SILTY CLAY- recklish brown, 10-15 From to haved plately to backy With methed gray Silty interballs CL 15-20' SILTY CLAY- Sinch red, 15-20 From hard, blocky, 11 to Slightly most, mothed gray May 11th at 22-23', grading 12 Soft, Clay ex, Slightly most, 150 25 SM 25-30 SILTY SAND - croagush to recklish brown, 100se, with 54" layers of pearly to mederately Gray and the procedure of the proce
CL 10-15' SILTY CLAY-recklyh brown, 10-15 From to havel platey to body, With method gray Silty interball, Cly, NSCO CME 19-20 CME 29-30 SiO' Recovery CME 29-30 SiO'
from to havel plater to bedry with methed gray silty interbally CL 15-201 SILTY CLAY- brick red, from hood, blocky, dry to Slightly most, mothed gray m part, NSO CL 20-25' SILTY CLAY- SAN becoming 20-25 ML 10-25' SILTY CLAY- SAN becoming 20-25 ML 10-25' SILTY CLAY- SAN becoming 20-25 ML 10-25' SILTY CLAY- SAN becoming 20-25 SML 50-25' SILTY CLAY- SAN becoming 20-25 SML 20-25' SILTY CLAY- SAN becoming 20-25 SML 20-25' SILTY CLAY- SAN becoming 20-25 SML 20-25' SILTY SAND - Orongish to 25-30 CME 29-30 SIO' Recovery 1.7 SML 25-30 SILTY SAND - Orongish to 25-30 CME 29-30 SIO' Recovery SSI Layers of poemly to moderately CAMENATE of poemly to moderately CAMENATE of Livery fire grained clay Mother with Livery fire grained clay Mother with Livery fire grained clay
With mathed gray sity intertails Chy NSCO Chy NSCO Chy NSCO Chy NSCO Chy NSCO Chy NSCO Chy NSCO Sighty Chay by to Sighty Most Mothed gray Mapped NSCO Che 24.25 Che
CL 15-20' SILTY CHAY- STICK red, 15-20 The hood, blocky thy to Slightly most mothed gray A part, NSO CLI 70-25' SILTY CHAY- SAA becoming 20-25 The hong silty at 22-23' grading to SILT-oranges brown from to Soft, Clayer, Slightly most, NSO 25 SM 25-30 SILTY SAND- oranges to reddish brown loose with 34" layers of poorly to moderately Camented, new fire grained clay Mother wet his or
CL 15-201 SILTY CNAY- Brick red, 15-20 The hard, blocky thy to Slightly most, mottled gray May 11th at 22-231, grading to SILT - crangesh brown from to Soft, clayer, slightly most, NSC 25 SM 25-30 SILTY SAND - crangesh to reddish brown loose, with 3411 layers of poorly to moderately Camented, new fire grained clay Mathy wet his of
firm - hard, blocky, By to Slightly most, mothed gray m part, NSO CLI 70-25' SILTY CLAY-SAA becoming 20-25 ML 18 11 St 22-23' Gracking by SILT - crangish brown from to Soft, Clayer, Slightly most, NSO 25 SM 25-30 SILTY SAND - crangish to reddish brown, loose, with 34" layers of poorly to moderately Comented, seny fire grained clay Mathy wet NSO Methy wet NSO
from hard, blocky thy to Slightly most, mottled gray In part, NSO CLI 20 -25' SILTY CLAY- SAA becoming 20-25 ML remy silty at 22-23', gracing to SILT-crangesh brown from to Soft, Clayer, Slightly most, NSO Soft, Clayer, Slightly most, NSO 25 -30 SILTY SAND-crangesh to reddish brown loose, with 34" Inyers of party to maderately Inyers of party to maderately Camented, new fire grained clay Mother wet his of
Slightly most mothed gray m part, NSO CLI 70-25' SILTY CLAY- SAA becoming 20-25 The youth of 22-23' grading to SILTY congraph brown from to SILT - oranged brown from to Soft, Clay of Slightly most, NSO Sm 25-30 SILTY SAND - oranged to 25-30 CME 29-30 3:0' Recarry reddish brown loose with 34" layers of party to maderately layers of party to maderately Convented ivery fire grained clay Mothy wet NSO Mothy wet NSO
20 CLI 70-25' SILTY CLAY-SAA becoming 20-25 ML 1/2 11th at 22-23' Gracking to SILT - crangish brown from to Soft, Clayer, Slighty most, ASO 25 SM 25-30 SILTY SAND - crangish to reddish brown loose, with 34" 1 reddish brown loose, with 34" 1 repers of poorly to moderately Camented, went fire grained clay Mothy wet 1500
CLI 70-25' SILTY CLAY- SAA becoming 20-25 ML 100 M SILTY CLAY- SAA becoming 20-25 ML 100 M SILTY CLAY- SAA becoming 20-25 SILTY - crangish brown 12 mm to 25-30 Soft, Clay ex Slightly most, ASC SM 25-30 SILTY SAND - crongish to 25-30 reciclish brown 1000e with 54" layers of powry to moderately layers of powry to moderately Comented 100 M to 100 M to 100 M Mother wet 1000
25 SM 25-30 SILTY SAND - crongish to reciclish brown, 1000000, 1000000, 1000000, 1000000, 100000, 100000, 1000000, 1000000, 10000000, 1000000, 1000000, 10000000, 1000000, 1000000, 10000000, 100000000
Soft, Clayed, Slightly most, NSO Some 25-30 SILTY SAND - arough to reciclish brown, loose, with 374" layers of pointy to moderately Comented, new fine grained, clay matrix wet NSO
Soft, Clayed, Slightly most, NSO Some 25-30 SILTY SAND - crongish to reciclish brown, loose, with 374" Since Recarded Silvery for acceptably Comented, new fire grained clay matrix wet NSO
25 SM 25-30 SILTY SAND - crongish to 25-30 CME 29-30 SiO' Receivery reciclish brown, loose, with 3-4" layers of poorly to moderately Comented, very fire grained clay motory wet 1500
SM 25-30 SILTY SAND - crongish to 25-30 reciclish brown, loose, with 54" layers of pourly to moderately Comented, new fine grained clay mother wet NSC
recidish brown, wase, with 37" layers of pointy to moderately comented, very fire grained, clay matrix wet 1500
namented, very fire grained, clay
30 matrix up t NSC
30 matrix upt visc
CL 30-33' SILT & SILTY CLAY - SAA 30-35 CME 33-75 4.5' Pewery
M2
SM 33-35' SILTY SAND - very SILLY
to clayer - SAA
35
-: SMI 35-40' SICTY SAND and SILT - 55-40 CME 57-40 4.2' Rescovery
me reddish brown, soft to firm
with layers of well comen ted, clayer, 5.0
40 went moist No
SM 40-451 SILTY SAND - recklish brain 40-45 CME 4445 4.2 Recovery
100se to soft, with 34 1 Cyers of
hand and well cemental salue
very firegraphed wet uses
45
SM/ 45-50' SILTY SAND. reddin 45-50 CMF 49.50 0.8' Receivery
ML loroun, locale to soft clayer,
grading to clever silt, west
50
NOTES: $TD = SO$ ' BGS - Below Ground Surface SS - Split Spoon Sample

BGS - Below Ground Surface

NS - Not Sampled

G - Grab Sample

TOC - Top of Casing

D - Drive Sample

PID - Photoionization Detector ppmv - Parts per Million, Volume per Volume

NSO- no stain or octor

PARSONS ENGINEERING SCIENCE, INC.

JOB NUMBER: BORING NUMBER: RIG TYPE: TEMPERATURE (°F): LOCATION:

COMMENTS:

729691 2-3933 FAILINGFIC 65-90°F FTA Z

CLIENT: BORING DIA.: CONTRACTOR: WEATHER: DRLG MED:

AFREE ENA | TINKER AFB DATE SPUD: QI"1111 QO "1163 DATE CMPL: AEI Corp. ELEVATION: DATUM: Sunny, Hot GEOLOGIST: Hulkow Stem Auger

7130197 7130197 Grand Surface Bruce Henry

					<u> </u>	Laboratory			
CL STOCKET - Clean brown costs Come 45' 3 to Pecceny Come 45' 3 to Pecceny Signify main to retain brown Signify main to serve the mode of the significant main to retain brown Come 415' 515' 515' 515' 515' 515' 515' 515'	Depth	Pro-	USCS		Sample	-	Sample	PID	Remarks
SITTY CLAY- brown to redd to Sugarthy most some bicks and the sugarthy most some bicks are bicks as the sugarthy most some bicks are bicks as the sugarthy to such some motival activated some clay bicks are bicks as the sugarthy to such some some clay bicks and bicks are bicks as the sugarthy and bicks are sugarthy most some clay bicks and bicks are sugarthy and bic	(ft.)	file		Geologic Description	Interval	Identification	Type	ppmv	
Second three to experience day to slightly made to slightly made to see block original productions to slight out of some block original productions to slight out of slightly made to slight out of slightly made to slightly made	1			0-1' TOPSOIL-doublown, roots	05'		CME	4-5'	3. E' Recovery
SIGNATIN Mesh send block organic machenal STO SUBMITTED LAY SAR with INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SILT CLAY SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC. 15° SAR WITH INC.			CL					1.4 11.1	,
5 - CL Consider the control of the c				brown, firm to very firm, dry to					
CL STET SITT CLAY SAR with 5-10 Checkent break cream mothers 1-2 In the control of base 1-2 In the control of base 1-2 In the control of base 1-2 In the control of base 1-2 In the control of base 1-2 In the control of base 1-2 In the control of base 1-2 In the control of contr						-			
abundant black carpain makeral apropriation to see a second to see apropriation to see appearance of see apropriation to see appearance of see appeara	5								1.0
according to Siet at bose 10			CL		5.10		CME		4.4' Recorn
10 - ML 11				abundant black organic material,	!				
mu locis's sict crossish brown, method growy dry blocky NSO with sictry dry blocky NSO with sictry carry say, blocky NSO with sictry carry say, blocky NSO with sictry carry say, blocky NSO sitry carry say, increasing is to sitry carry say, increasing is to sitry carry say, increasing is to sitry carry say, increasing is to sitry carry say, increasing is to sitry carry say, increasing is to soft to the sitry carry in part sound, in part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist has soft to the part moist have a soft to the part moist has soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft to the part moist have a soft hard a sound of the law of the part moist have a soft hard sound part of the part moist have a soft hard sound part of the part moist have a soft hard sound surface. So part of hard sound surface. So part of hard sound surface. So part of hard sound surface. So part of hard sound surface. So split spoon sample.				grading to SILT at base				1.2	
The control of the co	10		WL		{				
CL with Sutt CLAY recicish brown, from to hard and from the hard a	10			1. (2) (14.15	H. H. Para Maria
think to kind and The sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Sold to from clayer in part Social in part, month to very month and Sold to from clayer in part Sold to from clayer Sold		-	ML	10-16 SICI -crangin brum,	10-12		cine	2.01	4.4 MELLENY
think to kind and The sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Silty with sold interpretated system Sold to from clayer in part Social in part, month to very month and Sold to from clayer in part Sold to from clayer Sold			CL	Mothed gray dry blocky NSO			1	1.3	
15 - CL 15-20' SILTY CLAY-SAA increasing 15-20 Silty with soft inderbacked 34' Silt (weeks, clay bis to thing moint) Soft to trim, clayer in part, Soft to trim, clayer,				Care Jahrand State County	{			' -	
SILT CAMP SAN Increasing to the sixty with soft intersected 34 soft sitty with soft intersected 35 soft with soft intersection of soft soft soft soft soft soft soft s	15	<u></u>			i				
SILT Cyen's, Chry to stephy ML SILT Cyen's, Chry to stephy ME SILT Cyen's, Chry to stephy ME ST To ST TO STAND TO STAND Soft to Firm, Chayer in port So	15	_::	A	15- 201 Survice AN- SAA Transfina	15-20		CHE	19-20	2 Page Page
SILT (cyes) clay to stephing meist NSD Do 25' SILT reddish brown, 2025 Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm, clayer in part, Soft to firm for a clayer Soft to firm f		_	CL	sill the sold who delent 3 th	Ą		CINE	300/	Zil Ballery
20 The Lorenty NSD Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm, clayer in port, Soft to firm grained, LSSD Soft Soft Layers of maderally Commanded firm grained, Layer Soft to firm grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Soft Clay Mathry for grained, LSSD Soft Soft Clay Mathry for grained, LSSD Soft Soft Soft Soft Soft Soft Soft Soft		_:		SILLA COLLAR SCAL MARS PROCED 24	{				
20 The Lorest Start - reddish brown, Soft to time, clayer in port, Sort to time, clayer in port, Sort to time, clayer in port, Sort to time, clayer in port, Sort to time, clayer in port, Sort to time and the soft time and the soft time and time and the soft time and the soft time and tim				SEET TEYERS EVY 485 EGWING	 				
20-25' SILT-reddish brown, 20-25 Soft to From Clayer in part, Soft to From Clayer in part, Soft to From Clayer in part, Soft to From Clayer in part, Soft to From Clayer in part, Soft Manual Cose to Soft, with a few to Programmed of From Grands of Manual Cose to Soft, with a few to Programmed of From Grands of Manual Cose to Soft, with a few to Programmed of From Grands of Manual Cose to Soft, with a few to Programmed of From Grands of Manual Cose to Soft, with a few to Programmed of From Grands of Manual Cose to Soft of From Manual Cose to Soft of From Manual Cose to Soft of From Manual Cose to Soft of From Manual Cose to Manua	20			TANGET, NZO	i]		
Seft to firm, clayer in part, Sondy in goal, moist to very 25 Sm 25-30 SILTT SAND reddish thous, loose to soft with a tent 12' layer at moderally comented fire ground, clay matrix, wet NSD 30 30-35 SILTT SAND reddish Thous, loose to soft, with a comented fire ground, clay matrix fire ground, also 30 30-35 SILTT SAND reddish Thous, loose to soft, clay matrix fire ground, also matrix fire ground, also 35 SM 35-40' SAND reddish brown, 35-40 SM 35-40' SAND reddish brown, 35-40 SM 35-40' SAND reddish brown, 35-40 Come 35-40 SM 35-40' SAND reddish brown, 35-40 Come 35-40 SM 35-40' SAND reddish brown, 35-40 Come 35-40 SC Recomy Come 35-40 SC Recomy Tock Tock SM SM SM SM SM SM SM SM SM S	20		m t	20-25' SILT - redelah borne	20.25	i	CME	24-25	O.S' Recovery
25 Sondy in part, moist to very Moist, NSC 25-30 SILTT SAND, reddish Losse to soft with a two 1-2' levers of motoralety comented fire ground, clay matrix, wet NSC Son 30-35 SILTT SAND- reddish brown, lease to soft clay matrix the ground, NSC Son 30-35 SILTT SAND- reddish brown, lease to soft clay matrix the ground, NSC Son 35-40' SANN- reddish brown, soft clay fire to mechanically clay notice, party softed, clay notice, party softed, clay notice, party softed, clay to some fire fire to macked by comented fire to macked by comented fire to macked y comented fire to Mechan agained 45 Son 45-50 Con 13'-10' 13'-10' 2-3458 Con 2-3458 Con 34'-35' Con 35'-40' SANN- reddish brown, 35-4c Con 35'-40' 30'-30' Con 13'-40' 30'-30' Con 1									
25 SM 2S-30 SILTY SAND- reddish 25-30 long from loose to soft, with a few 1-2' loyers of motoralely economisted, fine ground, Ckey mathry, wet NSD 30-35 SILTY SAND- reddish 30-35 SILTY SAND- reddish brown, locate to soft, chey mathry fine ground, ASC (TOC) SM 35-40' SANN - reddish brown, were fine to mechan ground lovely fine to mechan ground lovely fine to mechan ground lovely fine to mechan ground lovely fine to mechan ground lovely fine to mechan ground lovely fine to mechan ground lovely sand-wet, with 12" of motoracting comented fine to mechan ground lovely fine to medium ground lovely fine to medium ground lovely fine to medium ground lovely lovely fine to medium ground lovely lovely fine to medium ground lovely lovely fine to medium ground lovely lovely fine to medium ground lovely lovely fine to medium ground lovely lovel				south is and assist to very				1.7	
25 30 SILTY SAND - reciclish 25:30 25-30 SILTY SAND - reciclish 25:30 forum, loose to soft, with a carranted, fire grained, Ckey mother, wet NSO 30-35 SILTY SAND - reciclish 534 (TOC) 30 SM 35-40' SANN - reciclish brown, loose to soft, cher mathy the grained, NSO 35-40 (TOC) 35 SM 35-40' SANN - reciclish brown, 35-40 (TOC) 35 SM 35-40' SANN - reciclish brown, 35-40 (TOC) 35 SM 35-40' SANN - reciclish brown, 35-40 (TOC) 35 SM 35-40' SANN - reciclish brown, 35-40 (TOC) 36 SM 36-40' SANN - reciclish brown, 35-40 (TOC) 36 SM 36-40' SANN - reciclish brown, 35-40 (TOC) SM 40-41' SAND - SAN- wet, with 1-2" of part 10 moderately carrented to the lock with 1-2" of moderately carrented to the loc	7.00			moist war					
tew 1-2' (cyers at moderally comented, fine grained, clay mothins, wet 1 NSD 30 SM 30-35 SILTY SAND- reddish brown, locate to soft, clay mathy three grained, ASC SM 35-40' SAND - reddish brown, 12 Wery fine to mechan grained, 12' at 12'	25				i			l	
tew 1-2' (cyers at moderally comented, fine grained, clay mothins, wet 1 NSD 30 SM 30-35 SILTY SAND- reddish brown, locate to soft, clay mathy three grained, ASC SM 35-40' SAND - reddish brown, 12 Wery fine to mechan grained, 12' at 12'		· - : :	500	deliche Mund TT SANA . Medeliche	25.30		- 45	29-30	1.8' Recovery
tew 1-2' (cyers at moderally comented, fine grained, clay mothins, wet 1 NSD 30 SM 30-35 SILTY SAND- reddish brown, locate to soft, clay mathy three grained, ASC SM 35-40' SAND - reddish brown, 12 Wery fine to mechan grained, 12' at 12'		:::	J				CME	2.8/	7
Somewheel fine growed, ckey methix, wet NSO 30-35 SILTY SAND-readish brown loase to soft char mathy the growns, NSO 35 SM SM 35-40' SAND-readish brown, wery fine to mechan crowd, ckey mathy postly softed, iset look with 1-2" of post to moderately comented your's SAND-SAN-wet with marky sand sand-wet with so To at SO' NOTES: To at SO' BGS-Below Ground Surface SS-Split Spoon Sample		٠			i			1.8	
Som 35-40' SAND - reddish brown, 12 35-40' SAND - reddish brown, 12 2-348 B 334 (TOC) Som 35-40' SAND - reddish brown, 35-40 (TOC) Som 35-40' SAND - reddish brown, 35-40 (TOC) Som 35-40' SAND - reddish brown, 35-40 (TOC) Som 35-40' SAND - reddish brown, 35-40 (TOC) Som 35-40' SAND - reddish brown, 35-40 (TOC) Som 40-40' SAND - SAND - 40-40 (TOC) Som 40-40' SAND - SAND - 40-40 (TOC) Som 40-40' SAND - SAND - 40-40 (TOC) Som 45-50 - 0.5' of hard SAND - 45-50 (T				comented for arranged ckey	i		1		
SM 30-35 SILTY SAND- neddish 30-35 SAY Thomas, loase to soft, clear methy, the ground, ASO SM 35-40' SAND - reddish brown, Wery line to mechan spained, Clear matrix, parriy softed, Lear loave with 1-2" of parr to moderately comented Mo-45' SAND-SAA-wet, with macrocaling thin layers of mechan grained Mo-45' SAND-SAA-wet, with Medium grained SP 45-50 - 0.5' of hard SAND- Line to medium grained well CL comented, mentying SICTY CLAY- reddish brown, hard MOTES: TO at 50' BGS-Below Ground Surface SS-Split Spoon Sample	30			matrix cust Non	1				
matry fre grained. NSC SM 35-40' SANA - reddish brim, 35-4c Levy fine to mechan grained, 25-4c Levy fine to mechan grained, 25-4c Levy fine to mechan grained, 25-393B/ SM 40-45' SAND-SAA-Wet with 10-45' SAND-Mechan grained SP 45-50 - 0.5' of hard SAND-Hosting SILTY CL Cymented, One-tying SILTY CLAY-medish brian, hard NOTES: TO at SO' BGS-Below Ground Surface SS-Split Spoon Sample		٠,٠	SM		3035	Z-305D)	CMF	34-35	2.0' Recorns
matry fire grained, NSC TO C) Matry fire grained, NSC Inc. SM 35-40' SANN - reddish brown, 35-4c Very fire to mechan grained, clay matry, poorly softed, incl lock with 1-2" of poor to moderately comented for soft increasing thin layers of moderately comented fire to mechan grained. SM 40-45' SANN - SAN - Wet with 10-45' S44 Increasing thin layers of moderately comented fire to mechan grained. SP 45-50 - 0.5' of hard SAND - 45-50 CHE 35-40 SO - Wety in the soft with 10-45' SANN - 10-45' SANN			J		1	334	-		,
SM 35-40' SANN - reddish brim, 35-46 Wery fine to mechan grained, Clay matrix, partly softed, Luet look with 1-2" of part to moderately camented Y0-45' SAND-SAR-Wet with Increasing thin layers of moderately camented, fine to Mechan grained SP 45-50 - 0.5' of hard SAND- Increasing thin grained well CL camented chertying SILTY CLAY-reddish brain, hard NOTES: TO at 50' BGS-Below Ground Surface SS-Split Spoon Sample		•••			1	(TOC)		1.2	
SM 35-40' SANN - reddyn brum, 35-40 Very fine to mechan grained, Clay matrix, party sorted, Luet, look with 1-2" of pear to moderately camented 40 SM 40-45' SANN-SAA-Wet with 40-45' SANN SANN-SAA-Wet with 10-45' SANN SANN SANN SANN SANN SANN SANN SAN]				
Levy that to mechanismand, Chay mother, party sorted, Luet, law with 1-2" of party to moderately comented Party SAND-SAR-Luet with NOTES: To at SO' BGS-Below Ground Surface Sorted, Chay mother, party sorted, Luet, law with 1-2" of party to moderately comented NOTES: To at SO' Sorted, Sorted, Sorted, Comented, moderately comented Sorted To at SO' BGS-Below Ground Surface Sorted, Sorte	35]				
Levy that to mechanismand, Chay mother, party sorted, Luet, law with 1-2" of party to moderately comented Party SAND-SAR-Luet with NOTES: To at SO' BGS-Below Ground Surface Sorted, Chay mother, party sorted, Luet, law with 1-2" of party to moderately comented NOTES: To at SO' Sorted, Sorted, Sorted, Comented, moderately comented Sorted To at SO' BGS-Below Ground Surface Sorted, Sorte			SM	35-40' SAND - reddish brown,	3546			39-40	30' Rocowy
ctay matrix poorly sorted, inet look with 1-2" of poor to moderately comented 10 SM 40-45' SAND-SAA-wet with increasing thin layers of more value comented fire to Medium agained 15 SP 45-50 - 0.5' of hard SAND- fine to modium agained well care mented medium agained well comented medium agained well CLAY-reddish bream, hard 10 SP 45-50 CRE 45-50				very fine to medium socional			Crie	2.51	
poor to moderately comented SM 40-45' SAND-SAA-wet, with Increasing thin layers of Moderately comented fine to Medium grained SP 45-50 - 0.5' of hard SAND- The to medium grained well CL comented, comentarying SICTY CLAY-reddish bream, hard MOTES: TO at 50' BGS-Below Ground Surface SS-Split Spoon Sample				clay matrix, pourly sorted,	Į			-	
SM 40-45' SAND-SAA-Wet with 40-45' Z-393B) CINE 4445' SHY INCREASING THIN layers OF Mark retary cerented, fire to Medium grained SP 45-50 - 0.5' of hard SAND- Fine to madium grained well CL cemented, cheritying SKLTY CLAY-reddish bream, hard NOTES: TO at 50' BGS-Below Ground Surface SS-Split Spoon Sample				livet loose with 1-2" of	Į			`	
SM 40-45' SAND-SAA-Wet with 40-45' Z-393B) CINE 4445' SHY INCREASING THIN layers OF Mark retary cerented, fire to Medium grained SP 45-50 - 0.5' of hard SAND- Fine to madium grained well CL cemented, cheritying SKLTY CLAY-reddish bream, hard NOTES: TO at 50' BGS-Below Ground Surface SS-Split Spoon Sample	40	• •		poor to moderately comented	Į		1		
More valey cerented fire to Medium grained 45 SP 45-50-0.5' of hard SAND-45-50 Che to medium grained well Che comented medium grained well Che comented medium security ing Section SO TO at 50' BGS-Below Ground Surface SS-Split Spoon Sample			sm	46-45, SAND-SAA-met, with	140-45	2-3938/	L		2.2' Reasony
Medium grained 45 SP 45-50 - 0.5' of hard SAND - 45-50 CL Comented, modium grained well CL Comented, modium starty CLAY - reddish bream, hard NOTES: TO at 50' BGS - Below Ground Surface SS - Split Spoon Sample				increasing thin layers of	Į	544	CHE		
Medium grained 45 45 45 45-50 - 0.5' of hard SAND - 45-50 Fine to madium grained well co mented, overlying SICTY CLAY - reddish brown, hard NOTES: TO at 50' BGS - Below Ground Surface SS - Split Spoon Sample				more rately comented fine to	Į	(Tac)		'	
SP 45-50 - 0.5' of hard SAND. The to making grained well CH comented, motiving SICTY CLAY-reddish bream, hard NOTES: TO at 50' BGS-Below Ground Surface SS-Split Spoon Sample				medium arained.	Į			06	
time to making grained well comented metrying Sicty CLAY - reddish bream, hard NOTES: TO at SO' BGS - Below Ground Surface SS - Split Spoon Sample	45			, , , , , , , , , , , , , , , , , , ,					1 4 0
CL Comented metring SICTY CLAY-reddish brown, hard NOTES: BGS - Below Ground Surface SS - Split Spoon Sample	<u> </u>		SP		42.20		EME		1.8' Kaccarey
SO CLAY - reddish brewn, heard NOTES: BGS - Below Ground Surface SS - Split Spoon Sample	<u> </u>	7			4			NR	,
NOTES: BGS - Below Ground Surface SS - Split Spoon Sample		1::.	CL		-	l	1		
NOTES: TO Q+ SO' BGS - Below Ground Surface SS - Split Spoon Sample	<u> </u>	7:		CLAY- reddish brewn, hard	-	,			
BGS - Below Ground Surface SS - Split Spoon Sample		<u> </u>	<u> </u>		<u></u>	i	L	L	1
	-								

NS - Not Sampled

G - Grab Sample

TOC - Top of Casing

D - Drive Sample

PID - Photoionization Detector ppmv - Parts per Million, Volume per Volume NSO - no stain or

ASSOCIATED ENVIRONMENTAL IND.

HOLE NO: CG39B9741

LOGGED: 07/28/97 at

TINKER AIR FORCE BASE SE/4 SEC 22 T11N R2W COMPANY AREA

PROSPECT N/A OPERATOR MM

DRILLING DATA

PERM DAT GL

ELEV AZIMUTH

INCLIN

HOLE POS HOLE DIA 9" 40' DEPTH

CASING DATA

CASE-TYP CASE-DTH CASE-DIA

CASE-THK

FLUID DATA

FLD TYPE FLUIDLEV AUGER

TD-3 LOGTYPE

TOOL: Tmrg--Merged Log

Serial Number: 886

Spacer (cm) : n/a

CHANNEL DATA

CHANNEL	AMPLITUDE		CHART	AREA	SENSOR OFFSET	FILTER CNST	PLOT TYPE
1 - GAMMA	0.0 -	200.0	0.0 -	10.0	328	4.0	4
2 - PR	0 -	2000	12.0 -	32.0	453	0.0	2
2 - SP	-200 -	200	0.0 -	10.0	453	0.0	2
3 - CALIPER	4.00 -	24.00	12.0 -	32.0	318	0.0	2

Calibration Data

3	CHANNEL	**************************************	slope	offset	C1	C2	C3	C4
	1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	API ohm mV IN	0.326204 -0.89890	0.000000 -569.169 3428.191 -6.11253	3568 0.000 0.000 3.100		100.0 1500 6.100	0.000 2051 2145 2977

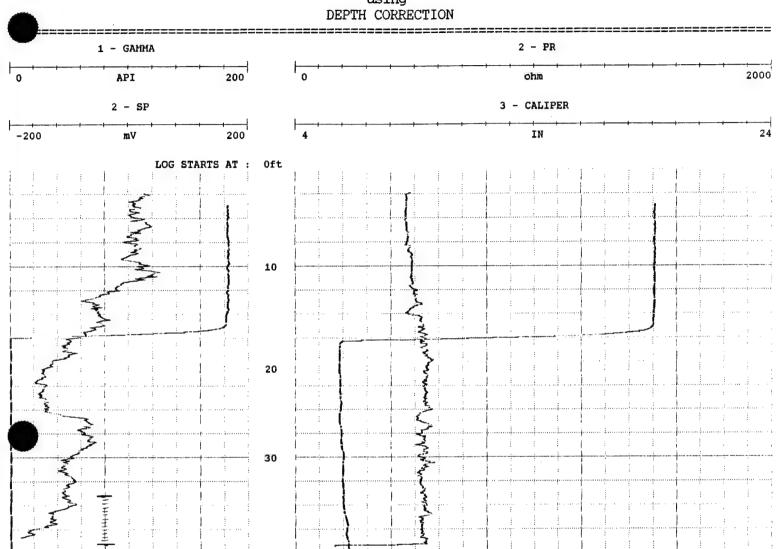
Log No. FILENAME

1 - CG3941 2

2 - CG3941 4 3 - CG3941 5

LOG ENDS AT : 40ft

SCALE: 10 ft/in using



ASSOCIATED ENVIRONMENTAL IND.

HOLE NO: CG39B9743

LOGGED: 07/29/97 at

TINKER AIR FORCE BASE SE/4 SEC 22 T11N R2W COMPANY AREA

PROSPECT N/A OPERATOR MM

DRILLING DATA

PERM DAT ELEV

AZIMUTH INCLIN

HOLE POS HOLE DIA DEPTH

50'

CASING DATA

CASE-TYP

CASE-DIH CASE-DIA CASE-THK

FLUID DATA

FLD TYPE FLUIDLEV AUGER

LOGTYPE TD-3

TOOL: Tmrg--Merged Log

Serial Number: 886

Spacer (cm) : n/a

CHANNEL DATA

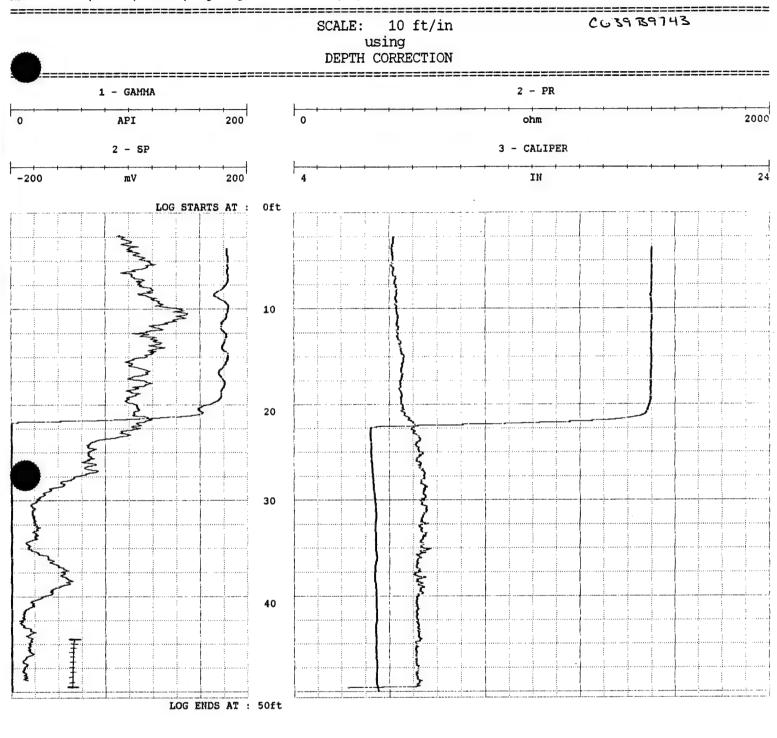
CHANNEL	AMPLITUDE		CHART	AREA	SENSOR OFFSET	FILTER CNST	PLOT TYPE
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	1100	200.0 2000 200 24.00	0.0 - 12.0 - 0.0 - 12.0 -	10.0 32.0 10.0 32.0	328 453 453 318	4.0 0.0 0.0 0.0	4 2 2 2 2

Calibration Data

CHANNEL		slope	offset	C1	C2	C3	C4
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	mV	0.326204 -0.89890 0.004103	-6.11253	3568 0.000 0.000 3.100	3814 2245	100.0 1500 6.100	2051 2145 2977

Log No. FILENAME

1 - CG3943 1 2 - CG3943 3 3 - CG3943 4



ASSOCIATED ENVIRONMENTAL IND.

HOLE NO: CG39B9743S

LOGGED: 07/30/97 at

TINKER AIR FORCE BASE SE/4 SEC 22 T11N R2W COMPANY AREA

PROSPECT N/A OPERATOR MM

DRILLING DATA

PERM DAT GL

ELEV AZIMUTH INCLIN

HOLE POS HOLE DIA 9" 371 DEPTH

CASING DATA

CASE-TYP CASE-DTH CASE-DIA CASE-THK

FLUID DATA

FLD TYPE FLUIDLEV AUGER

TD-3 LOGTYPE

TOOL: Tmrg--Merged Log Serial Number: 886 Spacer (cm) : n/a

CHANNEL DATA

CHANNEL	AMPLITUDE		CHART	AREA	SENSOR OFFSET	FILTER CNST	PLOT TYPE
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	0.0 - 0 - -200 - 4.00 -	200.0 2000 200 200 24.00	0.0 - 12.0 - 0.0 - 12.0 -	10.0 32.0 10.0 32.0	328 453 453 318	4.0 0.0 0.0 0.0	4 2 2 2 2

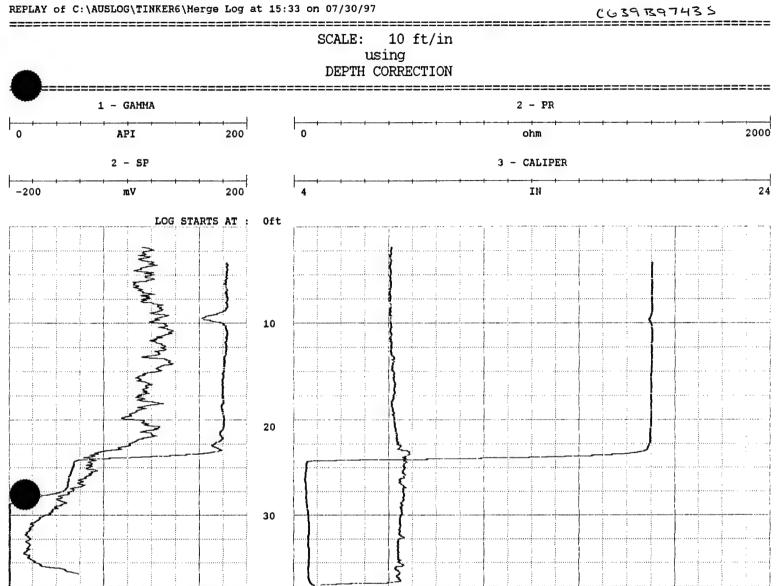
Calibration Data

The same and a second	CHANNEL		slope	offset	C1	C2	C3	C4
	1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	API ohm mV IN	0.326204 -0.89890 0.004103		0.000 0.000 3.100	1745 3814 2245	100.0 1500 6.100	0.000 2051 2145 2977

Log No. FILENAME

1 - CG3943S 2 2 - CG3943S 4 3 - CG3943S 5

LOG ENDS AT : 37ft



ASSOCIATED ENVIRONMENTAL IND.

HOLE NO: CG39B9745

LOGGED: 07/31/97 at

TINKER AIR FORCE BASE SE/4 SEC 22 T11N R2W COMPANY AREA

PROSPECT \tilde{N}/A OPERATOR MM

DRILLING DATA

PERM DAT GL

ELEV AZIMUTH INCLIN HOLE POS HOLE DIA 9" 45' DEPTH

CASING DATA

CASE-TYP CASE-DTH CASE-DIA CASE-THK

FLUID DATA

FLD TYPE AUGER FLUIDLEV

TD-3LOGTYPE

TOOL: Tmrg--Merged Log Serial Number: 886 Spacer (cm) : n/a

CHANNEL DATA

CHANNEL	AMPLITUDE		CHART	AREA	SENSOR OFFSET	FILTER CNST	PLOT TYPE
1 - GAMMA	0.0 -	200.0	0.0 -	10.0	328	4.0	4
2 - PR	0 -	2000	12.0 -	32.0	453	0.0	2
2 - SP	-200 -	200	0.0 -	10.0	453	0.0	2
3 - CALIPER	4.00 -	24.00	12.0 -	32.0	318	0.0	2

Calibration Data

CHANNEL	***************************************	slope	offset	C1	C2	C3	C4	
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	API ohm mV IN	0.326204 -0.89890 0.004103	0.000000 -569.169 3428.191 -6.11253	0.000 0.000 3.100	2245	100.0 1500 6.100	2051 2145 2977	

Log No. FILENAME

1 - CG3945 2 - CG3945 3 - CG3945

C639 B9745 10 ft/in SCALE: using DEPTH CORRECTION 1 - GAMMA 200 2000 API ohm 2 - SP 3 - CALIPER -200 mV 200 IN LOG STARTS AT : 10 20 30

40

LOG ENDS AT : 45ft

ASSOCIATED ENVIRONMENTAL IND.

HOLE NO: 2-392B

LOGGED: 07/29/97 at

COMPANY TINKER AIR FORCE BASE SE/4 SEC 22 T11N R2W N/A

OPERATOR MM

DRILLING DATA

PERM DAT GL ELEV AZIMUTH INCLIN HOLE POS HOLE DIA 9" 50' DEPTH

CASING DATA

CASE-TYP CASE-DTH CASE-DIA CASE-THK

FLUID DATA

FLD TYPE AUGER FLUIDLEV

TD-3LOGTYPE

TOOL: Tmrg--Merged Log Serial Number: 886 Spacer (cm) : n/a

CHANNEL DATA

CHANNEL	AMPLITUDE	RANGE	CHART	AREA	SENSOR OFFSET	FILTER CNST	PLOT TYPE
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	0.0 - 0 - -200 - 4.00 -	200.0 2000 200 24.00	0.0 - 12.0 - 0.0 - 12.0 -	10.0 32.0 10.0 32.0	328 453 453 318	4.0 0.0 0.0 0.0	4 2 2 2 2

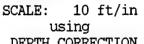
Calibration Data

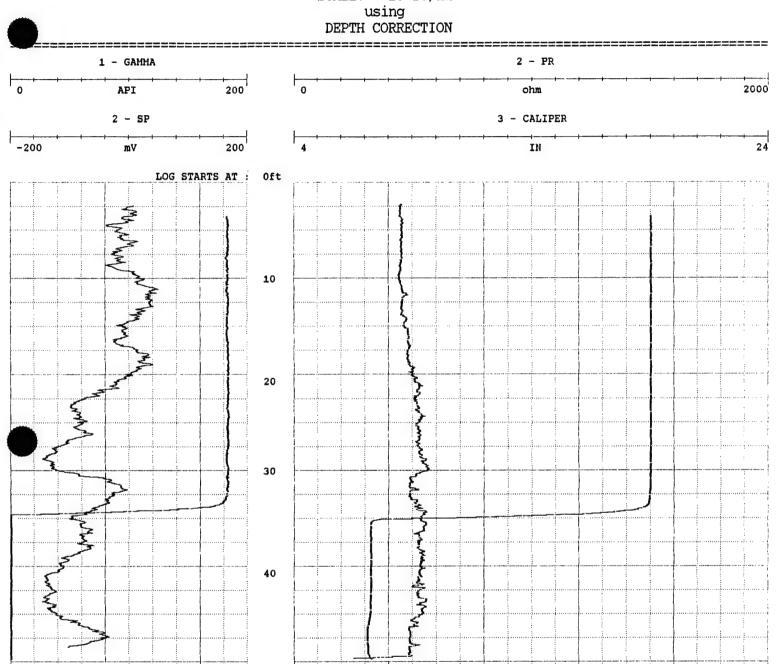
					***********			****
CHANNEL	***************************************	slope	offset	C1	C2	C3	C4	
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	API ohm mV IN	0.326204 -0.89890	0.000000 -569.169 3428.191 -6.11253	3568 0.000 0.000 3.100	0.000 1745 3814 2245	0.000 100.0 1500 6.100	0.000 2051 2145 2977	

Log No. FILENAME

1 - 2329B 2 - 2329B 3 - 2329B

LOG ENDS AT : 50ft





ASSOCIATED ENVIRONMENTAL IND.

HOLE NO: 2-393B

LOGGED: 07/30/97 at

TINKER AIR FORCE BASE SE/4 SEC 22 T11N R2W N/A COMPANY AREA PROSPECT N/A OPERATOR MM

DRILLING DATA

PERM DAT GL ELEV AZIMUTH INCLIN HOLE POS HOLE DIA 9" 50' DEPTH

CASING DATA

CASE-TYP CASE-DTH CASE-DIA CASE-THK

FLUID DATA

FLD TYPE FLUIDLEV AUGER

TD-3 LOGTYPE

TOOL: Tmrg--Merged Log Serial Number: 886 Spacer (cm) : n/a

CHANNEL DATA

CHANNEL	AMPLITUDE RAN	GE CHART	AREA	SENSOR OFFSET	FILTER	PLOT TYPE
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	0.0 - 200 0 - 20 -200 - 2 4.00 - 24.	00 12.0 - 00 0.0 - 00 12.0 -	- 10.0 - 32.0 - 10.0 - 32.0	328 453 453 318	4.0 0.0 0.0 0.0	4 2 2 2 2

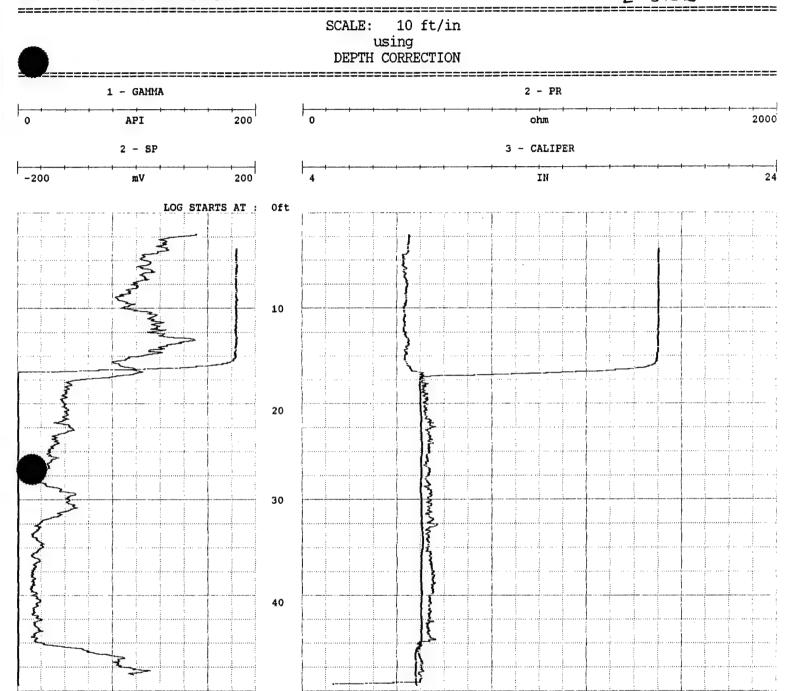
Calibration Data

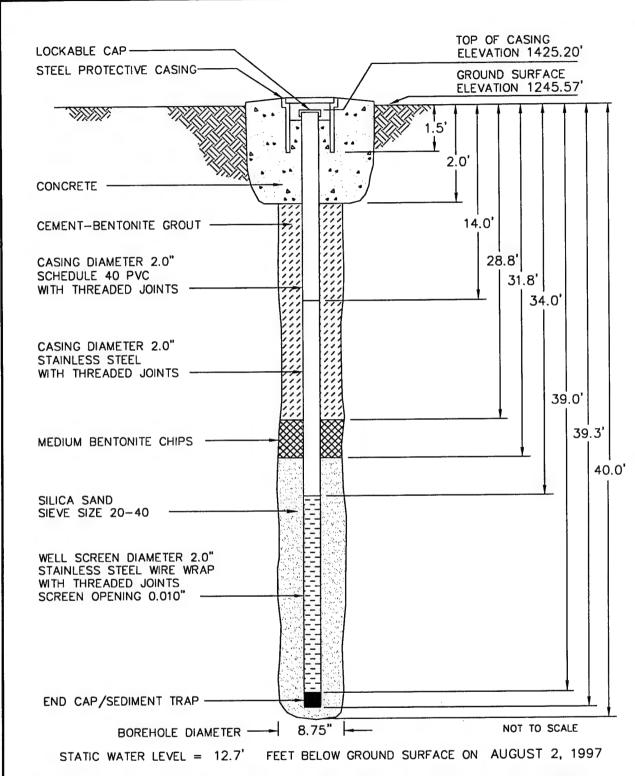
CHANNEL		slope	offset	C1	C2	C3	C4
1 - GAMMA 2 - PR 2 - SP 3 - CALIPER	API ohm mV IN	0.326204	0.000000 -569.169 3428.191 -6.11253	0.000 0.000 3.100	1745 3814 2245	0.000 100.0 1500 6.100	0.000 2051 2145 2977

Log No. FILENAME

1 - 2393B 2 - 2393B 3 - 2393B 13

LOG ENDS AT : 49ft



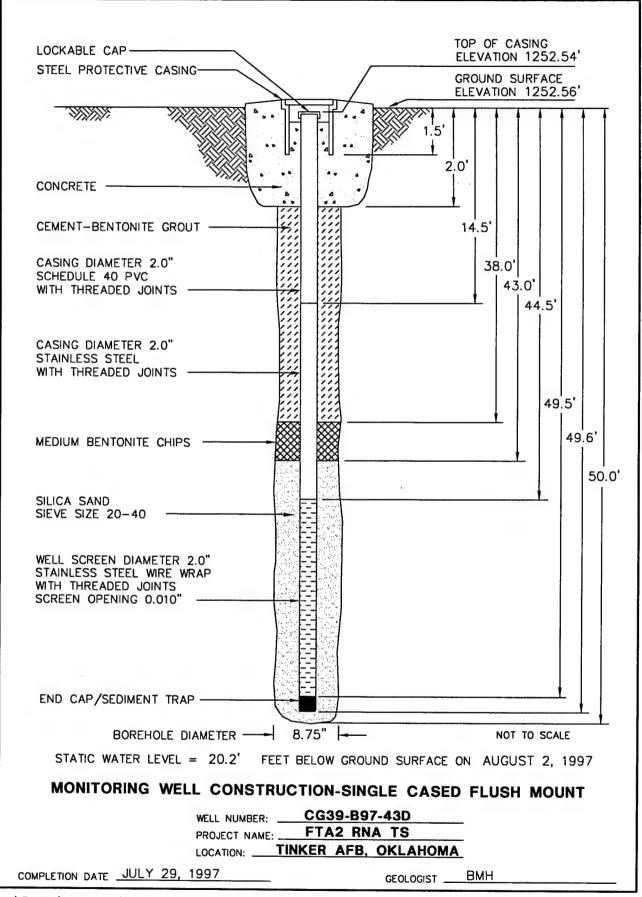


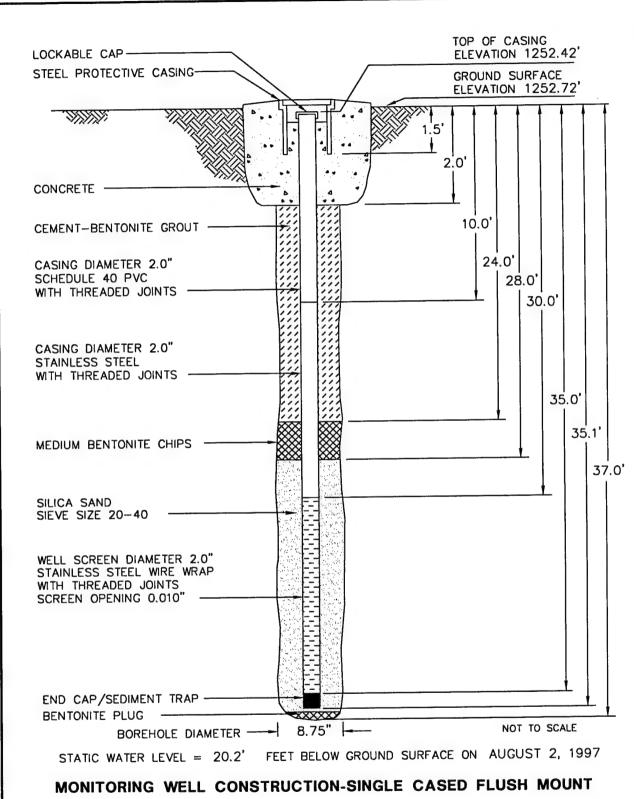
MONITORING WELL CONSTRUCTION-SINGLE CASED FLUSH MOUNT

WELL NUMBER: CG39-B97-41
PROJECT NAME: FTA2 RNA TS
LOCATION: TINKER AFB, OKLAHOMA

COMPLETION DATE JULY 28, 1997

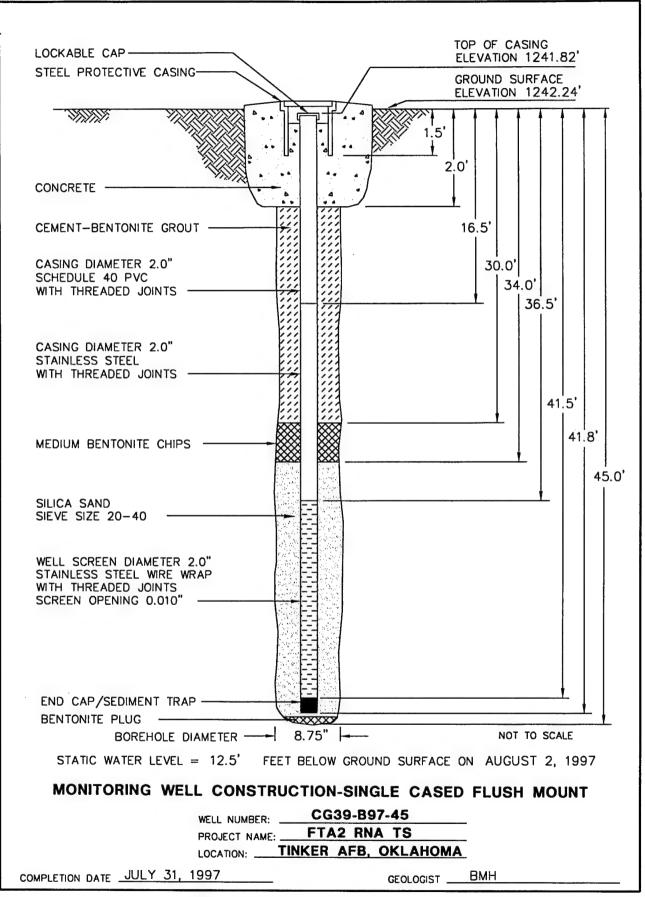
GEOLOGIST BMH

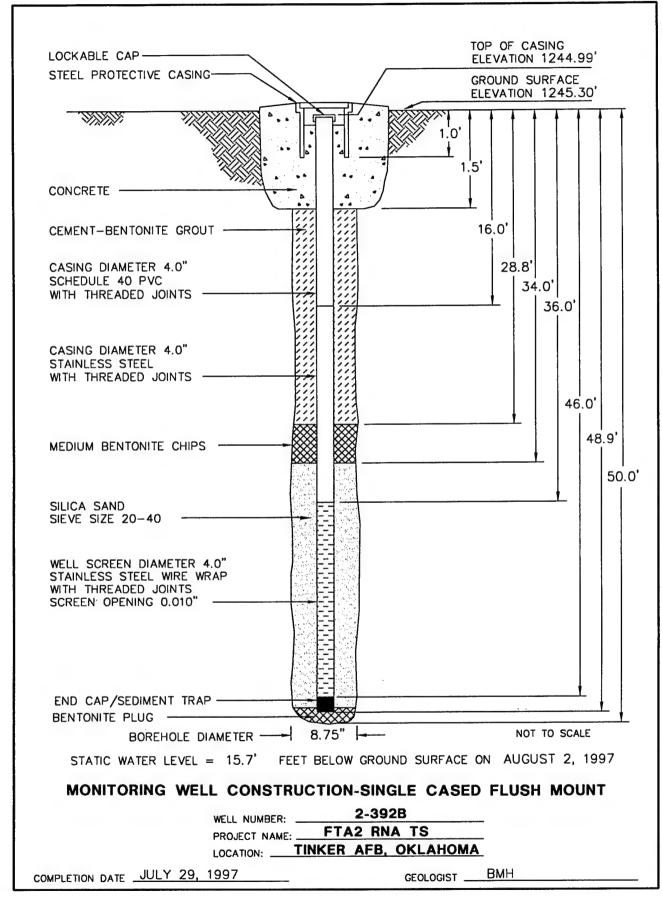


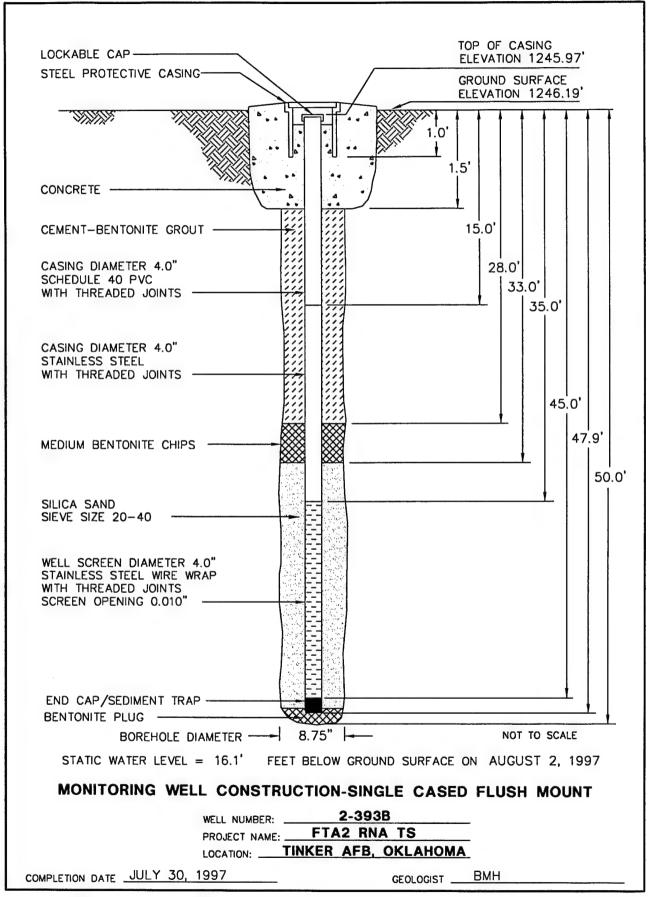


WELL NUMBER: CG39-B97-43S
PROJECT NAME: FTA2 RNA TS
LOCATION: TINKER AFB, OKLAHOMA

COMPLETION DATE JULY 30, 1997 GEOLOGIST BMH







	Project Nan Project Nun			AFB- FTA	2			
	Well Design	ation:	C639-	397-418				
	Developme	nt Method:	bail	pump sur				
	,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,	***************************************	***************************************		2			***************************************
Bailer Volu		NA		_) ² (L)(PI)(7.48			
Total Well	Depth (TD):	39 '		_ Water Colu	ımn (TD-WL): 24.7	1	
Static Wat	er Level (WL)	12.3	1	_ Casing Dia	meter:	2.0"		
Casing Volume: 4.35 gallons				_ CV =(CID/2	24) ² (TD-WL)	(PI)(7.48) g	allons/casin	g
Bails/Casing Volume: CV/BV								
Developm	ent Date: _&	11197						***************************************
	Casing No./	Temp.		Electrical	Dissolved	Redox	Turbidity	Visual
Time	Bailer No./	ت در	pН	Conductivity		Potential	(ntu)	Appearance
	Gallons			uslem	(mg/L)	(mv)		
1505	So	19.6	7,04	824	6.40		250	Cloudy
1238	100	19.5	7.18	82			6	Clear
१८८८	120	20.3	7.14	817	-		1.5	Ciear
				 				
							····	
				<u> </u>				
				-				
Total Volume Removed: ~34 (casing volumes) 150 (gallons) Weather: Suny and Hot, 905°F								
Weather: Notes:				y EPA Pa	sens ES		-	
Developed	і Ву:	BMH	AEI	Corp.		•		

	Project Nam Project Num			4FB - FTA 2 1. 34020	•			
	Well Design	ation:	C6 39	- 7897 - 435				
	Developmer	nt Method:	bail	pump surg	ge airlift	other		
	***************************************		***************************************		2		••	
Bailer Volu	ıme:	N ^p	\	- ` '	² (L)(PI)(7.48			
Total Well	Depth (TD):	3s'		Water Colu	mn (TD-WL): 15.1	,	-
Static Wat	er Level (WL)	19,9	<u>.'</u>	Casing Dia	meter:	"ت.2		•
Casing Vo	lume:	2.5 9	allens	CV =(CID/2	24) ² (TD-WL)	(PI)(7.48) ga	allons/casin	g
Bails/Casing Volume: CV/BV								
Developm	ent Date:8	314197						,
	Casing No./	Temp.		Electrical	Dissolved	Redox	Turbidity	Visual
Time	Bailer No./	ور_	pН	Conductivity		Potential	(ntu)	Appearance
	Gallons			uslem	(mg/L)	(mv)	` '	
2290	50	(9.0	7.07	1325	_	_	_	Clusely
1013	90	18.7	7.18	12105	_	-	-	Clardy
1025	125	18.5	7.20	1273			82.7	SI. Cludy
1035	150	18.2	7.24	1271			11.5	Clear
1042	175	18.2	7.27	1502			2.3	Clear
	me Removed:					175+		
Weather:	Suny	م لىص	m					
Notes:	Suny	Sampud	by Pa	sens Es EPI	a (See GW	Super For	ns)	
Developed	в Ву:	ТЗМН	AEI	C∞ρ.				

	Project Nan Project Nun			AFB - FA	2_			
	Well Design	ation:	C639	- B97 -43c	1			
•	Developme	nt Method:	bail	oump su	other			
Bailer Volu	ıma'	NA.		BV=(ID/2	4) ² (L)(PI)(7.48	R) gallons/ha	ilor	
					, , , , , ,		ilici	
Total Well Depth (TD): 44.5' Static Water Level (WL): 20.2'			5' 2'	_ Vvater Co _ Casing Di	lumn (TD-WL ameter:	2.0"		- -
Casing Vol	lume:	4.8	galluns.	CV =(CID	/24) ² (TD-WL)	(PI)(7.48) g	allons/casin	g
Bails/Casing Volume: CV/BV								
Developme	ent Date:&							
	Casing No./	Temp.		Electrical	Dissolved	Redox	Turbidity	Visual
Time	Bailer No./	ع_	pН	Conductivity	y Oxygen (mg/L)	Potential (mv)	(ntu)	Appearance
0940	2.5	19.5	7.62	1115	-		_	Closely
1005	75	(8.8)	1,58	1085	3-75	_	_	15
1020	125	18.2	7.49	1501	4.40		-	Single Hy Chiery
1030	175	18.4	7-41	1073	4.50	_	32.6	Vew St. Closery
1048	200	18.4	7.40	1074	4.55	_	6.4	Clear '
				ļ				
						ļ		
				-				
			-	- 				
Total Volu	me Removed:		42	_ ` _ `	ımes)	200+	(gallons)	
Weather:	_ Clear	Sumy	Hot.	90°F's		4 80 00		
Notes:	Samp	red aft	er pri	jing by E	PA Pascus	ES		
Developed	i By:	Bmit	AEI	Corp.		-		

	Project Nan Project Nun			AFB- FTA	12			
	Well Design	ation:	C639	- 397-45				
	Developme	nt Method:	bail	pump surg	ge airlift	other		
Bailer Volu	ma:	١٥		BV=(ID/24)	² (L)(PI)(7.48	l) nallons/ha	iler	
	•			- '				
Total Well Depth (TD): 42' Static Water Level (WL): 12.1'			,	_ Vvater Colu _ Casing Dia	mn (TD-WL meter:): 29.9 2.0°		
Casing Vol	ume:	4.9	gallens	_ CV =(CID/2	24) ² (TD-WL)	(PI)(7.48) g	allons/casin	g
Bails/Casing Volume: CV/BV								
Development Date:ちlчเจつ								
	Casing No./	Temp.		Electrical	Dissolved	Redox	Turbidity	Visual
Time	Bailer No./	0	pН	Conductivity	Oxygen (mg/L)	Potential (mv)	(ntu)	Appearance
1115	30	18.8	7.27	902	_	-	_	Closely Braun
1140	100	17.6	7.31	હ્યવ		•	85	SI. Clarry
1150	130	17.5	7.30	පිංහි		_	25.9	SI. Clarry
1500	175	17.5	7.32	BSY		(5.4	Clear
Total Volur	ne Removed:		~36	_(casing volun	nes)	175	(gallons)	
Weather:	Suny	arbt -	90's " F					
Notes:	Sompred	a Purped	by Epp	1 Parsons ES	>		•	
Developed	Ву:	Bm+ 1	AEI G	~ ~				

	Project Nan Project Nun	nber:	7290	AFS - FTAZ				
	Well Design	ation:	2-39	28				
	Developme	nt Method:	bail	oump sur	ge airlift	other		
***************************************	***************************************	***********************	••••••	•••••••••••	***************************************	***************************************	•••••	********************************
Bailer Vo	lume:	<u>\</u>		_ BV=(ID/24)) ² (L)(PI)(7. 4 8) gallons/ba	iler	
Total We	ll Depth (TD): ater Level (WL)	4	6'	_ Water Colu	_			
Static Wa	ater Level (WL)	15.	4'	Casing Dia	meter:	4.01	•	_
Casing V	olume:	ZC) gal.	CV =(CID/2	24) ² (TD-WL)	(PI)(7.48) g	allons/casin	ıg
Bails/Casing Volume: NA CV/BV								
			••••••			*******************	•••••	***************************************
Davalann	nent Date: _ 8	1. 1. 0.1	u 16 7					
Pevelohii			11.1					
T :	Casing No./	Temp.		Electrical	Dissolved	Redox	Turbidity	Visual
Time	Bailer No./ Gallons	نو	pН	Conductivity	Oxygen (mg/L)	Potential (mv)	(ntu)	Appearance
0912	80	22.1	7.82	739	5.63	NR	NR	Muddy Braw
1425	200	21.3	7.360	680	5.50	_	105	SI. CICLELY
1435	2.10	-	_	_	-	_	40	clear
1655	250	19.5	7.32	679	6.5		72_	51. Clary
1730	300	19.4	7.31	6 කිය	6.7		10,9	clear '
						-		
				-				
							L	<u></u>
Total Vol	ume Removed:		15	_(casing volun	nes)	3∞	(gallons)	
Weather:	Clear- 9	Pently CI	نفادل .	Hot				
Notes:		•	• .		. d	(ı	. la au . 2 laca	دا نا سادی
	Well P	mpecs c	304 0	20-40 gall	us crepe	miny an	have 1	30 min to
***************************************				M		د	cupie f	eet at Sta
Develope	ed By:	BMH.	AEI	Carp		De	relepted :	several tru
•	-						co I	-1. 150

	Project Nan Project Nun			AFB- FTA 11.34020	2				
	Well Design	ation:	2-393	3					
	Developme	nt Method:	bail	pump> surç	ge airlift	other			
		•••••••	••••••••••••••••••	DV 4D404	24 245027 46	······································			
Bailer Volu		AN_		- ` '	² (L)(PI)(7.48				
Total Well	Depth (TD):	(TD): Water Column (TD-WL): ZS, I'							
Static Water	er Level (WL)	15.	91	_ Casing Dia	meter:	4.0"			
Casing Vol	ume:	14.3	gal.	_ CV =(CID/2	24) ² (TD-WL)	(PI)(7.48) g	allons/casin	g	
Bails/Casin	g Volume:	AM.		_ CV/BV					
Developme	ent Date:	814147							
	Casing No./	Temp.		Electrical	Dissolved	Redox	Turbidity	Visual	
Time	Bailer No./ Gattons	<u>«</u>	pН	Conductivitys (cm	Oxygen (mg/L)	Potential (mv)	(ntu)	Appearance	
1240	50	18.6	7.31	1202	-	_	_	Closely Brein	
idos	200	18.2	7.60	11.30	1	•	4=	45.5	
1415	250	17.8	7.49	1185	~			SI. Cluxy	
1430	300	17.5	7.39	1179	_		2	- 40,8 cm+	
1445	350	17.5	7.38	1180	_		<	- lo.lo	
Total Volum	ne Removed:	14.	71	(casing volun	nes)	350+	(gallons)		
				_	,		. 19		
vveatner: Notes:	Clear to	Party (1	curly, Ho				•		
***************************************								***************************************	
Developed	Ву:	73mH]	ACT (ορ.		•			

GROUNDWATER SAMPLING RECORD

Sampling Location: <u>Tinker AFB - FTA2</u>

	Sampling Dates:	(31/97)
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL	2B
REASON FO DATE AND SAMPLE CO WEATHER:	OR SAMPLING: [Regular Sampling; [] Special Sampling; TIME OF SAMPLING: 7/31/47 at 16/26 a.m./p.m. OLLECTED BY: CH/DR of Parts of EPTH.	(number)
MONITORI	NG WELL CONDITION: [] LOCKED: WELL NUMBER (IS - IS NOT) APPARENT STEEL CASING CONDITION IS: 5000	D
	STEEL CASING CONDITION IS:	
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH DT who	
2[]	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	WATER DEPTH 14.5' btc. Measured with:	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Turbidity: Odor: Other Comments:	
4[]	WELL EVACUATION: Method: decl: decl gran for Volume Removed: Observations: Turbidity (clear) slightly cloudy Water level (rose fell no change) Water odors:	very cloudy)

Groundwater Sampling Record

Monitoring Well No	· <u>a-6213</u>	(Cont'd)
--------------------	-----------------	----------

5 []] SAMP	LE EXTRAC	CTION ME	THOD:			
		[] B [/] P [] C	Bailer made lump, type:_ Other, descri	of:	grunses		
		Samp	le obtained	is [X] GRA	B; [] COMF	POSITE SAMPLE	
6[]	ON-SI	TE MEASU	REMENTS				
	Time	16:28	16:30	16:33	6:36		Measured with
	Temp (°C)	18.8	15.9	18 8	18.8		
	pН	7.26	7.33	7.35	7.34		
	Cond (µS/cm)	1433	1437	1437	1941		
	DO (mg/L)	0.2	0.2	0.3	0.2		
	Redox (mV)	62.5	55.8	52.7	50.7		
,	Salinity						
7[]		TE SAMPLE	E TREATM	ENT: Method Method	<u>ت</u>	Containers:	
	[]	Preser	vatives add		Isuric acid I:un phoiphats	Containers: / Containers: / Containers: / Containers: /	oboral 2 serun
9[]	1 CONT	'AINER HAI	NDLING:				
· []	,	[/] []	Container Container	Sides Labeled Lids Taped Placed in Ice			
10 [) OTHE	R COMMEN	NTS:				·

GROUNDWATER SAMPLING RECORD

	Sampling Location: Tinker AFB - FTA2 Sampling Dates: 7/31/97		
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL 2-63 A		
REASON FO DATE AND SAMPLE CO WEATHER: DATUM FO	(number) OR SAMPLING: [1] Regular Sampling; [1] Special Sampling; TIME OF SAMPLING: 1/21/5/ at 1/41/05 a.m./p.m. DLLECTED BY: (seis Have of Paise of 65.1/25 Example) Sample Control		
	WELL CONDITION: LOCKED: WELL NUMBER (IS- IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS- IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):		
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH -> used dedicated situates pump. Items Cleaned (List):		
2[]	PRODUCT DEPTHFT. BELOW DATUM Measured with:		
	WATER DEPTH 55/55 54.8 btc FT. BELOW DATUM Measured with: depth refer		
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:		
4[]	WELL EVACUATION: Method: dedicated grantes Volume Removed: 7 callens Observations: Turbidity (clear slightly cloudy very cloudy) Water level (rose fell no change) Water odors:		

Groundwater Sampling Record

Monitoring Well No. 2-63 (Cont'd)

5[]	SAMPI	LE EXTRAC	TION MET	HOD:				
		[] B	ailer made of	f:				
		[TPI	imp, type:	closel.	co 1201	g runso	ſ	
		[] 0	ther, describe	e:				
		Sample	e obtained is	[X] GRAE	; [] COM	IPOSITE SA	MPLE	
	0)1 01	DE ME ACUTE	EL CELEC.					
6[]	ON-SI	IE MEASUR	EMEN 15:					
	Time	14:05	11:09	14:14	14:20			Measured with
	Temp (°C)	18.8	19.0	18.9				
	рН 7.21	7-2110	72735	7.2 stat	7.31			
	Cond (µS/cm)		660	655	654			
	DO (mg/L)	7.0	7.2	7.2	7.2			
	Redox (mV)	17.1	29.5	_	47.8			
	Salinity							
8[]	ON-SIT		on: M N vatives added M N	Method	the acid	Contair Contair Contair Contair Contair Contair	ners: 2 seemers: 4 - 1	m. 1 2507-12L
9[]	CONT	AINER HAN		zemou				
		ij	Container Si Container Li Containers P		Chest			
10 [) OTHE	R COMMEN	TS:		•			
[

GROUNDWATER SAMPLING RECORD

	Sampling Location: Tinker AFB - FTA2 Sampling Dates: 7/31/97
GROUNDW	VATER SAMPLING RECORD - MONITORING WELL 2 - 6 3 B (number)
DATE AND SAMPLE CO	OR SAMPLING: [x] Regular Sampling; [] Special Sampling; OTIME OF SAMPLING: 7/3/97 at 13:15 a.m./p.m. OLLECTED BY: 5/64 of 45664/Perces 65
	OR WATER DEPTH MEASUREMENT (Describe): 700 0 0.3 54 hgs
	NG WELL CONDITION:
	INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS- IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):
Check-off 1 [x]	EQUIPMENT CLEANED BEFORE USE WITH -> ded : and grantes pump Items Cleaned (List):
2 [٨]	PRODUCT DEPTHFT. BELOW DATUM Measured with:
	WATER DEPTH TOC - 12.1 (12.4' 655) FT. BELOW DATUM Measured with: Depth meter FT. BELOW DATUM
3 [x]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Turbidity: Odor: Other Comments:
4 [¢]	WELL EVACUATION: Method: Volume Removed: Observations: Turbidity (Clear) Water level (rose fell no change) Water odors: Water odors:
	Other comments:

Groundwater Sampling Record

Monitoring Well No. 2-63 B (Cont'd)

	SAMPI	LE EXTRAC	TION MET	HOD:				
		[] Ba [] Pu [] Ot	ailer made o imp, type: ther, describ	fi 1 Grun fa e:	د			
		Sample	e obtained is	X GRAI	3; [] COMP	OSITE SAM	PLE	
					,			
]	ON-SIT	TE MEASUR	REMENTS:					
Time		13:35	13:38	13:41	13:43		Measured wi	th
Temp (°	°C)	19.4	19.9	19.9	19.9		Trien 8	40
pН		_	_	7.33	7.35		Ocan 2	50 8
Cond (µ	S/cm)	15/3	1516	1517	1516		Or. on 120	2
DO (mg	/L)	0.5	0.4	0,4	0.6		Orien g	10
Redox (mV)	129.2	100.2	84.6	79.4		Otion 33	50 A
Salinity								
_					2 1:	50.ml- A	dups) of present / sulface acid /algore buttles!! of	Jul Furic
1		TE SAMPLE	TREATME			_ Container _ Container	Sulfaria ballies (1 u/	
1	ON-SIT	TE SAMPLE	TREATME	ENT: Method Method Method		_ Container _ Container	s:s:	
]	ON-SIT	TE SAMPLE	TREATME	Method Method Method d:	Surve encicl	Container Container Container Container Container Container Container	s:s:	1.50-~L
	ON-SIT	TE SAMPLE	TREATME	Method	Surve encicl	Container Container Container Container Container Container Container	s:	1.50-~L
]	ON-SIT	Filtrati Preserv	TREATME ion: I Vatives added I I I I I I I I I I I I Container S. Container L	Method Me	Sirie ewich	Container Container Container Container Container Container Container	s:	1.50-~L
	ON-SIT	Filtrati Preserv	TREATME ion: I vatives adde i I I I I I I I I I I I I	Method	Sirie ewich	Container Container Container Container Container Container Container	s:	1.50-~L

	Sampling Location: <u>Tinker A</u> Sampling Dates:	
GROUNDW	VATER SAMPLING RECORD - MONITORING WELL $\frac{2-64}{1}$	
DATE AND SAMPLE C	OR SAMPLING: [X] Regular Sampling; [] Special Sampling; Description of the sampling; at the sampling; and the sampling;	(number)
————	N WATER DEI III MEASUREMI (Describe).	
MONITOR	ING WELL CONDITION: [] LOCKED: [] LOCKED: WELL NUMBER (IS IS NOT) APPARENT STEEL CASING CONDITION IS: Good	
	INNER PVC CASING CONDITION IS:	
Check-off	EQUIPMENT CLEANED BEFORE USE WITH	
2 [4	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	WATER DEPTH 57.37 Lt a Measured with: water mater	FT. BELOW DATUM
3 [4	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: 51.gh7 orange color Turbidity: 122 - neel:222 Odor: 2022 Other Comments:	
414	WELL EVACUATION: Method: Volume Removed: Observations: Turbidity (clear slightly cloudy Water level (rose fell no change) Water odors: Other comments:	very cloudy)

		1			ampling R 2-6419		
5 [-	SAMPI	LE EXTRAC	TION MET	HOD:			
					3; [] COM	POSITE SAMPLE	
ا 6	ON-SI	TE MEASUR	EMENTS:				
		1.3	2.5	4.0	6.0		Measured with
				0842	0850		
	Temp (°C)	19.7	20.9	20.4	20.2		751 55
	Cond (µS/cm)	7.74	7.69	7.68	7.69		Orienzisch
	DO (mg/L)	665	667	७५८	666		Crich 122
	Redox (mV)	6.41	6.09	6.02	6.05		
	Salinity	NR	NAS.	125,3 NR	121.9 NR		Oyen 250A
7 [-	SAMPI	LE CONTAIN	NERS (mate	rial, number		WAS Synun Jas 250 Wl. Plastic	
8 [-		TE SAMPLE				Containore	7
[]		Filtration; of		Method 14 Method Seek on Prospers Method		Containers: S Containers: F	LOAL Exact Transporter
	[]	Preserv	atives adde	1 :			
			1	Method Method Method Method		Containers:	
9 [-	CONT	AINER HAN	DLING:				
		i i	Container L	ides Labeled ids Taped Placed in Ice	Chest		

OTHER COMMENTS: Scupes Hardled by EPA

10 [4]

Sampling Location: <u>Tinker AFB - FTA2</u>

	Sampling Dates: S/1	
GROUNDWA	ATER SAMPLING RECORD - MONITORING WELL 2-643	
REASON FO DATE AND SAMPLE CO WEATHER:_	R SAMPLING: [b] Regular Sampling; [] Special Sampling; IIME OF SAMPLING: 9/1/4 at a.m./p.m. 2810 78 LLECTED BY: 6H/cH of Persons 65.	(number)
DATUM FOR	R WATER DEPTH MEASUREMENT (Describe): 40 water dep	the measurement
	IG WELL CONDITION: [] LOCKED: [V UNLOCKED	out)
	WELL NUMBER (IS IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS - IS NOT) APPARENT DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR MONITORING WELL REQUIRED REPAIR (describe):	
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH DI water Items Cleaned (List):	
2[]	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	WATER DEPTH rose (broken roter) Measured with:	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:	
4[]	WELL EVACUATION: Method: Volume Removed: Observations: Turbidity (clear) Water level (rose fell no change) Water odors: Other comments:	very cloudy)

Groundwater Sampling Record

Monitoring Well No.	2-643	(Cont'd
---------------------	-------	---------

[]		SAMP	LE EXTRA	CTION ME	ETHOD:				
			[] I [/] I	Bailer made Pump, type: Other, descr	of:	ited grun	fus		
	,					B; [] COMPO	SITE SAMPL	Æ	
[]			TE MEASU		6,0				
Γ	Time	سو_رده.	0743		0806			Mea	sured with
ŀ	Temp (°C)	19:3	20.2	206			YSI	55
Ì	pН		Ne	NP.	NR				
Ì	Cond (ıS/cm)	853	848	548			01.	JA 122
Ī	DO (mg	g/L)	2.33	2.10	1.54			Ysj	5 55
Ī	Redox ((mV)	123.4	Ju. 3	123.5			Or:	on 250A
ſ	Salinity		NR	NR	MR				
[]		[]	TE SAMPL		Method		Containers:_		
		M	Prese	rvatives add	Method 5. Method 5.	elfuric coid	Containers:_ Containers:_	4 VOF.	
[]		CONT	AINER HA	NDLING:					
			[4] [1] [1]	Container	Sides Labeled Lids Taped s Placed in Ice				
] 0	1	OTHE	R COMME	NTS:					
~ L	•								

Sampling Location: <u>Tinker AFB - FTA2</u>

Sampling Dates: <u>δ/ι/ 9 </u>	
GROUNDWATER SAMPLING RECORD - MONITORING WELL &-65 A	
	(number)
REASON FOR SAMPLING: [Regular Sampling; [] Special Sampling;	
DATE AND TIME OF SAMPLING: $\frac{1}{2}$	
SAMPLE COLLECTED BY: Chi Dk of Parsons (USEP 17	
WEATHER. 384M/	
DATUM FOR WATER DEPTH MEASUREMENT (Describe): TOC	
MONITORING WELL CONDITION:	
[] LOCKED: [/] UNLOCKED	
WELL NUMBER (IS - IS NOT) APPARENT	
STEEL CASING CONDITION IS:	
STEEL CASING CONDITION IS:	
WATER DEPTH MEASUREMENT DATUM (15) - IS NOT) APPARENT	***************************************
 DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR MONITORING WELL REQUIRED REPAIR (describe): 	
[] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off	
Items Cleaned (List):	
items cicalica (bist).	
2 [] PRODUCT DEPTH have	FT. BELOW DATUM
Measured with:	
WATER DEPTH 61.5' btc	FT. BELOW DATUM
Measured with: water nater	
3 [] WATER-CONDITION BEFORE WELL EVACUATION (Describe):	
Color: orange	
Turbidity: 1:54	
Odor: none	
Other Comments:	
4 [] WELL EVACUATION:	
Method: declicated gransos	
Volume Removed:	
	ery cloudy)
Water level (rose fell no change)	
Water odors: none	
Other comments:	

Groundwater Sampling Record

Monitoring Well No. 2 - 65 A (Cont'd)

5 []	SAN	IPLE EXTRA	CTION MET	HOD:				
		[] E [/j P [] C	Bailer made o Pump, type: Other, describ	f:	d grunse	o\$		
		Samp	le obtained is	[X] GRA	B; [] CON	IPOSITE SAM	PLE	
6[]	ON-	SITE MEASU	REMENTS:					
о	volume(gal	SITE MEASU J.O /352	5.0	7.2				
	Time	1352	1402	1407			Measure	ed with
	Temp (°C)	19.1	19.9	19.9			Orion.	
	pН	7.69	7.46	7.67			Or:00 V	250A
	Cond (µS/cm	535	577	576			Orion	122
	DO (mg/L)	4.4	4.2	4.2			01:00	£40
	Redox (mV)	11.4	-45.0	-38.3			Orien	250A
	Salinity	NR	NR	NR				
8[]	ON- []	SITE SAMPLI Filtrat	tion: I	Method Method Method		Container	s:s:s:	
	.,]]]	Method <u>54</u> Method <u>500</u>	Ifuric acid	Container Container Container Container	s: 1 125-ml, s: 4 voA, s: 5:	2 serum
9[]	CON	TAINER HAI	NDLING:					
		[]	Container L	ides Labeled ids Taped Placed in Ice				
10 [] OTH	IER COMMEN	NTS:					

Sampling Location: <u>Tinker AFB - FTA2</u>

	Sampling Dates: S/1/97	
GROUNDWA	ATER SAMPLING RECORD - MONITORING WELL $\frac{2-65}{3}$	
DATE AND SAMPLE CO	R SAMPLING: [1] Regular Sampling; [1] Special Sampling; IME OF SAMPLING: 5/1/97 at 1305 a.m./p.m. LLECTED BY: CH/DR of Decisions & S/USEPH Sunny R WATER DEPTH MEASUREMENT (Describe): 70C	(number)
MONITORIN	G WELL CONDITION: [] LOCKED: WELL NUMBER (S) IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (S) IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH Kim-W:pc Items Cleaned (List): Probes	
2[]	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	WATER DEPTH 17.2 btc Measured with: Leter meter	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:	
4[]	WELL EVACUATION: Method: ded:cated grantes Volume Removed: ~15 sallons Observations: Turbidity (dear) slightly cloudy Water level (rose fell no change) Water odors: Other comments:	very cloudy)

Groundwater Sampling Record

Monitoring Well No. 2 ~ 6 5 3 (Cont'd)

5[]	SAMP	LE EXTRAC	CTION ME	THOD:					
		[]B [4]P []O	ailer made ump, type:_ ther, descri	of:	ted g	runfos			
		Sampl	e obtained	is [X] GRAE	; [] COM	POSITE S	AMPLE		
6[]	ON-SI	TE MEASUI	REMENTS:	15.0					
	wolumelgal)					ſ		Measured with	
	Time (9C)	1305		1335				Wicasured With	
	Temp (°C)	19.8	18.7						
				7.13					
	Cond (µS/cm) DO (mg/L)	877	966	2.3					
		0.3							
	Redox (mV)	100.4	93.0	91.2					
	Salinity	NR	NR	NR					
7[]	SAMP	LE CONTAI	NERS (mai	terial, number,	size):				
8[]	ON-SI	TE SAMPLE	TREATM	ENT:					
	[]	[] Filtration:		Method Containers: Method Containers: Method Containers:					
	M	Preser	vatives add	ed:					
				Method Social Method Method Method	ium phosp	Contai	ners: 4 ners: / ners: ners:	UDAS 125.ml, 2 Secum	
9[]	CONT	AINER HAN	IDLING:						
		į į	Container	Sides Labeled Lids Taped Placed in Ice	Chest				
10 [] OTHE	R COMMEN	TTS:						

	Sampling Location: Tinker AF Sampling Dates: Éjece	
GROUNDWA	ATER SAMPLING RECORD - MONITORING WELL Z-142 A	
REASON FO DATE AND SAMPLE CO	R SAMPLING: [1] Regular Sampling; [] Special Sampling; TIME OF SAMPLING: 8/2/57 at 1325 a.m./pm DLLECTED BY: DMIT/CIT of Passons Clear, Freezy, 1407 900 F	(number)
DATUM FOR	R WATER DEPTH MEASUREMENT (Describe): Top Dedicated F	mp
MONITORIN	IG WELL CONDITION: [] LOCKED: WELL NUMBER (S)- IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS:	
	WATER DEPTH MEASUREMENT DATUM(IS) IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off 1 [-	EQUIPMENT CLEANED BEFORE USE WITH DI water Items Cleaned (List): Prefree	
214	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	Measured with: Somet level Indicator	FT. BELOW DATUM
3[丁	Pure Ucil. = 55 ga). WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Reachish Brown Turbidity: Shanty Clearly Odor: Dom Other Comments:	
4[\$	WELL EVACUATION: Method: Dedicated Councilor Range Volume Removed: So gallow Observations: Turbidity (clear) slightly cloudy Water level (rose fell no change) Water odors: None Other comments:	very cloudy)

				undwater S Well No					
~ -	- 041	ADI E EVED			2-14274	(Cont a	,		
5 [~	SA		ACTION ME						;
		[]	Bailer made Pump, type:_	of:	1	•			
		[7]	Other, descri	be:	rea Crowd	165			
		Sar	nple obtained i	is [X] GRA	B; [] CON	APOSITE SA	MPLE		
6 [~	ON								•
	Time	3.0		22	35	45	56	Measured with	
	Temp (°C)	1500	1226	1240	1300	19.1	1320		
	pH	18.9		17.9	7.49	7.75	18.9	07 EX 32 SY	
	Cond (µS/cn	7・53 n) 7いマ	7.70 679	7.68	626	613	620	Orion 122	
	DO (mg/L)	0.62		0.59	1.26	1.34	1.20	48122	
	Redox (mV)		NR	NR	NR	24.5	24.3	CYENZEUA	
	Salinity	NR	WR	NR	NR	Nr.	NR	MR	
							ed Amp Do		
7 [=	- SA	MPLE CONT	'AINERS (mat	terial, numbe	r, size):2	2000	*		
						5 520 V	nl. Plastu		
8 [-	+ ON	-SITE SAME	LE TREATM	ENT:					
υį	•					a			
	[]	Filt	ration:	Method			Containers:		
				Method		Contain	ners:		
	עז	/ Dro	servatives add	ed:					
	H	FIC							
				Method Su	Afune Aciel	Contain	ners: Sym	m Jors, France Ivans	Cures
				Method Method	cun Phospha	Contair	ners: <u>CC/</u>	HÇ	
				Method					
0 F	,	NTAINER H	ANDI DIC.						
با 9	- co	NI AINEK H	ANDLING:						
				Sides Labeled	i				
		[]		Lids Taped Placed in Ice	Chest				
		[-7							
10	OT OT	HER COMM	ENTS: EA	A will be	nelle San	hes Mend	en		

Sampling Location: <u>Tinker AFB - FTA2</u>

	Sampling Dates: 312197	
GROUNDW	VATER SAMPLING RECORD - MONITORING WELL 2-1423	
DATE AND SAMPLE C	OR SAMPLING: [1]—Regular Sampling; [] Special Sampling; O TIME OF SAMPLING: 8/2197 at 1445 a.m./pm. COLLECTED BY: BM CI of Pasau ES COLLECTED BY: BM CI OF Pasau ES COLLECTED BY: BM CI	(number)
MONUTOR	INC. WELL CONDITION.	
	ING WELL CONDITION: [] LOCKED: WELL NUMBER (Q2- IS NOT) APPARENT STEEL CASING CONDITION IS:	
	INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS) IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off	EQUIPMENT CLEANED BEFORE USE WITH Decolor leaned (List): Robes	
2 H	PRODUCT DEPTH Neasured with:	FT. BELOW DATUM
	WATER DEPTH 13.36' Top well Cap 73 = 42' Measured with: Solvast Purje Wille =	FT. BELOW DATUM . 56 gal.
3 [12]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:	
4[]	WELL EVACUATION: Method: Declared Counciles Puro Volume Removed: Observations: Turbidity (clear slightly cloudy Water level (rose fell no change) Water odors: Other comments:	very cloudy)

				ndwater S			
					2-1423	(Cont'd)	
5 [~	SAMPI	LE EXTRAC	CTION MET	HOD:			
		[] B	ailer made o	f. Decheuh		. 0	
		1 [k]	ump, type: ther_describ	Declicat	ed Com	dfus	
			ŕ				
		Sampl	e obtained is	[X] GRAI	B; [] CON	MPOSITE SAMPLE	
6 [-	ON-SI	ΓΕ MEASUF	EMENTS:				
· [.	,	10.0	25	35	4६	92	
	Time	1352	1408	1420	1430	1440	Measured with
	Temp (°C)	16.8	16.5	16.9	17.1	16.9	75I55
	pН	Me	7.90	7.41	7.35	7.36	Aroks ward
	Cond (µS/cm)	1438	1460	1480	1481	1480	Orien122
	DO (mg/L)	2.40	3.25	3.48	3.35	3.06	48154
	Redox (mV)	NR	7411	55.6	53.7	52.60	Charzia
	Salinity	NR	Ne	Ne	NR	me	NR
a r .	SAMP	E CONTAI	NIEDS (mote	rial number	cize):	2 40/4	
1 [4	SAIVIP	LE CONTAI	MERS (mate	mai, number	, 5120)	2 Symin	Sas
						Zx 250m	A Playtic
8 [-	ON-SI	TE SAMPLE	TREATME	ENT:			
	[]	Filtrat	ion:	Method		Containers:	
	1.1	Intac		Method			
			- 1	Method		Containers:	
				۸٠			
	1	Preser	vatives adde	u.			
	1	Preser	vanves adde	Mathad		Containers:	
	け	Preser]	Method		Containers:	
	[<i>}</i>	Preser]	Method Method Method		Containers: Containers:	
	[<i>}</i>	Preser]	Method		Containers: Containers:	
9[4		Preser AINER HAN]]]	Method Method Method		Containers: Containers:	
9[=		AINER HAN]]] IDLING:	Method Method Method Method		Containers: Containers:	

OTHER COMMENTS:____

10 [·]

	Sampling Location: Tinker AFI Sampling Dates: 7/3/	B - FTA2 97
	ATER SAMPLING RECORD - MONITORING WELL $3 - 62 A$	(number)
WEATHER:	OR SAMPLING: [Regular Sampling; [] Special Sampling; TIME OF SAMPLING:	(
	WITTER DEL TITIVE AUGUST (DESORTO).	
MONITORIN	WELL CONDITION: [] LOCKED: WELL NUMBER (IS- IS NOT) APPARENT STEEL CASING CONDITION IS:	
	INNER PVC CASING CONDITION IS: _ 5000 (S) - IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH DI with Items Cleaned (List): f'obs	
2[]	PRODUCT DEPTH	FT. BELOW DATUM
	WATER DEPTH 57.36 6/c Measured with: water pater	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:	
4[]	WELL EVACUATION: Method: Volume Removed: Under Company Water level (rose fell no change) Water odors: Other comments:	

5[]	SAMP	LE EXTRAC	TION ME	THOD:				
		[] B: [] P: [] O	ailer made o ump, type:_ ther, descri	of:	ted gran	n Fes		
		Sample	e obtained i	s [X] GRAI	B; [] COM	IPOSITE SA	MPLE	
6[]	ON-SI	TE MEASUR	REMENTS:					
	Time	17:15	17:17	17:20	17:23			Measured with
	Temp (°C)	20.3	21.7	21.5	21.5			
	pН	7.35	7.42	7.42	7.42			
1	Cond (µS/cm)	789	786	184	782			
	DO (mg/L)	5.2	5.3	5.5	5.6			
	Redox (mV)	104.7		100.4	97.2			
	Salinity							
8[]	[]		ion:	Method Method Method		Contain	ers:ers:	
	[]	Preser	vatives adde		Fusic zeil Lium pheif	Contain Contain Contain Contain	ers: 2 52 ers: 4 1/ ers:	1 125-m
9[]	CONT	AINER HAN	IDLING:					
		Ĺĵ	Container 1	Sides Labeled Lids Taped Placed in Ice				
10 [] OTHE	R COMMEN	TTS:					

	Sampling Location: <u>Tinker AFI</u> Sampling Dates: <u>F/1/97</u>	
GROUNDWA	ATER SAMPLING RECORD - MONITORING WELL 2-167 B	(market)
DATE AND SAMPLE CO	R SAMPLING: [1] Regular Sampling; [1] Special Sampling; TIME OF SAMPLING: [1] [1] at 11 [1] a.m./p.m. PLLECTED BY: CH D	(number)
MONITORIN	WELL CONDITION: [LOCKED: WELL NUMBER (IS IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH DJ weder Items Cleaned (List): probes	
2[]	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	WATER DEPTH	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Company 25h Turbidity: Company 25h Odor: None Other Comments:	
4[]	WELL EVACUATION: Method:	

Groundwater Sampling Record

Monitoring Well No. $2 - 10 + 15$ (Cont	nitoring Well No.	2-16/13	_ (Cont'
---	-------------------	---------	----------

5[]	SAMP	LE EXTRA	CTION MI	ETHOD:					
		[] [[] [Bailer made Pump, type Other, desc	e of: : ribe:	grunfog				
		Samp	ole obtained	is [X] GRAB; [) COMPOSIT	E SAMPLE			
6[]	ON-SI	TE MEASU		S: 3 1.0					
ſ	Time	1150	1211	1227			Measured with		
ı	Temp (°C)	17.9	17.2	17.3					
ı	pН	7.20	7.14	9.11					
	Cond (µS/cm)	472	991	1008					
	DO (mg/L)	3.1	2.8	a .3					
	Redox (mV)	104.2	76.3	80.6					
	Salinity	11/2	NR	NR					
8[]	ON-SI [] [√	4			Method Containers: Method Containers: Method Containers:				
				Method Sockium Method Sulfur Method Method	Co	ntainers:	VOD dr-ml a serum		
9[]	CONT	AINER HA	NDLING:						
		[]	Container	r Sides Labeled r Lids Taped rs Placed in Ice Che	est				
10 [) OTHE	R COMME	NTS:		•				

	Sampling Location: <u>Tinker AFB - FTA2</u> Sampling Dates: <u>7/31/97</u>
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL $\frac{2-165}{3}$
DATE AND SAMPLE C	(number) OR SAMPLING: [A Regular Sampling; [] Special Sampling; OTIME OF SAMPLING: 7/31/97 at 15:34 a.m./p.m. OLLECTED BY: DK/CH of 165 EP17/Portent (S) OR WATER DEPTH MEASUREMENT (Describe): 700 0 3'095
	NG WELL CONDITION: [L] LOCKED: WELL NUMBER (IS- IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS - IS NOT) APPARENT DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR
	[] MONITORING WELL REQUIRED REPAIR (describe):
Check-off	EQUIPMENT CLEANED BEFORE USE WITH Di water Items Cleaned (List): ~2 12 5
2[]	PRODUCT DEPTH
	WATER DEPTH 12.56' 45 5tc FT. BELOW DATUM Measured with:
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: for of any the form of the high order of the high of th
4[]	WELL EVACUATION: Method: declicated grantes Volume Removed: 50° calless Observations: Turbidity (clear slightly cloudy very cloudy) Water level (rose fell no change) Water odors: 2002 Other comments:
	Note: Had to purge 50+ gallens For turbidity to lower and water to clear

 $L:\tinker\workplan\forms\gwsample.doc$

Groundwater Sampling Record

Monitoring Well No. 2-1680 (Cont'd)

5[]	SAMP.	LE EXTRA	CTION MET	HOD:				
		[] E [/] P [] C	Bailer made o ump, type: Other, describ	f:	od <u>g</u> ra	N Frost		
		Samp	le obtained is	[X] GRAI	B; [] COM	POSITE S.	AMPLE	
6[]	ON-SI	TE MEASU	REMENTS:					
ſ	Time	15:34	15:37	15:42	15:48			Measured with
ı	Temp (°C)	18.6	18.3	18.2.	18,3			
ı	pH	9.15	7.18	7.18	7.17			
Ì	Cond (µS/cm)	988	95.4	985	993			
Ì	DO (mg/L)	0.8	0.7	0.6	0.7			
	Redox (mV)	43.1	2.1.3	-5.8	15/28 -2	4.2		
	Salinity							
8[]	ON-SI	ΓΕ SAMPLI Filtra]	Method		Contai	ners:	
	[]	Preser	rvatives adde		Farl acie lium phoiphs	Contai Contai Contai	iners:	250-ml 2 Servi
9[]	CONT	AINER HAI	NDLING:					
		[/] []	Container L	ides Labeled ids Taped Placed in Ice	Chest			
10[] OTHE	R COMME	NTS:					

	Sampling Location: <u>Tinker AF</u> Sampling Dates: <u> </u>	
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL $2-272$	
DATE AND SAMPLE CO WEATHER:	OR SAMPLING: [Regular Sampling; [] Special Sampling; TIME OF SAMPLING: 84 1/97 at 10:00 a.m./p.m. OLLECTED BY: CH/OU of Parsons / USEPH Sunny R WATER DEPTH MEASUREMENT (Describe): 700	(number)
MONITORII	NG WELL CONDITION: [
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH DI water Items Cleaned (List): probes	
2[]	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	WATER DEPTH 17.6 Measured with:	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:	
4[]	WELL EVACUATION: Method: delicated granges Volume Removed: Observations: Turbidity (clear slightly cloudy Water level (rose fell no change) Water odors:	very cloudy)

5[]	SAMI	LE EXTRA	CTION MET	HOD:				
		[] B [v] P [] C	ailer made o ump, type: Other, describ	of: <u>cluel:ca</u> pe:	led grun	<u>1505</u>		
		Samp	le obtained is	[X] GRA	B; [] COM	POSITE SAN	MPLE .	
6[]	ON-S	TE MEASU	REMENTS:					
~ []	inclumel gal)	6.5	18.0	30.8	35,0			
	Time	1014	1032	1050	1056			Measured with
	Temp (°C)	17.7	17.7	17.8	17.8			Orion 840
	pН	7.31	9.25	7.34	7.24			Orion 250A
	Cond (µS/cm)	1088	1083	1075	1075			Orien 122
-	DO (mg/L)	3.0	3.6	4.1	4.1			Orion 840
	Redox (mV)	-32.1	-5.1	21.2	30.2			Or:01 250A)
	Salinity							
7 [] 8 []		TE SAMPLE	E TREATME	ENT: Method Method	r, size):	Y VOA Containe Containe	ers:	
	ľ	Preser	1			Containe	ers: / /a	25.ml, 250.um
9[]	CONT	AINER HAN	IDLING:					
- L J	, 33		Container S Container L	ides Labeled ids Taped Placed in Ice				
10 [] OTHE	ER COMMEN	JTS:					
٠								
	4-1-1-1-1-1-1							

		ling Location: Tinker AFB - FTA2 ling Dates: \$\frac{1}{3}\frac{1}{3}\frac{1}{3}
CDOLINIDA	ATER SAMPLING RECORD - MONITORING WEL	
REASON FOR DATE AND SAMPLE CONTROL WEATHER	OR SAMPLING: [1] Regular Sampling; [] Special TIME OF SAMPLING: [2] A GRAP OLLECTED BY: [CH OF CAPE O	(number) Al Sampling; As a.m./p.m. As a.m./p.m.
MONITORI	ING WELL CONDITION: [] LOCKED: WELL NUMBER (S) IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (S) [] DEFICIENCIES CORRECTED BY SAMPLE [] MONITORING WELL REQUIRED REPAIR (SIS NOT) APPARENT COLLECTOR
Check-off	EQUIPMENT CLEANED BEFORE USE WITH	DI
2[]	PRODUCT DEPTH Measured with:	FT. BELOW DAT
	WATER DEPTH 7.7' Ltc Measured with:	FT. BELOW DATE
3[]	WATER-CONDITION BEFORE WELL EVACUA Color:	ATION (Describe):
4[]	WELL EVACUATION: Method: Volume Removed: Observations: Turbidity (clear Water level (rose Water odors: Other comments:	Slightly cloudy very cloudy) fell no change)

Groundwater Sampling Record
Monitoring Well No. _ みっよりょ及 (Cont'd)

]. Of		Bailer made of Pump, type:_ Other, describole obtained in	of: lext: c oe:	ated go	W. F. F.	
ulumelo	Samp		<i>desd:</i> <u>(</u>)e:	atril gi	CAYUS.	
ulumelo	Samp					
ulumelo		ole obtained i	TTO ODA	D. [] CO) (
ulumelo	N-SITE MEASU		s [X] GRA	B; [] COME	OSITE SAMPLE	
	_		240			
	0545	-	0909	0915		Measured with
Temp (°C)		18.3	18.3	18:3		01.6540
pН	7.44	9.14	7.10	7.00		Orion 2504
Cond (µS/c		769	764	761		Octon 122
DO (mg/L)		0.3	02	Ø. 2		Octon 840
Redox (mV) -/3d.3	-141.2	-113.4	-105.1		Oran 250A
Salinity	NR	NR	NR	NR		
[]	Filtra ,		Method		_ Containers:	
[2	Y Prese	rvatives adde	ed:			
			Method 5000 Method 5000 Method Method		Containers: 4 Containers: 4 Containers: Containers:	501-m, 14125- UDA
] CO	ONTAINER HA	NDLING:				
	[M] []	Container I	Sides Labeled Lids Taped Placed in Ice			
[] 0	гнев сомме	NTS.				
	THER COMME.					

	Sampling Location: Tinker AFB - FTA2 Sampling Dates: 😕 397
GROUNDWA	ATER SAMPLING RECORD - MONITORING WELL 2 - 2713
REASON FO DATE AND ' SAMPLE CO WEATHER:	(number) R SAMPLING: [1] Regular Sampling; [1] Special Sampling; TIME OF SAMPLING: 2(3)97 at 1315 a.m./pcm. LLECTED BY: 75m4 (cit of Possons. Party Cloudy, 1401 85-95°F R WATER DEPTH MEASUREMENT (Describe): Top Deal control Prop Well Cop
MONITORIN	G WELL CONDITION: [] LOCKED: WELL NUMBER (S)- IS NOT) APPARENT STEEL CASING CONDITION IS:
	INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS - IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):
Check-off	EQUIPMENT CLEANED BEFORE USE WITH DI with
2[4	PRODUCT DEPTH
	WATER DEPTH 18.4' Top well cap TD=46' FT. BELOW DATUM Measured with: Solins + Pure W. = 55 gal.
3[4	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Noe Turbidity: Clear-very Slightly Clear Odor: Noe Other Comments:
4[4	WELL EVACUATION: Method: Declected concles Pup Volume Removed: 50+ gallow Observations: Turbidity (lear slightly cloudy very cloudy) Water level (rose fell na change) Water odors:

					ndwater S Well No 2			
5 [J	Monitoring Well No. <u>と-と7いる</u> (Cont'd) SAMPLE EXTRACTION METHOD:							
	•		[] 0	ther, describ	e:		MPOSITE SAM	
6[]	ı .	ON-SIT	TE MEASUF	REMENTS:				
υ <u>.</u> .	Gallen		10	१७	30	40	50	
	Time		1217	1230	1242	1258	1310	Measured with
	Temp (°	C)	16.7	16.7	16.9	17.0	16.9	42I 22
	pН		7.15	7.14	7.13	7.19	7.19	Awshord
	Cond (µ		948	1511	1195	1235	1238	Orian 122
	DO (mg	/L)	0.79	1.50	255	3.60	3.75	45I 55
	Redox (1	nV)	29.4	NR	60.7	63.8	61.8	ACZS NOW AC
	Salinity		NR	NR	we	Ne	NR	NR
8[4			TE SAMPLE					Plus tie
[]		[]	Filtration:		Method Method Method		Containers	S: S:
		[Preser	vatives added	đ:			
				N	Method /4. Method Sed. Method Method	u Phopphul	Containers	S: Symples Fixed Traganics S: ULM2 S:
9 [/	CONTA	AINER HAN	IDLING:				
			Ĺĺ	Container Li	ides Labeled ids Taped Placed in Ice	Chest		
10 Г	4	OTHER	COMMEN	TS:				

	Sampling Location: Tinker AFB - FTA2 Sampling Dates: £(3(97)
GROUNDV	VATER SAMPLING RECORD - MONITORING WELL 2-274 A
DATE ANI SAMPLE C WEATHER	OR SAMPLING: [X] Regular Sampling; [] Special Sampling; OTIME OF SAMPLING: 213197 at 1000 1000 1000 1000 1000 1000 1000 1
MONITOR	ING WELL CONDITION: [] LOCKED: WELL NUMBER (52- IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS - IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):
Check-off	EQUIPMENT CLEANED BEFORE USE WITH DI water / Alcrey Items Cleaned (List): Crancitos Puna / Protes -> Usoci Dedicated Itane Tusing
2 [-]-	PRODUCT DEPTH
3H	Measured with: To = 84' Tay Puc Puge Cost. = 45 gal. WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Stightly Recells because Turbidity: Stightly Clearly Odor: Day Other Comments:
4[]	WELL EVACUATION: Method:

Groundwater Sampling Record

			Monitoring	Well No.	2-274A	(Cont'd)			
5 [~	SAMP	LE EXTRAC	CTION MET	HOD:					
		[q P [] C Samp	other, describ	pe:		POSITE SAMPLE			
6 [~	ON-SI			30	45				
	Time	0907	0925	0940	1000		Measured with		
	Temp (°C)	20.7	20.60	20.2	20.1		42I22		
	pН	7.78	7.79	7.61	7.60		Aces wind		
	Cond (µS/cm)	527	833	516	516		CYULA 122		
	DO (mg/L)	7.5	7.35	€.80*	8.1		422 ES		
	Redox (mV)	62.8	60.7	625	63.1		AUSSAINO		
	Salinity	NR.	m	+ ane wal	we		NR		
8 [<u>-</u>	ON-SI	TE SAMPLE	ion: 1			Containers: Containers:	الجائز		
₽ Pr			Preservatives added: Method H. Goy Containers: Fixed Inagamus / Symun J Method Seekum Phosphake Containers: Uksis Method Containers:						
9 [-	CONT	ίí	NDLING: Container S Container L						
Containers Placed in Ice Chest 10 [] OTHER COMMENTS:									

Sampling Location: <u>Tinker AFB - FTA2</u>

	Sampling Dates: 812197	
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL てってって	
DATE AND SAMPLE CO	OR SAMPLING: [4] Regular Sampling; [] Special Sampling; TIME OF SAMPLING: 8 3197 at 0850 cm/p.m. OLLECTED BY: Sampling: of Passas. Parky closely warm, 2006	(number)
DATUM FO	R WATER DEPTH MEASUREMENT (Describe): Top Declarated Pure	p Well cap
MONITORIN	NG WELL CONDITION: [] LOCKED: WELL NUMBER (S- IS NOT) APPARENT STEEL CASING CONDITION IS:	
	WATER DEPTH MEASUREMENT DATUM (S)- IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off 1 [-	EQUIPMENT CLEANED BEFORE USE WITH DI water Items Cleaned (List): Probes	
2[4	PRODUCT DEPTH	FT. BELOW DATUM
	WATER DEPTH 16.72' Measured with: Solvast Page Ust = 50 go	FT. BELOW DATUM
3 [4	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Clear to Shiphty Cleary Turbidity: "" Odor: New Other Comments:	
4 [_]—	WELL EVACUATION: Method: Pechcated Cruncifes Purp Volume Removed: Stage Observations: Turbidity (clean slightly cloudy Water level (rose fell - no change) Water odors: Other comments:	very cloudy)

Groundwater Sampling Record

Monitoring Well No. 2-2743 (Cont'd)

SAMPLE EXTRACTION METHOD:

5 [4

						cifes	
		Samp	le obtained i	s [X] GRAI	B; [] CON	MPOSITE SAM	MPLE
6[]	ON-S	ITE MEASU	REMENTS:				
	Callans	10	76	30	40	50	
	Time	0805	08.5	0 225	0835	0850	Measured with
	Temp (°C)	17.5	17.5	17.5	17.6	ר.רו	45I55
	pН	we	6.82	7.09	7.17	17.20	Ass war
- [Cond (µS/cm)	906	915	918	920	919	Orica 122
ı	DO (mg/L)	5115	5.0	4.95	5.16	5.14	4555
- 1	Redox (mV)	Mr	106	79.8	49.0	20.0	1005 NOW
- 1	Salinity	NR	,vr	we.	W.	we	NR
8 [+]	ON-S	Filtrat	tion: vatives adde	Method Method Methoded:		Containe Containe	ers:ers:
				Method Sand	em Phospher	Containe	ers: Synm Jos / Foxed I
				Method		Containe	ers:
				Method		Containe	ers:
9[-]	CON	TAINER HAI	NDLING:				
		[] [] []	Container I	Sides Labeled Lids Taped Placed in Ice			
10 [] OTH	ER COMMEN	NTS:				

Sampling Location: Tinker AFB - FTA2

Sampling Dates: £13197 GROUNDWATER SAMPLING RECORD - MONITORING WELL Z-3553 (number) REASON FOR SAMPLING: [A Regular Sampling; [] Special Sampling; DATE AND TIME OF SAMPLING: 8 13197 at 1600 a.m./pm SAMPLE COLLECTED BY: 5 mit cit of Passars.

WEATHER: Party Classer | 1 tot 90-95 F

DATUM FOR WATER DEPTH MEASUREMENT (Describe): Top PUC - well head compression not completed MONITORING WELL CONDITION: LUNLOCKED [] LOCKED: WELL NUMBER (IS - IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS - IX NOT) APPARENT DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR MONITORING WELL REQUIRED REPAIR (describe): Check-off EQUIPMENT CLEANED BEFORE USE WITH DI wester Access 1 [4 Items Cleaned (List): Pup / wire PRODUCT DEPTH FT. BELOW DATUM 214-Measured with: WATER DEPTH 15.73' Top Pic Rise TD > 39.24' TPIC FT. BELOW DATUM

Measured with: Solinst Project. = 4c gal. WATER-CONDITION BEFORE WELL EVACUATION (Describe): 3[-Color: Brown Turbidity: Clarky Other Comments: WELL EVACUATION: 4[4 Method: Connales Pupul dechecuted HDPE Tubing Volume Removed: 50 591. Turbidity (dear slightly cloudy very cloudy) Observations: Water level (rose en no change) Other comments:

			Monitoring	Well No	Sampling 1 Z-35535	Record (Cont'd)	·
5 🗠	SAMF						
		[M] P	ailer made o ump, type: other, describ	and	, fos		
		Samp	e obtained is	[X] GRA	B; [] CO	MPOSITE SAN	MPLE
6[] ON-SI	TE MEASU			43.00	~~	
	Time	10		30	1545	1355	Measured with
	Temp (°C)	1506	1525	1535			YSI 55
	pH	17.8	178	7.23	17.8	17.8	ConcresoA
	Cond (µS/cm)	7.21	7.22		1485	7, 23	Oranizz
	DO (mg/L)	1450	1460	1480		1490	45I S5
	Redox (mV)	9.05	8.90	7.86	7.90	7.85	
	Salinity	NR	122.8	124.9 NR	NR	116.7 NR	Orien250A NR
, L	SAME		`			2 Symmy	rs Plasta
						2 230 MI	* raser
8 [ON-SI	TE SAMPLE	ETREATME	ENT:			
	[]	Filtrat	ion:	Method		Containe	ers:
	()	11111		Method		Containe	ers:
			- 1	Method		Containe	ers:
	N	Preser	vatives adde	d:			
]]	<u>حيح</u> Method	LSO4	Containe Containe	ers: Byrun Tax (Franch Triangene ers: USAz
o t	⊬ CONπ	AINER HAI		Method		Containe	ers:
9 🗗	T CONI	AIINEK FIAI		ides Labeled	ı		
		[]	Container L				

10 OTHER COMMENTS:_

		ampling Location: Tinker Alampling Dates: 5/3/47	FB - FTA2
GROUNDW	ATER SAMPLING RECORD - MONITORING	WELL 2-3723	*
DATE AND SAMPLE CO	OR SAMPLING: [X] Regular Sampling; [] SI TIME OF SAMPLING: 813 197 at OLLECTED BY: 5mlt 1011 of Poso Posty Cloudy, 11st, 90-95°F	1815 a.m./pm	(number)
DATUM FO	R WATER DEPTH MEASUREMENT (Describe): Top Puc Casing	· Noton North Sich
MONITORI	NG WELL CONDITION: [] LOCKED: WELL NUMBER (S- IS NOT) APPARENT STEEL CASING CONDITION IS:		
	INNER PVC CASING CONDITION IS:	(S) IS NOT) APPARENT PLE COLLECTOR	
Check-off 1 [EQUIPMENT CLEANED BEFORE USE WIT Items Cleaned (List): Purp		
2[4	PRODUCT DEPTH Measured with:		FT. BELOW DATUM
	WATER DEPTH Water = 15.43' Measured with: Solvent	70 = 460 - 3'Sup Pupe Ud. = 60 gal.	FT. BELOW DATUM
3 []	WATER-CONDITION BEFORE WELL EVAN Color: Nave Turbidity: Sighty Closely Odor: Nave Other Comments:		
4[]	Water level (ros	slightly cloudy se Gell no change)	very cloudy)

Groundwater Sampling Record

Monitoring Well No. 2-392B (Cont'd)

5 [-]		SAMPI	E EXTRA	CTION MET	THOD:			
			[] []	Bailer made o Pump, type:_ Other, descri	of:be:	S		
			Samp	le obtained i	s [X] GRAI	B; [] COM	IPOSITE SAI	MPLE
6 [-]	_	ON-SIT	TE MEASU	REMENTS:				
			10	20	3∅	46	so	
ļ	Time		1710	1725	1750	1800	1816	Measured with
	Temp (°C)	18.3	18.8	(8.6	18.6	18.5	५प्र ६८
ı	pН		7.13	7.17	7.15	7.17	7.18	A285 MYC)
	Cond (μS/cm)	693	698	696	695	697	Onien 122
	DO (m	g/L)	8.05	7.68	4.68	6.05	6.00	YSISS
	Redox	(mV)	سعد	NR.	NR	135.2	138.7	AGSwin
	Salinit	у	NR	ML	NR	M	pe	Ne
8 [J		[]	Filtra		Method Method Method		Contain	ers: ers:
		H	Prese	rvatives add			Contain	ers: <u>Synun Tas Fixed</u> Inagaun ers: <u>UON</u> s ers:
9 [-]	<u> </u>	CONT	AINER HA	NDLING:				
			17 [] []	Container	Sides Labeled Lids Taped Placed in Ice			
10 [١.	OTHE	R COMME	NTS:				

	Sampling Location: Tinker AFB - FTA2 Sampling Dates: 81447
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL Z-393 3
DATE AND SAMPLE CO	(number) R SAMPLING: [/] Regular Sampling; [] Special Sampling; TIME OF SAMPLING: 814147 at 1500 a.m./pcm. LLECTED BY: Tomy DK of Process EPA Clear, Sumy Hot
DATUM FO	R WATER DEPTH MEASUREMENT (Describe): Top DC Casing - Notch Worth Side
MONITORIN	IG WELL CONDITION: [] LOCKED: WELL NUMBER (IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS:
	WATER DEPTH MEASUREMENT DATUM (IS - IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):
Check-off 1	EQUIPMENT CLEANED BEFORE USE WITH DI wester / Averox Items Cleaned (List): Probes / Purps + Tobing
2 H	PRODUCT DEPTHFT. BELOW DATUM Measured with:
	WATER DEPTH 15.9' Top PVC FT. BELOW DATUM Measured with: Solvet
314	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: ^ C^ C Turbidity: C ker Odor: reve Other Comments:
4 H	WELL EVACUATION: Method: Crucifes Pup Volume Removed: Developed 350 galkus + 15 galkus to marker stability Observations: Turbidity Clear slightly cloudy very cloudy) Water level (rose fell northange) Water odors: New Other comments:

					Sampling Rec 2-393 B			
5 [Y SAMP	LE EXTRA	CTION ME	THOD:				
		[](Other, descri	be:	AB; [] COMPO	OSITE SAMPLE		
6 [~	ON-SI		REMENTS:					
	Diveloped 3	1445	1480	1422			Measured with	
	Temp (°C)	18.3	18.4				YSI SS	
	рН	7-33	7.21	7.23			Churcey	
	Cond (µS/cm)	1195	1186	1134			Orien 12Z	
	DO (mg/L)	7.73	7.25	7,00			48I 22	
	Redox (mV)	NR	105.3	110.9			Orienzian.	
	Salinity	N	ne	NS	·			
8 K	SAMP ON-SI	TE SAMPLI				2 synun jas 2 250mi. Pl	astic	
	[]	[] Filtration:		Method		Containers: Containers:		
	14	Prese	eservatives added:					
				Method_Sec Method_	to SO.1	Containers: Containers:	water Fixed Inorgania	
9 [-	CONT	AINER HA	NDLING:					
		[4] [] []	Container l	Sides Labele Lids Taped Placed in Ic				

OTHER COMMENTS:___

10[]

	Sampling Location: <u>Tinker AFB</u> Sampling Dates: <u>をリリクリ</u>	<u>- FTA2</u>
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL <u>C639 - 139</u>	7-41
REASON FO	OR SAMPLING: [Regular Sampling; [] Special Sampling; TIME OF SAMPLING: 8/4 at 9:50 a.m./p.m. OLLECTED BY: CH/Ok of Parsens/ LASEPA	(number)
	R WATER DEPTH MEASUREMENT (Describe): Tex	
MONITORII	NG WELL CONDITION: [] LOCKED: WELL NUMBER (IS - IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER DEPTH MEASUREMENT DATUM (IS SIS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR	
	[] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off	EQUIPMENT CLEANED BEFORE USE WITH DI water Items Cleaned (List): probes	
2[4]	PRODUCT DEPTH	FT. BELOW DATUM
	WATER DEPTH 12.33' Measured with:akrakr	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:	
4[]	WELL EVACUATION: Method:	very cloudy)

Groundwater Sampling Record

Monitoring Well No. <u><<39-897-41</u> (Cont'd)

5[]] SAMPLE EXTRACTION METHOD:									
[] Bailer made of:										
	Sample obtained is [X] GRAB; [] COMPOSITE SAMPLE									
6[]	1 .	ON-SI	TE MEASUI	REMENTS:	calibrate	d pHz a	cles			
	Time	rigal)	9.0	10:38	4 12.0	11.0			Measured with	7
		90)	10:14	-						-
	Temp (°C)		21.1	21.7						┪
	Cond (µS/cm)		7.11	6.92	6.99	6.88				-
			862	863	861)	861				-
	DO (mg		3.7	3.6	(3.9)	3.9				-
	Redox		153.8	109,1	156.0					-
	Salinity	7	NR_	NR	NR	NR				
7[]]	SAMP	LE CONTAI	NERS (mat	erial, number	, size):	Voit 125-1 Servi	w r		
8[]	ON-SITE SAMPLE TREATMENT:									
		[]	Filtrat		Method Method Method		Contain	ers:		
		[Y	Preser	reservatives added:						
					Method Sum Method Method Method	lfuric act	Contain Contain Contain Contain	ers:	Serum 112 VOAz	<u>\$</u> -,L
9[CONTAINER HANDLING:									
			[] []	Container 1	Sides Labeled Lids Taped Placed in Ice	Chest				
10[]		OTHER COMMENTS: Conductivity standard checked out exactly								
				,	cec): 61	wheel to	mithin	0.02		
				Ded	the coulib	ruped to	mithin.	10 mV		

GROUNDWATER SAMPLING RECORD

	Sampling Dates:(639-857-435 6/4/97
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL	514/97 A97-435
		(number)
DATE AND SAMPLE CO	OR SAMPLING: [1] Regular Sampling; [1] Special Sampling; TIME OF SAMPLING: 3/4 at 1/50 a.m./p.m. OLLECTED BY: CH/DU of Parsons/USEPA Clear Lat (~75°F) OR WATER DEPTH MEASUREMENT (Describe): TOC	(
MONTTONT	NO WELL CONDITION.	
MONITORI	NG WELL CONDITION: [] LOCKED: WELL NUMBER (IS - IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS:)
	WATER DEPTH MEASUREMENT DATUM (\$\subseteq IS NOT) APPARE [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	ENT
Check-off	EQUIPMENT CLEANED BEFORE USE WITH DI witer Items Cleaned (List): probes prop	
2[]	PRODUCT DEPTH Measured with:	FT. BELOW DATUM
	WATER DEPTH 30.64 Measured with: water noter	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: (
4[]	WELL EVACUATION: Method: Volume Removed: Turbidity (clear) Water level (rose fell no change) Water odors: Other comments:	

Groundwater Sampling Record
Monitoring Well No. 639-697-435 (Cont'd)

5[]	SAMP	LE EXTRAC	CTION MET	HOD:				
		[] Ba [[] Pu [] O	ailer made o ump, type: ther, describ	of: grunsos oe:				
		Sample	e obtained is	s [X] GRA	B; [] COM	POSITE SAMP	LE	
6[]		TE MEASUR		5.0	9.0			
1		1,5			T		Measured v	with
	Time		11:55		12.21		Wicasured v	VILII
	Temp (°C)	19.5	20.2	22.8	21.5			
	pН		7.16	<i>5.83</i>	6.85			
	Cond (µS/cm)	1265	1262	1269	12.78			
	DO (mg/L)	5.1	4.8	4.8	4.8			
	Redox (mV)	122.7	117.3	109.1	116.7			
	Salinity	NR	NR	NR				
8[]	on-si	TE SAMPLE	ion:]			Containers:		
	M	Preserv	vatives adde					
]	Method_ <u><ue< u=""> Method_<u>suf</u> Method Method</ue<></u>		Containers: Containers: Containers: Containers:	4 VOAS	2 Se1"
9[]	CONT	AINER HAN	IDLING:					
		ii	Container L	ides Labeled ids Taped Placed in Ice				
10 [] OTHE	R COMMEN	ITS:					

GROUNDWATER SAMPLING RECORD

	Sampling Location: Tinker AFB - FTA2 Sampling Dates: 9/4/57
GROUNDW	ATER SAMPLING RECORD - MONITORING WELL
DATE AND SAMPLE CO	(number) OR SAMPLING: [4] Regular Sampling; [] Special Sampling; TIME OF SAMPLING: 8/4197 at 0900 pp./p.m. OLLECTED BY: 3m4 (C1 of Pascus Porty Clady, Hot, 85-90 F R WATER DEPTH MEASUREMENT (Describe): Top pue Reserved to be to Side
	· · · · · · · · · · · · · · · · · · ·
MONITORII	WELL CONDITION: [] LOCKED: WELL NUMBER(IS) IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS:
	WATER DEPTH MEASUREMENT DATUM (IS - IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):
Check-off 1	EQUIPMENT CLEANED BEFORE USE WITH Do was to leaned (List): Pup & Probect
2[4—	PRODUCT DEPTHFT. BELOW DATUM Measured with:
	WATER DEPTH Water = 20.19 To = 49.93 (messed) Measured with: Solvest Que Source = 15gal.
3 H	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color: Stightly Reciden Brown Turbidity: Stightly Classy Odor: Occ. Other Comments:
4[4	WELL EVACUATION: Method: (Michigan Pun) Volume Removed: 16 + gallers Observations: Turbidity (clear slightly cloudy very cloudy) Water level (rose fell nochange) Water odors: Nac. Other comments:

Groundwater Sampling Record

Monitoring Well No. C659-697-43d (Cont'd)

5 [-	<u> </u>	SAMPI	E EXTRAC	TION MET	HOD:				
			[of Pi	imp, type:	E BMIL Borthood	ed Crus	the Pup	- Docheated Tobing	
			Sample	e obtained is	[X] GRAE	3; [] CON	POSITE SAN	MPLE	
ان] 6	r	ON-SIT	TE MEASUR	EMENTS:					
			3.℃	8.0	12.0	160	18.0	Measured with	
	Time		0180	०१८६	0335	2842	0880		
	Temp (°	C)	19.6	18.7	18.6	18.50	18.40	75.T 35	
	pН		7.75	7.14	7.09	7.08	7.07	Onas 250A	
	Cond (µ	.S/cm)	1115	1110	1110	1105	1112	CAMEN 12Z	
	DO (mg	/L)	3.29	333	3.36	3.35	3.34	TSI SS	
	Redox (mV)	NR	m	134	134.6	129.8	Chien 25ch	
	Salinity		me.	ne	we.	we	w	NR	
8H-		ON-SITE SAMPLE TREATMENT: [] Filtration: Method Containers: Method Containers: Containers:							
		Method Section Phosphak Containers: UCAS Method Hasoy Containers: Symmisus / Kinece Method Containers: Method Containers: Method Containers:							
9[CONT	AINER HAN	IDLING:					
			ίi	Container L	ides Labeled ids Taped Placed in Ice	Chest			
10 []	OTHE	R COMMEN	ITS:					

GROUNDWATER SAMPLING RECORD

Sampling Location: <u>Tinker AFB - FTA2</u>

	Sampling Dates: 5/4/	9.1
GROUNDWA	ATER SAMPLING RECORD - MONITORING WELL C 6 3 9 - 13 9	7-45 (number)
REASON FO DATE AND SAMPLE CO WEATHER: DATUM FOR	R SAMPLING: [Regular Sampling; [] Special Sampling; TIME OF SAMPLING:	(number)
MONITORIN	IG WELL CONDITION: [] LOCKED: WELL NUMBER (IS - IS NOT) APPARENT STEEL CASING CONDITION IS: INNER PVC CASING CONDITION IS: WATER PERTIL NEAS SUPERIENT DATES OF SOME APPARENT.	
	WATER DEPTH MEASUREMENT DATUM (S) IS NOT) APPARENT [] DEFICIENCIES CORRECTED BY SAMPLE COLLECTOR [] MONITORING WELL REQUIRED REPAIR (describe):	
Check-off 1 []	EQUIPMENT CLEANED BEFORE USE WITH DI water Items Cleaned (List): probes, 5,00,505 pa	mp
2[]	PRODUCT DEPTH	
	WATER DEPTH 12.14 Measured with:	FT. BELOW DATUM
3[]	WATER-CONDITION BEFORE WELL EVACUATION (Describe): Color:	
4[]	WELL EVACUATION: Method: Volume Removed: Volume Remove	very cloudy)

Groundwater Sampling Record

Monitoring Well No. 639-1397-45 (Cont'd)

5[]	SAMPI	LE EXTRAC	CTION MET	THOD:		•			
		[] E [;/ P	sailer made o ump, type:_ Other, describ	of:grung	03				
		Samp	le obtained is	s [X] GRAE	3; [] CON	MPOSITE SA	MPLE		
6[]	volumo (gal)	TE MEASU	REMENTS:	11.0	14.0	20.0			
1	Time	1410	1418		1441	1458		Measured with	
Δ	Temp (°C)	19.4	19.2	19.3	19.3	19.4			
	pН	7.02	7.16	64 (.2)	7.14	6.91			
	Cond (µS/cm)	881	882	874	871	872			
	DO (mg/L)	7.3	7.1	6.60	6.2	6 .5			
	Redox (mV)	114.6	105.1	102.8	104.0	103.2			
	Salinity	NR	NR	NR	NR	NR			
7[] 8[]			E TREATME	ENT:		Contain Contain	ers:		
	M	Preser						Serum 125-m	Ĺ
9[]	CONT	AINER HAI	NDLING:						
		[] [] []	Container L	ides Labeled ids Taped Placed in Ice	Chest				
10 [) OTHE	R COMMEN	VTS:						

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/20/97 13:00:30 TEST DESCRIPTION Data set..... CG-41-1.DAT Output file..... CG-41-1.OUT Data set title.... MONITORING WELL CG39-B97-41: RUN 1 Company..... PARSONS ES Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time..... min Test Well Data Initial displacement in well.... 3.7 Radius of well casing..... 0.083 Radius of wellbore......... 0.354 Aquifer saturated thickness..... 27.67 Well screen length..... 5 Static height of water in well... 26.67 Gravel pack porosity..... 0.35 Effective well casing radius.... 0.2199 Effective wellbore radius..... 0.354 Log (Re/Rw) 2.394 Constants A, B and C..... 2.000 , 0.294, 0.000 No. of observations........... 182 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aguifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 6.1279E-003 ft/min

y0 = 6.0960E + 000 ft

CLIENT: TINKER AFB	COMPANY: PARSONS ES	
LOCATION: OKLAHOMA	PROJECT: 729691,34050	
MONITORING WELL (CG39-B97-41:	RUN 1
	AOH	DATA SET: CG-41-1.DAT 11/20/97
10.		AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice
1	H = S S =	TEST DATA: H0= 3.7 ft rc= 0.083 ft rw= 0.354 ft I = 5 ft
(13)	iet	b = 27.67 ft H = 26.67 ft
Displacement		PARAMETER ESTIMATES: K = 0.006128 ft/min y0 = 6.096 ft
1. http://www.linihii.linihiii.linihiii.		
Time (min)		TESOLU

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/20/97 13:06:41 TEST DESCRIPTION Data set..... CG-41-2.DAT Output file..... CG-41-2.OUT Data set title.... MONITORING WELL CG39-B97-41: RUN 2 Company..... PARSONS ES Project...... 729691.34050 Client..... TINKER AFB Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length.... ft Time..... min Test Well Data Initial displacement in well.... 3.2 Radius of well casing...... 0.083 Radius of wellbore..... 0.354 Aquifer saturated thickness..... 27.67 Well screen length..... 5 Static height of water in well... 26.67 Gravel pack porosity..... 0.35 Effective well casing radius.... 0.2199 Effective wellbore radius..... 0.354 Constants A, B and C..... 2.000 , 0.294, No. of observations............ 163 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aguifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 8.2784E-003 ft/min

y0 = 5.5668E + 000 ft

COMPANY: PARSONS ES	PROJECT: 729691,34050	MONITORING WELL CG39-B97-41: RUN 2	DATE SET: 10-20-97 1-2	
CLIENT: TINKER AFB	LOCATION: OXLAHOMA	TINOM	Displacement (ft) 0 10 10 10 10 10 10 10 10 10	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/20/97 13:20:09

TEST DESCRIPTION

Data se	et				CG-43S-1.DAT
Output	file.				CG-43S-1.OUT

Data set title.... MONITORING WELL CG39-B97-43S: RUN 1

Units of Measurement

Length..... ft Time.... min

Test Well Data

ANALYTICAL METHOD

Bouwer-Rice (Unconfined Aquifer Slug Test)

RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 1.2560E-002 ft/min

y0 = 3.6813E + 000 ft

COMPANY: PARSONS ES	PROJECT: 729691.34050	RING WELL CG39-B97-43S: RUN 1	CG-T3S-1.DAT 11.20-77 11.20-77 AQUITER MODEL: Unconfined SULUTION METHOD: Bouwer-Bice TEST DATA: Te= 0.063 ft Te= 0.063 ft Te= 0.063 ft Te= 0.063 ft Te= 0.063 ft The Description of the confinence of the conf	ESOLU
CLIENT: TINKER AFB	LOCATION: OXLAHOMA	MONITORING	Time 4	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

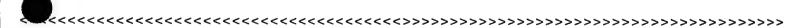
11/20/97 13:28:50 TEST DESCRIPTION Data set..... CG-43S-2.DAT Output file..... CG-43S-2.OUT Data set title.... MONITORING WELL CG39-B97-43S: RUN 2 Company..... PARSONS ES Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time.... min Test Well Data Initial displacement in well..... 3.1 Radius of well casing..... 0.083 Radius of wellbore......... 0.354 Aquifer saturated thickness..... 16.1 Well screen length..... 5 Static height of water in well... 15.1 Gravel pack porosity...... 0.35 Effective well casing radius.... 0.2199 Effective wellbore radius..... 0.354 Log (Re/Rw) 2.192 Constants A, B and C..... 2.000 , 0.294, No. of observations........... 183 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aguifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 1.9713E-002 ft/min

y0 = 2.8198E + 000 ft



PARCING FC	CG39-B97-43S: RUN 2	DMTM SET: CG-43S-2.DAT 11.20-97 AQUIFER HODEL: Uncomf ined SOLUTION METHOD: Bouwer-Rice TEST DATH: HO = 3.1 ft r_ = 0.083 ft r_ = 0.083 ft r_ = 0.083 ft t	
TINKER AFR	 MONITORING WELL	Displacement (ft) 0.0 0.1 0.0 0.0 1.2 1.2 1.3 1.4 1.5 1.5 1.5 1.5 1.5 1.5 1.5	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/17/97 15:29:03 TEST DESCRIPTION Data set..... CG-43S-3.DAT Output file..... CG-43S-3.OUT Data set title..... MONITORING WELL CG39-B97-43S: RUN 3 Company..... PARSONS ES Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time..... min Test Well Data Initial displacement in well.... 3.5 Radius of well casing..... 0.083 Radius of wellbore......... 0.354 Aquifer saturated thickness..... 16.1 Well screen length..... 5 Static height of water in well... 15.1 Gravel pack porosity..... 0.35 Effective well casing radius.... 0.2199 Effective wellbore radius..... 0.354 Log (Re/Rw) 2.192 Constants A, B and C..... 2.000 , 0.294, 0.000 No. of observations...... 181 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aquifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING VISUAL MATCH PARAMETER ESTIMATES

Estimate

v0 =

1.0860E-002 ft/min

3.4957E+000 ft

		13	DATA SET: CG-43S-3.DAT 11/17/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 3.5 ft L = 5. ft L = 5. ft H = 15.1 ft H = 15.1 ft H = 15.1 ft Year 0.354 ft L = 5. ft U = 5.	SOLU
COMPANY: PARSONS ES	PROJECT: 729691,34050	97-43S: RUN	DATA SET: CG-43S-3.DAT 11/17/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 3.5 ft r_= 0.083 ft r_= 0.083 ft r_= 0.083 ft r_= 0.08	
COMPANY:	PROJECT:	WELL CG39-B97-43S:	4. 5.	•
		MONITORING W	1. 2. 3. Time (min)	
CLIENT: TINKER AFB	LOCATION: OXIAHOMA		Displacement (ft) Displacement (ft)	•

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/17/97 14:57:40

TEST DESCRIPTION

Data set..... CG-43D-1.DAT Output file..... CG-43D-1.OUT

Data set title.... MONITORING WELL CG39-B97-43D: RUN 1

Company..... PARSONS ES Project...... 729691.34050 Client..... TINKER AFB Location..... OKLAHOMA Test date..... AUGUST 1997

Units of Measurement Length..... ft Time..... min

Test Well Data

Initial displacement in well.... 3.5 Radius of well casing..... 0.083 Radius of wellbore..... 0.354 quifer saturated thickness..... 29.81 Well screen length..... 5 Static height of water in well... 29.31 Gravel pack porosity..... 0.35 Effective well casing radius.... 0.2199 Effective wellbore radius..... 0.354

Constants A, B and C..... 2.000 , 0.294,

No. of observations..... 182

ANALYTICAL METHOD

Bouwer-Rice (Unconfined Aquifer Slug Test)

RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

1.8731E-002 ft/min

v0 =3.5280E+000 ft



2	020	D: RUN 1	DATA SET: CG-43D-1.DAT 11/17/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 3.5 ft r_0= 0.083 ft r_0= 0.083 ft L = 5. ft L = 5. ft L = 5. ft K = 0.354 ft K = 0.358 ft FARAMETER ESTIMATES: K = 0.01873 ft/min y0 = 3.528 ft	NIOS
COMPANY: PARSONS ES	PROJECT: 729691,34050	WELL CG39-B97-43D:		5. 6. 7. 8. (min)
		MONITORING		1. 2. 3. 4 Time
CLIENT: TINKER AFB	LOCATION: OKLAHOMA		Displacement (ft)	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/21/97 09:42:38 TEST DESCRIPTION Data set..... CG-43D-2.DAT Output file..... CG-43D-2.OUT Data set title.... MONITORING WELL CG39-B97-43D: RUN 2 Company..... PARSONS ES Project..... 729691.34050 Client..... TINKER AFB Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time..... min Test Well Data Initial displacement in well.... 3.2 Radius of well casing..... 0.083 Radius of wellbore......... 0.354 Aguifer saturated thickness..... 29.81 Well screen length..... 5 Static height of water in well... 29.31 Gravel pack porosity...... 0.35 Effective well casing radius.... 0.2199 Effective wellbore radius..... 0.354 Constants A, B and C..... 2.000, 0.294, No. of observations........... 185 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aquifer Slug Test) _______ RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 2.5238E-002 ft/min

y0 = 3.2824E + 000 ft

COMPANY: PARSONS ES	PROJECT: 729691,34050	3 WELL CG39-B97-43D: RUN 2	The control of the	ESOLU
CLIENT: TINKER AFB	LOCATION: OKLAHOMA	MONITORING WI	·	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/21/97 09:48:24 TEST DESCRIPTION Data set..... CG-45-1.DAT Output file..... CG-45-1.OUT Data set title.... MONITORING WELL CG39-B97-45: RUN 1 Company..... PARSONS ES Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time.... min Test Well Data Initial displacement in well.... 5.5 Radius of well casing...... 0.083 ladius of wellbore.......... 0.354 Aquifer saturated thickness..... 29.89 Well screen length..... 5 Static height of water in well... 29.39 Gravel pack porosity..... 0.35 Effective well casing radius.... 0.2199 Effective wellbore radius..... 0.354 Constants A, B and C..... 2.000 , 0.294, 0.000 No. of observations........... 185 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aquifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

3.0218E-002 ft/min

v0 =2.9465E+000 ft



	.34050	45: RUN 1	DATA SET: CG-45-1.DAT 11/21/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 5.5 ft r_m = 0.354 ft L = 5. ft b = 29.89 ft H = 29.39 ft H = 29.39 ft FARAMETER ESTIMATES: K = 0.03022 ft/min y0 = 2.947 ft	NTOSK .
COMPANY: PARSONS ES	PROJECT: 729691,34050	MONITORING WELL CG39-B97-45:		
CLIENT: TINKER AFB	LOCATION: OKLAHOMA		Displacement (ft) 19 19 19 19	

AOTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/21/97 10:04:43 TEST DESCRIPTION Data set..... 2-355B-1.DAT Output file..... 2-355B-1.OUT Data set title.... MONITORING WELL 2-355B: RUN 1 Company..... PARSONS ES Project..... 729691.34050 Client..... TINKER AFB Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time.... min Test Well Data Initial displacement in well.... 1.71 Radius of well casing..... 0.167 adius of wellbore.......... 0.354 Aquifer saturated thickness..... 15.27 Well screen length..... 10 Static height of water in well... 14.17 Gravel pack porosity..... 0.35 Effective well casing radius.... 0.249 Effective wellbore radius..... 0.354 Log (Re/Rw) 2.511 Constants A, B and C..... 2.390 , 0.388, 0.000 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aguifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING VISUAL MATCH PARAMETER ESTIMATES

Estimate

5.3140E-003 ft/min

y0 =2.4405E+000 ft



			TIMATES:	NTOS.
COMPANY: PARSONS ES	PROJECT: 729691,34050	ING WELL 2-355B: RUN 1	DATA SET: 2-355B-1.DAT 11/21/97 11/21/97 AQUIFER MODEL: Uncomfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 1.71 ft r_0= 0.167 ft r_0= 0.167 ft r_0= 0.167 ft the 1.71 ft t	
CLIENT: TINKER AFB	LOCATION: OXLAHOMA	MONITORIP	Displacement (ft) 0.1 0.1 0.1 1.2 1.2 1.6 1.6 1.7 1.6 1.7 1.6 1.7 1.7	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

TEST DESCRIPTION

Data set 2-355B-2.DAT
Output file..... 2-355B-2.OUT
Data set title.... MONITORING WELL 2-355B: RUN 2

Company...... PARSONS ES
Project....... 729691.34050
Client...... TINKER AFB
Location..... OKLAHOMA
Test date..... AUGUST 1997

Units of Measurement
Length..... ft
Time.... min

11/17/97

ANALYTICAL METHOD

Bouwer-Rice (Unconfined Aquifer Slug Test)

RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 4.5351E-003 ft/min

y0 = 2.7593E + 000 ft

	20	RUN 2	DATA SET: 2-355B-2.DAT 11/17/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 1.854 ft r_d= 0.167 ft I = 10. ft L = 10. ft L = 10. ft E = 15.27 ft H = 14.17 ft PARAMETER ESTIMATES: K = 0.004535 ft/min y0 = 2.759 ft	SOLU
COMPANY: PARSONS ES	PROJECT: 729691,34050	MONITORING WELL 2-355B: F	7	
CLIENT: TINKER AFB	LOCATION: OKLAHOMA		Displacement (ft)	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/21/97 10:26:56 TEST DESCRIPTION Data set..... 2-392B-1.DAT Output file..... 2-392B-1.OUT Data set title.... MONITORING WELL 2-392B: RUN 1 Company..... PARSONS ES Project..... 729691.34050 Client..... TINKER AFB Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time.... min Test Well Data Initial displacement in well..... 3.1 Radius of well casing..... 0.167 Radius of wellbore......... 0.3554 Aquifer saturated thickness..... 30.57 Well screen length..... 10 Static height of water in well... 30.57 Gravel pack porosity...... 0.35 Effective well casing radius.... 0.2497 Effective wellbore radius..... 0.3554 Log (Re/Rw) 3.176 Constants A, B and C..... 0.000 , 0.000, 1.910 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aquifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 8.5770E-004 ft/min

v0 = 3.9670E-001 ft



ES	34050	RUN 1	DATA SET: 2-392B-1.DAT 11/21/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 3.1 ft r_a= 0.167 ft L= 10. ft b = 30.57 ft H = 30.57 ft H = 30.57 ft SOLUTION METHOD: Bouwer-Rice SOLUTION METHOD: B	SOLU
COMPANY: PARSONS ES	PROJECT: 729691,34050	MONITORING WELL 2-392B:	3. 4. 5. 6. 7. 8. 9. 10. Time (min)	
CLIENT: TINKER AFB	LOCATION: OKLAHOMA		Displacement (ft) ei 1. 9. 9. 9. 1. 1. 1. 1. 1. 2. 1. 2. 1. 2. 1. 2. 3. 4. 5. 6. 6. 7. 7. 8. 9. 9. 1. 1. 1. 1. 1. 1. 1. 1	

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/17/97 13:46:58 TEST DESCRIPTION Data set..... 2-392B-2.DAT Output file..... 2-392B-2.OUT Data set title.... MONITORING WELL 2-392B: RUN 2 Company..... PARSONS ES Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time.... min Test Well Data Initial displacement in well.... 3.1 Radius of well casing..... 0.167 Radius of wellbore......... 0.3554 Aquifer saturated thickness..... 30.57 Well screen length..... 10 Static height of water in well... 30.57 Gravel pack porosity...... 0.35 Effective well casing radius.... 0.2497 Effective wellbore radius..... 0.3554 Constants A, B and C..... 0.000, 0.000, No. of observations...... 182 ANALYTICAL METHOD Bouwer-Rice (Unconfined Aguifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 6.2383E-004 ft/min

y0 = 6.1331E-001 ft

	20	RUN 2	DATA SET: 2-392B-2.DAT 11/17/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice	H0= 3.1 ft r _c = 0.167 ft r _w = 0.3554 ft L = 10. ft b = 30.57 ft H = 30.57 ft	PARAMETER ESTIMATES: K = 0.0006238 ft/min y0 = 0.6133 ft	\(\tau_{1083}\)
COMPANY: PARSONS ES	PROJECT: 729691,34050	MONITORING WELL 2-392B: 1				4. 6. 8. 10.
CLIENT: TINKER AFB	LOCATION: OKLAHOMA		19.	(13)	Displacement	0.1

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

11/21/97 10:43:43 TEST DESCRIPTION Data set..... 2-393B-1.DAT Output file..... 2-393B-1.OUT Data set title.... MONITORING WELL 2-393B: RUN 1 Company..... PARSONS ES Location..... OKLAHOMA Test date..... AUGUST 1997 Units of Measurement Length..... ft Time.... min Test Well Data Initial displacement in well..... 1.475 Radius of well casing..... 0.167 Radius of wellbore..... 0.354 Aguifer saturated thickness..... 29.14 Well screen length..... 10 Static height of water in well... 29.14 Gravel pack porosity...... 0.35 Effective well casing radius.... 0.249 Effective wellbore radius..... 0.354 Log (Re/Rw) 3.153 Constants A, B and C........... 0.000 , 0.000, 1.914 No. of observations............ 185 ______ ANALYTICAL METHOD Bouwer-Rice (Unconfined Aguifer Slug Test) RESULTS FROM VISUAL CURVE MATCHING

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 1.2220E-002 ft/min

y0 = 1.2606E + 000 ft

S ES	.34050	RUN 1	DATA SET: 2-393B-1.DAT 11/21/97 AQUIFER MODEL: Unconfined SOLUTION METHOD: Bouwer-Rice TEST DATA: H0= 1.475 ft r_= 0.167 ft r_= 0.167 ft L = 10. ft b = 29.14 ft H = 29.14 ft H = 29.14 ft H = 29.14 ft OBRAMETER ESTIMATES: K = 0.01222 ft/min y0 = 1.261 ft	SOLU
COMPANY: PARSONS ES	PROJECT: 729691,34050	MONITORING WELL 2-393B:	2. 3. 4. 5. 6. 7. 8. Time (min)	
CLIENT: TINKER AFB	LOCATION: OKLAHOMA		Displacement (ft) 0. 1. 2. 2. 2. 2. 2. 2. 2. 2. 2. 2. 2. 2. 2.	•

AQTESOLV RESULTS Version 2.0

Developed by Glenn M. Duffield (c) 1993, 1994 Geraghty & Miller, Inc.

(c) 1993, 1994 Geraghty & Miller, Inc.									
11/21/97 10:51:19									
TEST DESCRIPTION									
Data set									
Units of Measurement Length ft Time min									
Test Well Data Initial displacement in well									
ANALYTICAL METHOD									
Bouwer-Rice (Unconfined Aquifer Slug Test)									
RESULTS FROM VISUAL CURVE MATCHING									
VICIDI, MATCU DADAMETED ECTIMATES									

VISUAL MATCH PARAMETER ESTIMATES

Estimate

K = 1.2421E-002 ft/min

y0 = 1.1485E + 000 ft

B COMPANY: PARSONS ES	A PROJECT: 729691,34050	MONITORING WELL 2-393B: RUN 2	Defin SET: 2-3538-2.Dof 11-3538-2.Dof 11-3	\(\text{SOT}\)
CLIENT: TINKER AFB	LOCATION: OKLAHOMA			

LEMKE LAND SURVEYING, LLC

3625 W. Main, Suite 106 - Norman, OK 73072
Ph (405) 366-8541 FAX (405) 360-6670 E-mail LLSSKL@aol.com
C.A. No. 2471 (expires 6/30/99)

PARSONS ENGINEERING SCIENCE, INC
FORMER FIRE TRAINING PITS - 6 MONITOR WELLS
TINKER AIR FORCE BASE
BRUCE HENRY
HORIZONTAL - NAD27 (US SURVEY FEET)
VERTICAL - NGVD29 (US SURVEY FEET)

NORTHING (FT)	EASTING (FT)	ELEVATION (FT)	DESCRIPTION
151,085.60	2,181,167.63	1,241.82	CG39-B97-45 TOC
151,088.16	2,181,169.18	1,242.24	CG39-B97-45 GND
150,861.52	2,181,770.08	1,252.42	CG39-B97-43S TOC
150,863.36	2,181,772.35	1,252.72	CG39-B97-43S GND
150,866.94	2,181,777.04	1,252.54	CG39-B97-43B TOC
150,869.21	2,181,779.13	1,252.56	CG39-B97-43B GND
150,440.90	2,182,126.70	1,245.20	CG39-B97-41 TOC
150,443.53	2,182,126.53	1,245.57	CG39-B97-41 GND
151,434.08	2,181,280.87	1,245.97	2-393B TOC
151,436.52	2,181,282.08	1,246.19	2-393B GND
150,696.27	2,181,430.89	1,244.99	2-392B TOC
150,697.93	2,181,432.62	1,245.30	2-392B GND

HORIZONTAL DATUM: NAD 27 (US SURVEY FEET) VERTICAL DATUM: NGVD 29 (US SURVEY FEET) CONTROL MONUMENTS: SE30, SE32, SE28

Date of Survey: September 11, 1997

Crew: Arce / Young

Steven K. Lemke

Weather: Clear, calm, warm

Oklahoma Registered Land Surve

Page 1 of 1

LEMKE

APPENDIX C LABORATORY ANALTYICAL DATA



Ref: 97-LB53 August 6, 1997

Dr. Don Kampbell National Risk Management Research Laboratory Subsurface Protection and Remediation Division U.S. Environmental Protection Agency P.O. Box 1198 Ada, OK 74820

THRU: Steve VandegriftS

Dear Don:

Please find attached the analytical results for the Tinker AFB Service Request SF-3-276 requesting the analysis of ground water samples to be analyzed for benzene, toluene, ethylbenzene, p-, m-, and o-xylene, 1,3,5-, 1,2,4-, and 1,2,3-trimethylbenzene, and total fuel carbon. We received your 24 samples, most in duplicate, August 4, 1997 in capped, lead lined 40 mL VOA vials. The samples were analyzed on August 4-5, 1997. Samples were stored at 4°C until analyzed. All samples were acquired and processed using the Millennium data system. A 3 point (1-100 ppb) external calibration curve was used to determine the concentration for all compounds.

RSKSOP-133 "Simultaneous Analysis of Aromatics and Total Fuel Carbon by Dual Column/Dual Detector Gas Chromatography in Ground Water Samples" was used for these analyses. Autosampling was performed using a Dynatech-Precision autosampler in-line with a Tekmar LSC 2000 sample concentrator.

Sincerely,

Lisa R. Black

xc: R.L. Cosby

G.B. Smith

J.L. Seeley

SampleName	BENZENE	TOLUENE	ETHYLBENZENE	p-XYLENE	m-XYLENE	o-XYLENE	1,3,5-TMB	1,2,4-TMB	1,2,3-TMB	Fuel Carbon
					:					
10 PPB	9.7	9.7	8.6	2.6	9.6	6.6	6.6	6.6	10.0	N/A
QC, OBSERVED, PPB	19.7	20.3	19.0	20.8	19.9	21.2	20.9	21.9	19.1	N/A
QC, TRUE VALUE, PPB	20.0	20.0	20.0	20.0	20.0	20.0	20.0	20.0	20.0	N/A
GC LAB BLANK, PPB	2	2	2	2	2	2	Q	2	Q	A/A
2-62A	2	2	2	9	2	2	2	2	2	2
2-62B	0.9	1.2	9	2	Q	BLQ	BLQ	Q	Q	1140
2-63A	윤	2	9	2	Q	2	2	Q	2	2
2-63B	2	2	2	2	2	Q	Q	2	2	2
2-64A	2	2	2	2	2	2	2	Q	2	2
2-64A Duplicate	2	2	2	Q	2	Q	2	2	2	2
2-64B	2	Q	2	2	2	2	2	Q	2	Q
2-65A	2	2	2	2	2	2	2	2	2	2
2-65B	Q	2	9	2	2	2	Q	2	2	2
2-167B	2	2	9	2	2	2	2	Q	9	2
2-168B	2.5	2	2	2	2	2	2	2	2	5.8
1 PPB	1.0	6.0	1.0	1.0	1.0	6.0	0.0	1.0	1.1	∀ N
2-272B	2	2	2	2	9	2	<u>Q</u>	2	2	Q
2-273B	2	2	2	2	2	2	2	2	2	2
2-273B Duplicate	2	2	2	2	2	Q	2	g	2	2
2-142A	2	2	2	2	2	2	2	2	2	Q
2-142B	2	2	2	2	2	Q	9	2	2	<u>Q</u>
2-2718	Q	Q	2	2	Q	Q	Q	2	2	2
2-274A	2	2	2	2	2	Q	2	2	2	2
2-274A Duplicate	2	2	2	Q	2	2	2	2	2	2
2-274B	Q	2	2	Q	9	Q	2	9	2	2
2-392B	S	2	2	Q	2	Q	2	2	2	2
2-393B	BLQ	2	2	2	2	Q	2	2	2	BLQ
50 PPB	50.4	20.7	50.5	6.03	50.4	51.2	51.1	51.9	51.9	A/N
2-355B	2	2	2	Q	2	2	2	2	2	2
CG-39-B97-41	2	Q	2	Q	2	Q	9	2	2	Q
CG-39-B97-43S	2	2	2	Q	2	2	2	2	2	2
* CG-39-B97-43D	2	2	2	S	2	2	2	2	2	9
CG-39-B97-45	2	2	2	S	2	<u>Q</u>	2	2	2	2
1 PPB	1.0	6.0	1.0	6.0	6.0	6.0	0.8	6.0	1.0	A/A

^{*} Septa was inverted on the original sample and it's duplicate.

ND = None Detected; BLQ = Below Limit of Quantitation, 1 ppb; N/A = Not Analyzed



Ref: 97-SH51

August 6, 1997

Dr. Don Kampbell National Risk Management Research Laboratory Subsurface Protection & Remediation Division U.S. Environmental Protection Agency

P.O. Box 1198 Ada, OK 74820

THRU: S.A. Vandegrift 5

Dear Don:

Attached are TOC results for 25 liquids submitted August 6, 1997 under Service Request #SF-3-276. Sample analysis was begun August 6, 1997 and completed August 6, 1997 using RSKSOP-102.

Blanks, duplicates, and AQC samples were analyzed along with your samples, as appropriate, for quality control. If you have any questions concerning this data, please feel free to ask me.

Sincerely,

Sharon Hightower

xc: R.L. Cosby

G.B. Smith
J.L. Seeley

KAMPBELL TINKER LIQUIDS SF-3-276

SAMPLE '	MG/L TOC
2-62A 2-62B 2-63A 2-63B 2-64A 2-64B 2-65A 2-65B 2-142A 2-142B 2-167B 2-168B 2-271B 2-272B 2-273B 2-273B 2-274A 2-274A 2-355B 2-392B 2-393B CG39-B97-41 DUP	1.53 3.30 .745 11.2 .152 .795 <.4 1.96 2.14 .929 .488 3.01 .939 .448 .638 <.4 <.4 <.4 <.4 <.4 <.4 <.4 <.4 <.4 <.4
	<.4
	.539
WS37	58.7
	2-167B 2-168B 2-271B 2-272B 2-273B 2-274A 2-274B 2-355B 2-392B 2-393B CG39-B97-41 DUP CG39-B97-43D CG39-B97-45 CG39-B97-435



Ref: 97-MW50/vg 97-LP80/vg

August 6, 1997

Dr. Don Kampbell National Risk Management Research Laboratory Subsurface Protection & Remediation Division. U.S. Environmental Protection Agency P.O. Box 1198 Ada, OK 74820

THRU: S.A. Vandegrift 5V

Dear Don:

Attached are the results of 24 field samples from Tinker AFB, OK submitted to MERSC as part of Service Request #SF-3-276. The samples were received on August 4 and 5, 1997 and analyzed immediately. The methods used for analysis were EPA Methods 353.1 for NO2 and NO₃, 350.1 for NH₃, and Waters capillary electrophoresis Method N-601 for Cl and SO₄. Quality assurance measures performed on this set of samples included spikes, duplicates, known AQC samples and blanks.

If you have any questions concerning these results, please feel free to contact us.

Mark White

Synda Pennington

xc. R.L. Cosby G.B. Smith J.L. Seeley SV

Sample	mg/l NO ⁻ 2+NO ⁻ 3(N)	mg/l NH ₃ (N)	mg/l Cl	mg/1 SO ₁ -2
2-62A	4.02	<.05	7.54	6.51
2-62B	3.92	<.05	173	127
2-63A	1.91	<.05	20.5	8.46
2-63B	2.39	<.05	94.4	65.7
2-64A	1.29	<.05	21.5	7.96
2-64A Dup			21.6	7.86
2-64B	3.42	<.05	49.3	27.0
2-65A	0.72	<.05	25.6	9.89
2-65B	3.18	<.05	78.5	44.4
2-65B Dup	3.19	<.05		
2-167-B	0.27	<.05	4.49	18.1
2-168-B	0.58	<.05	14.4	42.4
2-272-B	1.95	<.05	130	37.6
2-273-B	0.35	<.05	16.5	12.1
2-273-B Dup	0.34	<.05	16.4	12.2
2-142-A	<.05	<.05	28.4	9.72
2-142-B	2.00	<.05	218	76.1
2-142-B Dup	2.03	<.05		
2-271-B	1.11	.15	117	29.7
2-274-A	.62	<.05	18.0	10.4
2-274-B	3.08	<.05	74.6	24.0
2-392-B	2.57	<.05	10.6	5.91
2-392-B Dup	2.63	<.05		
2-393-B	3.70	<.05	148	14.1
2-355-B	3.28	<.05	187	48.8
CG-39-B97-41	2.01	<.05	5.73	15.4
CG-39-B97-43S	4.22	<.05	159	49.0
CG-39-B97-43D	2.55	<.05	52.0	44.1
CG-39-B97-43D D	up		51.7	45.0
CG-39-B97-45	4.80	<.05	46.9	11.3
Blank	<.05	<.05	<.1	<.1
AQC	.58	.25	35.0	43.9
True Value	.62	.26	34.8	44.0
Spike Rec.	99%	99%	100%	99%

•



Ref: 97-LH40

August 8, 1997

Dr. Don Kampbell
National Risk Management Research Laboratory
Subsurface Protection and Remediation Division
U.S. Environmental Protection Agency
P.O. Box 1198
Ada, OK 74820

THRU: S.A. Vandegrift 5

Dear Don:

As requested in Service Request #SF-3-276, gas analysis was performed for methane, ethylene, and ethane on samples from Tinker AFB. The samples were received on August 4 and 5, 1997, and analyzed on August 5, 1997. These analyses were performed as per RSKSOP-194, and the calculations were done as per RSKSOP-175.

If you should have any questions, please feel free to contact me.

Sincerely,

Lisa Hopkins

xc:

R.L. Cosby

J.T. Wilson

G.B. Smith

J.L. Seeley

Sample	Methane	Ethylene	Ethane
100ppm CH4	100.8	**	**
100ppm C2H4	**	96.1	**
100ppm C2H6	**	**	102.5
HPHe	**	**	**
Lab Blank	**	**	**
2-62A	**	**	**
2-62B	1.05	**	**
2-63A	**	**	**
2-63B	0.002	***	**
2-64A	**	**	**
2-64A	**	**	**
Lab Dup			
2-65A	**	**	**
2-64B	**	de de	**
2-65B	0.002	***	**
2-167B	0.022	**	**
2-168B	0.020	**	**
2-168B	0.020	**	療療
Field Dup	0.020		
10ppm CH4	10.3	**	**
2-272B	**	**	**
2-272B	0.005	**	**
2-142A	**	**	**
2-142B	**	**	**
2-271B	**	**	**
2-271B	**	**	**
Lab Dup			
2-274A	**	**	**
2-274B	**	**	**
2-355B	**	**	**
2-392B	**	**	**
2-393B	**	**	**
2-393B	**	**	**
Field Dup			,
10ppm C2H4	**	9.2	**
CG-39-B97-41	**	**	**
CG-39-B97-43S	**	**	**
CG-39-B97-43D	**	**	**
CG-39-B97-45	**	**	**
CG-39-B97-45	**	**	**
Lab Dup			
100ppm CH4	100.6	**	**
100ppm C2H4	**	99.3	**
100ppm C2H6	**	**	100.9
Lower Limit of Q		0.000	0.000
	0.001	0.003	0.002

Units for the samples are mg/L.
Units for the standards are parts per million.
*** denotes None Detected.

^{*} denotes Below Limit of Quantitation.



97-MAB9/vg August 11, 1997

Dr. Don Kampbell
National Risk Management Research Laboratory
Subsurface Protection & Remediation Division
U.S. Environmental Protection Agency
P.O. Box 1198
Ada, OK 74820

THRU: S.A. Vandegrift 5

Dear Don:

As per Service Request #SFTA-3-132, total alkalinity, hydrogen sulfide, ferrous iron, carbon dioxide, total phenols, and manganese were measured in 24 water samples from Tinker AFB, OK. The analysis began on July 31, 1997 and was completed on August 4, 1997.

	WELL	CO ₂	Fe +2	Total Alk.	Phene	ols l	Mn	H ₂ S
	2-63B	75	<.01	460	<.01	<.01	<.0)1
	2-63A	50	1.0	380	<.01	<.01	<.0)1
	2-168B	65	<.01	380	<.01	<.01	<.(D1
	2-62B	55	<.01	260	<.01	<.01	<.(01
	2-62A	45	0.10	310	<.01	<.01	<.(01
	2-64B	40	<.01	260	<.01	<.01	<.(01
	2-64A	45	0.10	270	<.01	<.01	<.()1
	2-73B	45	0.20	280	<.01	<.01	<.0)1
	2-272B	45	<.01	280	<.01	<.01	<.(01
	2-167B	75	0.10	560	0.10	<.01	<.0)1
	2-65B	75	<.01	350	<.01	<.01	<.(01
	2-65A	25	0.30	190	<.01	<.01	<.(01
	2-142A	45	<.01	270	0.20	<.01	<.(01
	2-142B	55	0.80	380	0.20	<.01	<.0)1
	2-274B	45	0.20	310	0.20	<.01	<.0)1
	2-274A	25	<.01	210	0.40	<.01	<.0	01
	2-271B	65	0.60	420	0.30	<.01	<.0	01
	2-355B	50	0.40	380	0.30	<.01	<.(01
	2-392B	60	<.01	350	0.20	<.01	<.(01
	39-B97 - 43D	75	<.01	450	0.20	<.01	<.0	01
	39-B97 - 41	60	<.01	520	<.01	<.01	<.	01
	39-B97-43S	48	<.01	330	0.20	<.01	<.	01
,	39-B97-45	75	<.01	380	<.01	<.01	<.	01
	2-393B	110	<.01	190	<.01	<.01	<.	01

ManTech Environmental Research Services Corporation

Note: All the data is in Mg\L.

If you have any questions, please see us at your convenience.

Sincerely,

Brad Scroggins

Mark Blankenship

xc: R.L. Cosby G.B. Smith J.L. Seeley



Ref: 97-DK22/vg August 11, 1997

Dr. Don Kampbell
National Risk Management Research Laboratory
Subsurface Protection and Remediation Division
U.S. Environmental Protection Agency
P.O. Box 1198
Ada, OK 74820

THRU: S.A. Vandegrift

Dear Dr. Kampbell:

This report contains the results of my GC/MSD analysis of methylene chloride core sample extracts for Service Request #SF-3-276. The following target compound quantitations were requested: trichloroethene (TCE) and tetrachloroethene (PCE).

The analytical method was a modification of RSKSOP-124. Cool on-column injection (0.1 μ l) was used with electronic pressure control set for a constant flow of 1.0 ml/min. The capillary GC column consisted of a 30m X 0.25mm Restek Stabilwax (Crossbonded Carbowax-PEG, 0.5 μ m film) plus an SGE 0.3m X 0.53 mm ID deactivated Carbowax deactivated capillary precolumn. SIM mode GC/MSD was used with the ions chosen from those listed in EPA method 524.2 Revision 3.0, where available. Multiple ions were acquired and ion ratios used to verify the accuracy of target compound identification. Standards calibration ranged from 0.025 to 250 μ g/ml and was divided into a low level (0.025 to 2.5 μ g/ml) and high level (2.5 to 250 μ g/ml) curve for improved quantitative accuracy. Complete reports detailing the acquisition method and calibration curves have been recorded. The extracts were received on August 4, 1997 and SIM mode analyses for quantitation of target compounds performed August 5, 1997.

The quantitative results are as follows (mg/kg):

Sample	<u>TCE</u>	<u>PCE</u>
{1135 MPI 18-19' (Rep. 1)	0.019	ND
21135 MPI 18-19' (Rep. 2)	0.070	ND
(0855 MPL 29-30' (Rep. 1)	ND	ND
0855 MPL 29-30' (Rep. 2)	ND	ND
Check Standards (ug/ml)		
0.25	0.24	0.24
25	26.1	26.2
QC (expected)	50	N/A
QC (observed)	47.7	N/A

If you require further information, please feel free to contact me.

xc: R.L. Cosby

J.L. Seeley

ManTech Environmental Research Services Corporation

R.S. Kerr Environmental Research Laboratory, P.O. Box 1198, 919 Research Drive Ada, Oklahoma 74821-1189 405-436-8660 FAX 405-436-8501

C659-697



Ref: 97-DF38

Aug. 13, 1997

Dr. Don Kampbell
National Risk Management Research Laboratory
Subsurface Protection and Remediation Division
U.S. Environmental Protection Agency
P.O. Box 1198
Ada, OK 74820

THRU: S.A. Vandegrift ≤√

Dear Don:

As requested in Service Request SF-3-276, GC/MS analysis for phenols and aliphatic/aromatic acids was done on one water sample from the Tinker AFB. This sample was labeled: CG-39-B97-43S. The sample was received on Aug. 6, 1997 and was extracted and derivatized by Amy Zhao on Aug. 8, 1997. The extracts were analyzed by GC/MS on Aug 8, 1997. RSKERL SOP 177 was used for the extraction, derivatization and GC/MS analysis of the samples.

Table I provides the concentrations of the phenols and aliphatic/aromatic acids found in the Tinker AFB water sample. Derivative and extraction blanks, extraction recoveries and 100 ppb check standards are also included in the table.

If you should have any questions, please feel free to contact me.

Sincerely,

Dennis D. Fine

xc: J. Seeley

R. Cosby

D. Fine

J. Wilson

Table I. Quantitative Report and QC Data for Phenois and Aliphatic and Aromatic Acids in Water Samples from Tinker AFB, Service Request SF-3-276.

Concentration ppb % Recovery 106CG39B.TXT 88MTHBLK.TX 98EXTBLK.TXT 116EXTBL.TXT 97EXTREC.TX 115EXTRE.TXT 110100AA.TXT File Name: Method Extraction Extraction 50 ppb Extr. 50 ppb Extr. 100 ppb Std. CG-39-897-43S Compound Sample Name Blank Blank Blank % Recovery % RecNvery % Recovery Propanoic Acid 23 52 97 11 ••• *** 2-Methylpropanoic Acid 33 28 93 ... Trimethylacetic Acid 14 15 12 110 107 96 ••• **Butyric Acid** 5 *** 37 28 97 ••• 2-Methylbutyric Acid *** 76 71 96 *** ••• 3-Methylbutryic Acid 72 67 98 3,3-Dimethylpropanoic Acid 103 90 96 Pentanoic Acid 82 74 90 2,3-Dimethybutyric Acid *** *** 99 92 91 *** 2-Ethylbutyric Acid *** 104 89 91 *** 2-Methylpentanoic Acid 107 92 96 *** *** 3-Methylpentanoic Acid *** 102 93 92 4-Methylpentanoic Acid *** 105 94 93 Hexanoic Acid 5 ••• 14 5 128 101 91 2-Methylhexanoic Acid ••• 115 89 82 Phenol *** *** 97 79 87 Cyclopentanecarboxylic Acid *** *** 86 79 94 5-Methylhexanoic Acid *** *** 114 95 93 o-Cresol *** *** 105 97 96 2-Ethylhexanoic Acid 21 *** *** *** 122 103 94 *** Heptanoic Acid 6 *** 123 98 92 *** m-Cresol *** *** 107 95 94 *** p-Cresol *** 102 93 91 1-Cyclopentene-1-carboxylic Acid *** 81 73 93 o-Ethylphenol *** 103 96 91 Cyclopentaneacetic Acid *** *** 107 94 92 2,6-Dimethyphenol 86 83 92 2,5-Dimethylphenol 96 88 90 Cyclohexanecarboxylic Acid *** 107 94 93 3-Cyclohexene-1-carboxylic Acid ••• *** 89 85 91 2,4-Dimethylphenol 71 74 95 3,5-Dimethylphenol & m-Ethylphenol *** 107 92 93 Octanoic Acid 11 3 129 99 92 2,3-Dimetholphenol 95 85 93 p-Ethylphenol 105 98 93 Benzoic Acid 9 117 101 92 3,4-Dimethyphenol 95 89 92 m-Methylbenzoic Acid 89 82 89 1-Cyclohexene-1-carboxylic Acid 108 91 93 Cyclohexaneacetic Acid 110 93 91 2-Phenylpropanoic Acid 113 95 92 o-Methylbenzoic Acid 109 96 95 *** Phenylacetic Acid 109 94 90 m-Tolylacetic Acid 176 124 134 o-Tolylacetic Acid 102 99 108 2,6-Dimethylbenzoic Acid 114 99 89 p-Tolylacetic Acid 111 93 91 p-Methylbenzoic Acid 106 95 94 3-Phenylpropanoic Acid 102 87 96 2,5-Dimethylbenzoic Acid 109 91 94 Decanoic Acid 22 111 89 92 2,4-Dimethylbenzoic Acid 104 98 91 3,5-Dimethylbenzoic Acid 119 89 93 2,3-Dimethylbenzoic Acid 108 95 94 4-Ethylbenzoic Acid 107 93 94 2.4.6-Trimethylbenzoic Acid 110 95 93 3,4-Dimethylbenzoic Acid 87 108 93 2,4,5-Trimethylbenzoic Acid 93

indicates not found.

^{***} indicates concentration of extract was below lowest calibration standard (3 ppb)



Ref: 97JAD51

August 14, 1997

Dr. Don Kampbell National Risk Management Research Laboratory Subsurface Protection and Remediation Division U.S. Environmental Protection Agency P.O. Box 1198 Ada, OK 74820

THRU: S.A. Vandegrift $\leq V$

Dear Don:

As requested in Service Request # SF-3-276, headspace GC/MS analysis of 24 Tinker AFB water samples for chlorinated volatile oraganics was completed. The samples were received on August 4 and 5, 1997 and analyzed on August 8-9, 1997. RSKSOP-148 (Determination of Volatile Organic Compounds in Water by Automated Headspace Gas Chromatography/Mass Spectrometry (Saturn II Ion Trap Detector) was used for this analysis.

An internal standard calibration method was established for the 15 compounds. The standard curves were prepared from 1.0 to 2000 ppb. The lower calibration limits were 1.0 ppb.

If you should have any questions, please feel free to contact me.

John Allen Daniel

xc: R.L. Cosby

G.B. Smith

D.D. Fine J.L. Seeley

J.T. Wilson

Table 1. Quantitation Report for S.R. # SF-3-276 from Tinker AFB.

Concentration = ppb

2-142A 2-142B 2-271B		2-62B 2-63A Field Dup 1/10 Dil ND ND ND ND ND ND ND ND ND ND ND ND ND
CG39-B97 2-1 -435	1.1 0.1 1.0 0.1 1.0 0.1 0.1 0.1 0.1 0.1	ND 1.7 ND 9.5 ND 76.2 ND 76.2 ND 1200 ND ND
CG39-B97 C -41		CG39-B97 2 -45 ND ND ND ND 1.4 153 ND ND ND ND ND ND ND ND ND ND ND ND ND
2-355B	DN	CG39-B97 -43D ND ND S:4 ND ND ND ND ND ND ND ND ND ND ND ND ND
2-393B Field Dup	N	2-274B Field Dup ND ND S3.4 ND ND 1.4 506
2-393B	ON 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	2-274B Lab Dup ND ND 51.4 ND ND 1.2 529
2-392B	22222222 2222	2-274B ND 2.2 ND 51.7 ND 1.1 518
Compound	VINYL CHLORIDE 1,1-DICHLOROETHENE 1,1-DICHLOROETHENE 1,1-DICHLOROETHANE C-1,2-DICHLOROETHENE CHLOROFORM 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE 1,2-DICHLOROETHANE TRICHLOROETHENE TRICHLOROETHENE TETRACHLOROETHENE CHLOROBENZENE 1,3-DICHLOROBENZENE 1,4-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,2-DICHLOROBENZENE	VINYL CHLORIDE 1,1-DICHLOROETHENE 1,1-DICHLOROETHANE 1,1-DICHLOROETHANE C-1,2-DICHLOROETHANE CHLOROF ORM 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE 1,2-DICHLOROETHENE TRICHLOROETHENE TETRACHLOROETHENE TETRACHLOROETHENE THICHLOROETHENE THICHLOROETHENE THICHLOROETHENE THICHLOROETHENE THICHLOROETHENE THICHLOROETHENE THICHLOROETHENE THICHLOROETHENE THICHLOROBENZENE 1,3-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,2-DICHLOROBENZENE

ND = None Detected --- = Below Calibration Limit(1.0 ppb) **** = Above Calibration Limit(2000 ppb) Dup = Duplicate Dil = Dilution

Table 2. Quantitation Report for S.R. # SF-3-276 from Tinker AFB.

Concentration = ppb

QC0808B 200 ppb	189 195 195 188 204 204 207 195 205 232 238 238	
2-273B	222222222222222	
2-272B	0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	
2-168B	D	
2-1678	222222222222222	₹
2-65B Lab Dup	1.5 2.5 36.9 ND ND ND 87.7 50.7 9.6 1.8	BL0808A ND ND
2-65B	1.3 2.5 37.0 ND ND 3.2 90.0 52.0 9.5 1.8	20 ppb 17.0 24.5 24.5 24.5 21.0 20.0 20.7 20.2 20.3 20.3 20.3
2-65A	222222222222222	QC0808E 20 ppb 18.1 25.3 20.6 20.2 19.3 19.3 19.0 19.0 20.2 20.2 20.3
2-64B	17.6 ND 154 ND ND 154 1.3 914	QC0808D 200 ppb 218 194 186 203 204 204 194 206 195 191 200 234 234
2-64A	222222222222222	20 ppb 16.6 16.6 19.9 19.9 19.5 18.7 18.5 18.5 18.0 20.3 20.3
Compound	VINYL CHLORIDE 1,1-DICHLOROETHENE 1,1-DICHLOROETHENE 1,1-DICHLOROETHANE C-1,2-DICHLOROETHENE CHLOROFORM 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE 1,2-DICHLOROETHENE TRICHLOROETHENE TETRACHLOROETHENE TETRACHLOROETHENE CHLOROBENZENE 1,3-DICHLOROBENZENE 1,4-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,3-DICHLOROBENZENE	VINYL CHLORIDE 1,1-DICHLOROETHENE 1,1-DICHLOROETHENE 1,1-DICHLOROETHANE C-1,2-DICHLOROETHANE CHLOROF ORM 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE 1,2-DICHLOROETHENE TETRACHLOROETHENE TETRACHLOROETHENE TETRACHLOROETHENE THICHLOROBENZENE 1,3-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,2-DICHLOROBENZENE 1,2-DICHLOROBENZENE

ND = None Detected --- = Below Calibration Limit(1.0 ppb) QC = Quality Control Std. BL = Blank Dup = Duplicate



Ref: 97-SH53

August 18, 1997

Dr. Don Kampbell National Risk Management Research Laboratory Subsurface Protection & Remediation Division U.S. Environmental Protection Agency P.O. Box 1198 Ada, OK 74820

THRU: S.A. Vandegrift SV

Dear Don:

Attached are TOC results for six Tinker soils submitted August 9, 1997 under Service Request #SF-3-276. Sample analysis was begun August 11, 1997 and completed August 18, 1997 using RSKSOP-102 and RSKSOP-120.

Blanks, duplicates, and AQC samples were analyzed along with your samples, as appropriate, for quality control. If you have any questions concerning this data, please feel free to ask me.

Sincerely,

Sharon Hightower

xc: R.L. Cosby

G.B. Smith

J.L. Seeley

KAMPBELL TINKER SOIL SF-3-276

	FILTRATE O.C.	SOLIDS %TOC	TOTAL SOIL % TOC	MEAN STD %TOC DEV
MP2 0855,1-1 1-2 1-3	.832	<.001 <.001 <.001	.894 .832 .869	.865 .031
MW1 1335,1-1 1-2 1-3		<.001 <.001 <.001	.969 .632 .888	.830 .176
MW1 1410,1-1 1-2 1-3	.860	<.001 <.001 <.001	.806 .860 .790	.819 .037
Z-395B,1-1 1-2 1-3		<.001 <.001 <.001	.857 .638 .697	.731113
Z-593B,1-1 1-2 1-3	.681 .747 .869	.003 .008 .004	.684 .755 .873	.771 .095
MP1 1135,1-1 1-2 1-3	.939 .803 .805	.013 .013 .014	.952 .816 .819	.862 .078
WS37 LECO STD	55.3	.921		

WS37 std t.v.=60.0 Leco soil std t.v.=.90+/-.04

Attention: Scott Boling

Brown & Root Environmental \$500 South Douglas Blvd. Oaklahoma City, OK 73150 Date of Report: 10/21/97
Date Collected: 09/04/97
Date Received: 09/05/97
Collected By: AJS

Cuet. Proj: Tinker AFB INW&B#1

Sample Desc: CG39 2-356B

Lab ID: 867403

				Det.	Diluth		Test
		Result	Dnit	Limit	Factor	Procedure	Date
		***************************************				• • • • • • • • • • • • • • • • • • • •	
DRGANIC							
	TILES - GC/MS						
	1 Chloride	BDL	DG/L	10	1	8270	09/27/97
-	Methacrylate	BDL	UG/L	10	1	8270	09/27/97
	phthene	BDL	UG/L	10	1	8270	09/27/97
	phthylene	BDL	UG/L	10	2	9270	09/27/97
••••	eneos	BDL	UG/L	10	1	8270	09/27/97
Benzi		BDL	DG/L	15	1	8270	09/27/97
Benzo	(a) anthracone	BDL	UG/L	15	1	\$270	03/27/97
Benzo	(a) pyrche	BDL	DG/L	15	1	8270	09/27/97
Benze	(b) fluoranthene	BDL	UG/L	15	1	8270	09/27/97
Benzo	(g,h.i)perylene	BDL	UG/L	20	1	8270	09/27/97
Benzo	(k) fluoranthene	BDL	UG/L	30	1	2270	09/27/97
bis (2	-Chlorosthoxy) methane	BDL	UG/L	10	1	8270	09/27/97
bia(2	-Chloroethyl) ether	BDL	UG/L	10	x	8270	09/27/97
bis (2	-Chloroisopropyl) cther	EDL	UG/L	15	1	8270	09/27/97
bis(2	-Ethylhexyl) phthalate	BDL	UG/L	15	1	\$270	09/27/97
4-Bro	mophenyl phenyl ether	BDL	DG/L	10	1	827 0	09/27/97
Bucyl	benzyl phthalate	BDL	UG/L	15	1	8270	09/27/97
2-ch1	oronaphthalane	EDL	DG/L	10	1	8270	09/27/97
4-Ch1	orophenyl phenyl ether	EDL	UG/L	15	1	8270	09/27/97
Custo	one	BDL	UG/L	15	1	6270	09/27/97
Diber	zo(a, h) anthracene	BDL	UG/L	20	1	8270	09/27/97
1,2-0	ichlorobenzene	EDL	ng/r	15	1	8270	09/27/97
1,3-5	ichlorobenzene	BDL	ug/L	15	1 .	8270	09/27/97
1,4-D	oichlorobenzene	BDL	Ua/L	10	1	8270	09/27/97
3,3'-	Dichlorobenzidine	BDL	UG/L	15	1	8270	09/27/97
Dieth	yl phthalate	BDL	UG/L	10	1	8270	09/27/97
Dimet	hyl phthalate	BDL	UG/L	15	1	8270	09/27/97
Di-n-	butyl phthalate	BDL	UG/L	40	1	8270	09/27/97
2,4-D	initrotoluens	BDL	UG/L	10	1	8270	09/27/97
2,6-0	initrotoluene	BDL	Ua/L	10	1	9270	09/27/97
•	octyl phthalate	BDL	Up/L	20	1	8270	09/27/97
	iphenylhydrazine	BDL	NG/L	10	1	8270	09/27/97
Fluor	anthene	BDL	DG/L	15	1	6270	09/27/97
Fluor	rene	BDL	UG/L	10	1	6270	09/27/97
Vevec	hlorobenzene	BDL	UG/L	10	1	8270	09/27/97



Attention: Scott Boling

Brown & Root Environmental 5500 South Douglas Blvd. Oaklahoma City, OK 73150 Date of Report: 10/21/97 Date Collected: 09/04/97 Date Received: 09/05/97

Collected By: AJS

Cust. Proj: Tinker APB IMW&B#1

Eample Desc: CG39 2-356B

Lab ID: 867403

				Det.	Diluen		Test
		Result	Unit	Limit	Factor	Procedure	Date
		••					
RGANI	c						
VOLA:	IILES - GC/MS						
	1,1,1,2-Tetrachloroethane	EDL	ng/r	. 8	1	8260	09/12/97
	1,2,3-Trichloropropane	EDL	DG/L	1	1	0260	09/12/97
	Dromobenzene	BDL	UG/L	. 5	1	8260	09/12/97
	Dibromomethane	BDL	UG/L	. 8	1	8260	08/12/97
	Xylenes (Total)	BDL	UG/L	. 5	1	8260	09/12/97
	2,2-Dichloropropane	BDL	ug/L	, 5	1	8260	09/12/97
	1,1-Dichloropropene	BDL	us/L	1	1	8360	09/12/97
	1,3-Dichloropropane	BDL	UG/L	. 8	1	8260	09/12/97
	Isopropylbenzena	MDL	DG/L	. 8	1	8250	09/12/97
	t-Butylbenzene	DDL	UG/L	. 8	1	8260	09/12/97
	n-Propylbenzene	BDL	ひな/と	. 6	1	8260	08/12/97
	sec-Butylbenzene	BDL	UG/L	. 8	1	8260	09/12/97
	1,3,5-Trimethylbensana	BDL	UG/L	. 5	1	8260	09/12/97
	1,2,4-Trimethylbenzene	PDL	UG/L	. 5	1	8260	09/12/97
	p-Isopropyltoluene	BDL	DG/L	, 8	1	8360	09/12/97
	n-Bucylbenzene	BDL	ug/L	. 8	1	\$260	09/12/97
	1,2-Dichlorobenzenc	DDL	UG/L	. 5	1	8260	09/13/97
	1,1-Dichlorobenzene	BDL	UG/L	. 5	1	8260	09/12/97
	1,4-Dichlorobenzene	BDL	UG/L	. 5	1	8260	09/12/97
	Hexachlorobutadiene	BDL	UG/L	. \$	1	8260	09/12/97
	Maphthalene	EDL	UG/L	. 5	1	9260	09/12/97
	1,2,4-Trichlorobenzene	BDL	va/L	. 5	1	\$260	09/12/97
	1.2.3-Trichlorobenzene	BDL	UG/L	. 8	1	8260	09/12/97

Note 'BDL' = Below Detection Limit

Attention: Scott Boling

Stoum & Root Environmental \$500 South Douglas Blvd. Oaklahoma City, OK 73150 Date of Report: 10/21/97
Date Collected: 09/04/97
Date Received: 09/05/97

Collected By: AJS

Cust. Proj: Tinker AFB INW48#1

Sample Desc: CG39 2-356B

Lab ID: 967403

				Dec.	Dilutn		Test
		Result	Unit	Limit	Pactor	Procedure	Date
			•••••				
ORGANIC							
VOLATI	LES - GC/MS						
3	enzena	EDL	ua/L	. 8	1	8260	09/12/97
3	romoform	BDL	UG/L	.5	1	8260	09/12/97
c	arbon tetrachloride	BDL	UG/L	1	1	8260	09/12/97
C	hlorobeniene	BDL	UG/L	. 5	1	\$26 0	09/12/97
ם	ibromochloromethane	BDL	ng/r	. 5	1	8260	09/12/97
C	hloroethane	BDL	UG/L	, 5	1	8260	09/12/97
c	hloroform	BDL	UG/L	, B	1	8260	09/12/97
B	romodichloromethane	BDL	UG/L	. 5	1	8260	09/12/97
D	ichleredifluoromethane	BDL	ne/r	1	1	8260	09/12/97
1	,1-Dishloroethans	BDL	ng/r	. \$	1	8260	09/12/97
1	,2-Dichloroathane	BDL	DG/L	. 8	1	8260	09/12/97
1	,1-Dichloroethene	EDL	UG/L	. 8	1	5260	09/12/97
1	,2-Dichloropropane	EDL	UG/L	11	1	8260	09/12/97
•	is-1,3-Dichloropropene	adl	UG/L	. 5	1	8260	09/12/97
В	thylbenzene	BDL	DG/L	. 5	1	8260	09/12/97
В	romomethane	BDL	UG/L	. 5	1	8260	09/12/97
c	hloromethane	BDL	UG/L	1	1	8260	09/12/97
M	athylene chloride	, 9JB	UG/L	2	1	8260	09/12/97
1	,1,2,2-Tetrachloroethane	EDL	UG/L	. 5	1	B260	09/12/97
T	etrachloroethene	BDL	Ug/L	. 8	1	8260	09/12/97
T	oluene	BDL	UG/L	. 8	1	2260	09/12/97
7	rans-1,2-Dichloroethene	BDL	UG/L	. \$	1	2260	09/12/97
1	,1,1-Trichloroethana	BDL	AG\T	. 8	1	9260	09/12/97
1	,1,2-Trichloroethane	DDL	UG/L	. *	1	8260	09/12/97
T	richloroethene	BDL	UG/L	. 6 "	1	8260	09/12/97
T	richlorofluoromethanc	BDL	UG/L	1	1	8260	09/12/97
v	inyl chloride	BDL	DG/T	1	1	8260	09/12/97
	romochloromethane	BDL	UG/L	1	1	8260	09/12/97
c	ia-1,2-Dichloroethene	BDL	UG/L	. 5	1	8260	09/12/97
2	-Chlorotoluene	EDL	UG/L	. 5	1	8260	09/12/97
4	-Chlorotoluene	BDL	UQ/L	. 5	1	6260	09/12/97
1	.2-Dibromoethane (Ethylana dibromide)	BDL	UG/L	. 9	1	0260	09/12/97
T	rans-1,3-Dichloropropene	EDL	DG/L	. 6	1	82 60	09/12/97
s	tyrene	BDL	UG/L	. 5	1	8260	09/12/97
1	,2-Dibromo-3-chloropropane (DBCP)	BDL	UG/L	2	1	8260	D9/12/97



Attention: Scott Boling

Brown & Root Environmental 5500 South Douglas Blvd. Oaklahoma City, OK 73150

Data of Report: 10/21/97

Date Collected: 03/04/97 Date Received: 09/05/97

Collected By: AJS

Cust. Proj: Tinker APB IMW45#1

Sample Desc: CG39 2-356B

Lab ID:

867403

				Det.	Dilutn		Test
		Result	Unit	Limit	Factor	Procedure	Date
		************		*****			
organi	re .					•	
3BM I	-volatiles - GC/MS						
	Hexachlorocyclopentadicne	BDL	UG/L	10	1	8270	09/27/97
•	Hexachloroethanc	BDL	DG/L	15	1	6270	09/27/97
	Indeno(1,2,3-cd)pyrane	BDL	UG/L	25	1	8270	09/27/97
	Isophorone	BDL	UG/L	10	1	8270	09/27/97
	Mitxobenzene	BDL	UG/L	10	1	8270	Q9/27 /97
	N-Mitrosodimethylamine	BDL	UG/L	10	1	8270	09/27/97
	N-Nitroso-di-n-propylamino	BDL	UQ/L	10	1	8270	09/27/97
	N-Nitrosodiphenylaminc(1)	BDL	ng/r	15	1	8270	09/27/97
	Phenanthrene	BDL	DG/L	10	1	8270	09/27/97
	Pyrene	BDL	UG/L	15	1	8270	09/27/97
_	1,3,5-Trichlorobenzene	BDL	UG/L	10	1	8270	09/27/97
	Pentachloronitrobenzene	BDL	UG/L	15	1	8270	09/27/97
	1,2,3,4-Tetrachlorobenzeno	BDL	Ua/L	10	1	8270	09/27/97
	1,2,4,5-Tetrachlorobchacho	BDL	tg/L	35	1	8270	09/27/97
	Anilins	BDL	UG/L	15	1	8270	09/27/97
	Bentyl alcohol	BDL	UG/L	15	1	8270	09/27/97
	4-Chloroaniline	BDL	UG/L	10	1	8270	09/27/97
	Dibenzofuran	BDL	UG/L	10	1	8270	09/27/97
	2-Methylnaphthalene	BDL	UG/L	10.	1	8270	09/27/97
	2-Witrosniline	EDL	UG/L	15	1	8270	09/27/97
	3-Nitroaniline	BDL	UG/L	15	1	8270	09/27/97
	4-Nitroaniline	BDL	ng/r	20	1	8270	09/27/97
	Pyridine	BDL	UG/L	15	1	2370	09/27/97
	1-Naphthylamine	BDL	UG/L	30	1	8270	09/27/97
	2-Naphthylamine	BDL	UG/L	20	1	8270	09/27/97
	4-Aminobiphenyl	BDL	UG/L	10	1	9270	09/27/97
	p-Dimethylaminozzobenzene	BDL	UG/L	10	1	8270	09/27/97
	2-Acctylaminofluorene	EDL	UG/L	15	1	8270	09/27/97
	4,4'-Mcthylenc-bis(2-chloroaniline)	BDL	UG/L	15	1	8270	09/27/97
	5-Mitro-O-toluidine	BDL	DG/L	10	1	8270	09/27/97
	N-Nicrosodiethylamine	BDL	UG/L	15	1	8270	09/27/97
	Acetophenone	BDL	UG/L	15	1	8270	09/27/97
	M-Nitrosopiperidine	BDL	ng/r	10	1	8270	09/27/97
	safrole	BDL	ua/L	10	1	8270	09/27/97
	1,4-Naphthoquinone	BDL	UG/L	25	1	3270	09/27/97

Attention: Scott Boling

Brown & Root Environmental \$500 South Douglas Elvd. Oaklahoma City, OK 73150 Date of Report: 10/21/97
Date Collected: 09/04/97
Date Received: 09/05/97
Collected By: AJS

Cust. Proj: Tinker AFE IMW&B#1

Sample Desc: CGJ9 2-356B

Lab ID: 867403

				Dat.	Diluth	•	Test
		Result	Unit	Limit	Pagtor	Procedure	Date
		***************************************		22410		7,0094416	2 40e
ORGANI	đ	•					
SEMI	-Volatiles - GC/MS						
	Pentachloroethane	BDL	UG/L	15	1	8270	09/27/97
	Pentachlorobenzenc	BDL	DG/L	10 '	1	8270	09/27/97
	1,3,5-Trinitrobanzene	BDL	UG/L	20	1	8270	09/27/97
	2-Picoline	BDL	ua/L	25	1	8270	09/27/97
	o-Toluidine hydrochloride	BDL	UG/L	25	1	8270	09/27/97
	1,2,3,5-Tetrachlorobensene	BDL	UG/L	35	1	8270	09/27/97
	Sthyl methanesulfonate	BDL	UG/L	10	1	8270	09/27/97
	Methapyrilene	BDL	DG/L	45	1	8270	09/27/97
	7,12-Dimethylbenzanthracene	BDL	UG/L	15	1	8270	09/27/97
	Sensal chloride	BDL	UG/L	10	1	8270	09/27/97
_	Einophos	BDL	UG/L	10	1	8270	09/27/97
	Pronamide	BOL	UG/L	10	1	8270	09/27/97
	Chlorobenzilate	BDL	UG/L	15	1	8270	09/27/97
	Isosafrole	BDL	UG/L	15	1	6270	09/27/97
	N-Nitrocopyrrolidine	EDL	UG/L	15	1	B270	09/27/97
	Armmite	BDL	UG/L	10	1	8270	09/27/97
	Trans-Diallate	BDL	UG/L	10	1	8270	09/27/97
	Cis-Diallate	BDL	ud/L	10	1	8270	09/27/97
	3,3'-Dimethexybenzidine	BDL	UG/L	20	1	8270	09/27/97
	Benzotrichloride	BDL	DG/L	10	1	8270	09/27/97
	Nitrosomathylethylemine	BDL	UG/L	45	1	8270	09/27/97
	N-Mitroso-di-h-butylamine	BDL	UG/L	15	1	8270	09/37/97
	Hexachloropropene	BDL	UG/L	10	1	8270	09/27/97
	Phonecetin	BDL	UG/L	15,	1	\$270	09/27/97
	Dimethoate	BDL	UG/L	10	1	8270	09/27/97
	Paraldehydc	BDL	UG/L	15	1	8270	09/27/97
	Mathyl methanesulfonate	BDL	ua/L	10	1	8270	09/27/97
	N-Nitrosomorpholine	BDL	UG/L	15	1	8270	09/27/97
	3,3'-Dimethylbengidine	BDL	UG/L	20 .	1	8270	09/27/97
	Cyclophosphamide	BDL	ug/L	10	1	8270	09/27/97
	1-Chloronaphthalene	BDL	DG/L	15	1	8270	09/27/97
	3-Methylcholanthrone	BDL	UG/L	10	1	8270	08/27/97
	Dibenz(a,j)acridine	BDL	UG/L	20	1	8270	09/27/97
	Diphenylamine	BDL	UG/L	15	1	8270	09/27/97
	i-Methylnaphthalone	BDL	UG/L	10	1	8270	09/27/97



Attention: Scott Boling

Brown & Root Environmental \$500 South Douglas Blvd. Oaklahoma City, OK 73150

Date of Report: 10/21/97

Date Collegted: 09/04/97 Date Received: 09/05/97

Collected By: AJS

Cust. Proj: Tinker AFB IMW&B#1

Sample Desc:

CG39 2-356B

Lab ID:

867403

				Det.	Diluta		Test
		Result	Unit	Limit	Factor	Proceduro	Date

ORGAN	e						
Semi	-volatiles - GC/MS						
	1,3-Dinitrobenzene	BDL	Ug/L	10	1	8270	09/27/97
	Pamphur	BDL	UG/L	20	1	8270	09/27/97
	Isodrin	BDL	UG/L	10	1	8270	09/27/97
	4-Nitroquincline 1-oxide	BDL	UG/L	100	1	\$270	09/27/97
	0,0,0-Triathyl phosphorothicata	EDL	UG/L	9 0	1	8270	09/27/97
	2-Chlorophenol	BDL	ug/L	25	1	8270	09/27/97
	2,4-Dichlorophonol	BDL	UG/L	30	1	8270	09/27/97
	2,4-Dimethylphenol	BDL	UG/L	20	1	8270	09/27/97
	4,6-Dinitro-2-methylphonol	BDL	ug/L	80	1	8270	09/27/97
	2,4-Dinitrophenol	BDL	UG/L	96	1	8270	09/27/97
	2-Nitrophenol	BDL	UG/L	30	1	\$270	09/27/97
	4-Nitrophenol	BDL	UG/L	35	1	8270	09/27/97
	4-Chloro-3-methylphenol	BDL	UG/L	20	1	8270	09/27/97
	Pentachlorophenol	BDL	DG/L	25	1	8270	09/27/97
	Phenol	BDL	ng/r	10	1	5270	09/27/97
	2,4,6-Trichlorophenol	BDL	UG/L	55	1	8270	09/27/97
	3-Mathylphenel	DDL	UG/L	20	1	9370	09/27/97
	1-Methylphonol	BDL	DG/L	120	1	8270	09/27/97
	4-Methylphenol	BDL	UG/L	120	1	8270	09/27/97
	Benzoic acid	BDL	UG/L	45	1	8270	09/27/97
	2,4,5-Trichlorophenol	BDL	UG/L	65	1	\$270	09/27/97
	2,3,4,6-Tetrachlorophenol	BDL	ng/L	70	1	8270	09/27/97
	2,6-Dichlorophenol	BDL	UG/L	65	1	8270	09/27/97

Note 'BDL' = Below Detection Limit

APPENDIX D COST CALCULATIONS

LTM Well Installation Cost Estimate Site FTA-2 RNA TS Tinker AFB, Oklahoma 729691.3405

Design/Construction of 8 LTM/POC Wells		Description	Unit	Qty.	Unit Price	Subtotal	Total
Cost Calculations Number of wells: Depth (ft) each:	8 60		lump sum feet drum lump sum	1 480 20 1	\$500 \$40 \$45 \$700	\$19,200 \$900	

Billing Category Cost Code/(Billing Category)	Billing Rate	Task 1 (hrs)	Install New LTM Wells (\$)	Task 2 (hrs)	Procurement (\$)	Task 3 (hrs)	Reporting & PM (\$)
Word Processor 88/(15)	\$30	0	\$0	6	\$180	8	\$240
CADD Operator 58/(25)	\$47	0	\$0	2	\$94	10	\$470
Technician 42/(50)	\$40	70	\$2,800	0	\$0	18	\$720
Staff Level 16/(65)	\$57	70	\$3,990	10	\$570	30	\$1,710
Project Level 12/(70)	\$65	8	\$520	6	\$390	8	\$520
Senior Level 10/(80)	\$85	2	\$170	2	\$170	4	\$340
Principal 02/(85)	\$97	0	\$0	0	\$0	0	\$0
Total Labor (hrs \$)		150	\$7,480	26	\$1,404	78	\$4,000

ODCs .	Task 1	Task 2	Task 3
Phone	\$50	\$20	\$10
Photocopy	\$10	\$0	\$20
Mail	\$0	\$10	\$20
Computer	\$0	\$50	\$100
CAD	\$0	\$0	\$100
WP	\$0	\$20	\$100
ravel	\$1,600	\$0	\$0
Per Diem	\$1,050	\$0	\$0
Eqpt. & Supplies	\$200	\$0	\$0
Total ODCs	\$2,910	\$100	\$350

Outside Services	Task 1	Task 2	Task 3
	\$20,600		
LTM Well Installation Costs Surveying	\$2,000		
Other: Maintain Institutional Controls	\$0	\$0	\$0
Total Outside Services	\$22,600	\$0	\$0

Proposal Estimate	Task 1	Task 2	Task 3
Labor	\$7,480	\$1,404	\$4,000
ODC's	\$2,910	\$100	\$350
Outside Services	\$22,600	\$0	\$0
Total by Task	\$32,990	\$1,504	\$4,350

Total Labor	\$12,884
Total ODCs	\$3,360
Total Outside Services	\$22,600
Total Project	\$38,844

Task 1: Field Work (Installation of LTM Wells)

Task 2: Subcontracting/Permitting

Task 3: Reporting/PM per Event.

					Jo	b Number	Page 1 of 2
	PARSONS	Calculati	ion Page		729	9691.34050	
Rev	Date B		Ck	Subject:	Cost Calculat	tions for Alternative 1	
	3/24/99	MLP			RNA + LTM		
					FTA-2, Tinke	r AFB	
	l Costs						***
Design	/Construct 8 LTM/	POC Wel	ls in 1999			Present Worth Cost	\$38,844
Groun	dwater Sampling	- Years	1999 - 2033				
	30 Long-Term Mo	onitoring \	Vells				
	2 Surface-Water	r Stations					
	10 QA/QC (3 dup	l, 1 field b	olank, 2 trip b	olank, 2 MS, 2	MSD)		
	42 Total Samples	3					
Sampli	ng Labor		180	hours x	\$60	/hour	\$10,800
Analytic	cal Subcontractor						
•		42	VOCsa/		\$160	/each	\$6,720
		42	Methane/Etl	nane/Ethene	\$101	/each	\$4,242
		42	Nitrate/Nitrit	е	\$20	/each	\$840
		32	Field Param	eters	\$20	/each	\$640
Supplie	es				\$700	lump sum	\$700
Travel					\$800	lump sum	\$800
Per Die	em		24	days x	\$88	/day	\$2,112
Data M	lanagement		40	hours x	\$60	/hr	\$2,400
	alidation		20	hours x	\$60	/hr	\$1,200
Reporti	ing/Project Manag	ement La	bor				
	Word Process	ing	10	hours x	\$25	/hour	\$250
	CADD		15	hours x	•	/hour	\$750
	Reproduction		8	hours x	\$20	/hour	\$160
	Staff Level		50	hours x		/hour	\$3,000
	Proj. Manager	•	20	hours x		/hour	\$1,600
	Editor		5	hours x	•	/hour	\$300
Reporti	ing/Project Manag	ement OI	OCs		\$400	lump sum	\$400
•					Total	for 1 Sampling Event	\$36,914

a/ VOC analysis includes BTEX, CAHs, napgthalene, and MTBE.

					Jo	b Number	Page 2 of 2
PA	RSONS	Calculation F	Page		729	691.34050	
Rev I	Date	By Ck	Subje	ct:	Cost Calculati	ions for Alternative 1	
	3/3/99	•			RNA + LTM		
					FTA-2, Tinker	AFB	
Summary	of Capita	l and Present	Worth Cost	<u>ş</u>			
Capital Co	sts						
Design/Cor	nstruct 8 LTN	M/POC Wells in	1999 P/F i	=7% n=1		Present Worth Cos	\$38,84
Monitoring							
Annual	_		d 2 surface w	ater stat)13 (15 events)	
		Cost per Event			\$36,914		
		P/A i=7%, n=1.0		1999			
		P/A i=7%, n=2.0			\$32,242.12		
		P/A i=7%, n=3.6		2001			
		P/A i=7%, n=4.0		2002	·		
		P/A i=7%, n=5.0	ס		\$26,319.17		
		P/A i=7%, n=6			\$24,597.36		
		P/A i=7%, n=7			\$22,988.18		
		P/A i=7%, n=8			\$21,484.28		
		P/A i=7%, n=9		2007	•		
		P/A i=7%, n=10		2008	•		
		P/A i=7%, n=11			\$17,537.58		
		P/A i=7%, n=12			\$16,390.26		
		P/A i=7%, n=13			\$15,318.00		
		P/A i=7%, n=14			\$14,315.89		
		P/A i=7%, n=15			\$13,379.33		
			Tot	al Presei	nt Worth Cost	\$336,210	
Biennia	l Monitorin	_				2033 (10 events)	
		P/A i=7%, n=17			\$11,686.03		
		P/A i=7%, n=19			\$10,207.03		
		P/A i=7%, n=21		2019			
		P/A i=7%, n=23		2021	\$7,786.89		
		P/A i=7%, n=25		2023	\$6,801.37		
		P/A i=7%, n=27		2025	\$5,940.58		
		P/A i=7%, n=29		2027	\$5,188.74		
		P/A i=7%, n=31		2029	\$4,532.04		
		P/A i=7%, n=33		2031	\$3,958.46		
		P/A i=7%, n=35		2033	\$3,457.47 nt Worth Cost	\$68,474	
			100	ai Fiesei	it vvoitii cost	Ψ00,474	
Site Ma	nagement (every year (35 y	/ears)				
		Annual Cost		\$6,000			
		P/A i=7% n=35			12.9476723		• *
				Prese	nt Worth Cost	\$77,686	
		sent Worth Co			***	\$521,213	7

P	ARSONS		Job Number		Page 1 of 2	
C P	Calculation Page		729691.34050			
lev	Date By Ck		lations for Alternati			
	3/3/99 MLP		M, Soil Gas Source		ion/	
			ization, and Source	Removal		
		FTA-2, Ti	nker AFB			
Capital C						
Source Ar	ea Identification/ Characterization					
	Survey - Initial Investigation					
- 50 ft no	ode centered grid for an area of $920,000 \text{ft}^2 = 412 \text{ san}$					
	Geoprobe Rental (8 pts. Per day)	52 days x	\$1,000 /d	-	\$52,00	
	Soil Gas Analysis (EPA TO-3)	412 samples x	\$130 /sa	ample	\$53,56	
	Sampling Labor	832 hours x	\$60 /h	our	\$49,92	
	Sample Shipping		\$500 lui	mp sum	\$50	
	Per Diem	104 days x	\$88 /d	-	\$9,15	
		104 days x	\$200 lui	•	\$20	
	Travel			-		
	Equipment Rental (Meters)		\$2,000 lui	-	\$2,00	
	Progress Report	40 hours x	\$70 /h	our	\$2,80	
	Contingency (10%)				\$11,81	
		Subtotal for I	initial Soil Gas Inve	stigation:	: \$129,94	
Soil Gas S	urvey - Hot Spot Investigation Investigation					
- 20 ft no	ode centered grid for 3 areas of 25,600ft ² = 243 samp	oling points.				
20 11 11	Geoprobe Rental (8 pts. Per day)	31 days x	\$1,000 /d	av	\$31,00	
	Soil Gas Analysis (EPA TO-3)	243 samples x	\$130 /sa		\$31,59	
	•	496 hours x	\$60 /h	_	\$29,76	
	Sampling Labor	490 Hours X	\$500 lui		\$50	
	Sample Shipping	(2) 1	\$88 /d		\$5,45	
	Per Diem	62 days x		-		
	Travel		\$200 lui	-	\$20	
	Equipment Rental (Meters)		\$2,000 lui	-	\$2,00	
	Progress Report	20 hours x	\$70 /h	our	\$1,40	
	Contingency (10%)				\$7,09	
		Subtotal for Hot	Spot Soil Gas Inve	stigation:	\$77,99	
Geophysic	al Survey					
	Time and Materials	5 days x	\$1,600 /da	ay	\$8,00	
	Mod/Demob		\$3,500 lui	-	\$3,50	
	Data Analysis/Reporting		\$2,500 lui	•	\$2,50	
			Ψ2,500 ια	np sam	\$1,40	
	Contingency (10%)	Subt	otal for Geophysica	al Survey:		
					, ,	
Slug Tests		6 wells x	\$300 /w	ell	\$1,80	
	Time, Materials, Labor, Data Analysis, Reporting	o wells X	\$300 /W	CII		
	Contingency (10%)		Subtotal for S	lug Tests:	\$18 \$1,98	
			Danielli IVI D	www.	41,50	
racer Tes		211	£2 000 /	.a11	¢ ດ ດ	
	Install 3 Monitoring Well	3 well x	\$3,000 /w		\$9,0	
	Tracer (3 kg reagent grade Lithium bromide)		\$200 lui	-	\$20	
			\$2,000 lur	np sum	\$2,00	
	Job Supplies			-		
	Job Supplies Data Collection (Total present worth cost of data co	ellection for life of test)	\$36,741 lui	-	\$36,74	
	Data Collection (Total present worth cost of data co	ellection for life of test)	\$36,741 lur	mp sum*		
		ollection for life of test)		mp sum*	\$36,74 \$5,00 \$5,29	

MANA A PHILI	Ul VIII THE	THE PARTY AND		101		or warm a rogram -	Ψυ~,,L1
I.TM Plan	(from Alter	rnativo 1)		Tot	al Present V	Vorth of LTM Program =	
SOII EXCAVA	mon:					Total Capital Costs =	
Tracer Test Soil Excava							\$38,23 \$253,08
Slug Tests							\$1,98 \$58,23
Geophysica							\$15,40
•	oil Gas Invest	tigation:					\$77,99
	Gas Investiga						\$129,94
	of Capital C					<u>*</u>	
						btotal for Soil Excavation:	
	g Oversight	10 02 0001	,	120		\$50.00 /Hour	\$6,000.
	ce (RSMeans			1667		\$6.69 /yd ²	\$11,150.
_			h Sand (RSMeans 17 03 040	8,325.0		\$8.45 /yd ³	\$70,346.
	_		Arapahoe Disposal)	8,325.0	yd ³	\$15.00 /yd ³	\$124,875.
	Transport (R				20 yd³/mile²/	\$3.44 /20 yd ³ /mile	\$28,638.
Soil Excava	ation (RSMea	ns 17 03 02	.78)	8,325.0	yd ³	\$1.45 /yd ³	\$12,071.
ITEM (W	ith Costing Re	eference)		# Units	Unit	Unit Cost ^{a/}	Total
			Total Weight of Soil to be	e Disposed of:	10,406.3	tons	
			(Assume Contaminant thick	ness of 15 ft):	8,325	•	
			Total Volume of Soil to be	e Disposed of			
Tota	al Weight of	Soil to be E	xcavated (Weight of soil = 1	1.25 tons/yd^3):	10,406.3	tons	
					8,325	yd ³	
	Total Vol	ume of Soil	to be Excavated (to an assur	ned 15 ft bgs):	225,000		
			1012	TATES OF SILE.	1,666.7		
Source Are Excavation	ea Removal		Total	l Area of Site:	15,000	\mathbf{ft}^2	
					FTA-2, Tink	er AFB	
						ation, and Source Removal	
	3/3/99	MLP			, ,	Soil Gas Source Identificat	ion/
Rev	Date	Ву	Ck	Subject:	Cost Calculat	tions for Alternative 2	
	RSONS	Calcula	tion Page	1		729691.34050	1

^{*} The lump sum amount presented here is the total present worth cost for tracer sampling, as presented on the cost table 2-1 (Tracer Sampling).

Remedial Alternative 3 Cost Estimate Site FTA-2 RNA TS Tinker AFB, Oklahoma 729691.3405

Author: MLP Date: 3/24/99 Checked by: Date:

nstallation of	the Groundwater Extraction System	040,000 luman aum	¢10.00
	Design/perform/analyze 24-hour pumping test (lump sur	m) \$10,000 lump sum	\$10,00
	Well Installation (7 4-inch wells) 7 wells x	\$4,000 /well	\$28,00
	Building/Controls/Piping/Accessories	\$20,000 lump sum	\$20,00
	Equipment Costs 7 submersible pumps, air stripper, electronics, PLC valves, and guages	\$600,000 lump sum	\$600,00
	System Installation	\$50,000 lump sum	\$50,00
		subtotal	\$708,00
nnual Opera	iting CostsPump and Treat System (10 years)		
	Labor		
	Weekly System Checks (2 hr x 52 wk/yr x \$60/hr)		\$6,24
	Monthly Sampling (12 hr x 12 mo x \$60/hr)		\$8,64
	Monthly Reporting (8 hr x 12 mo x \$60/hr)		\$5,76
	Maintenance (8 hr x 12 mo x \$60/hr)		\$5,76
	Analytical VOCs (air) (2/mo x 12 mo x \$250/ea)		\$6,00
	VOCs (air) (2/mo x 12 mo x \$250/ea) Sample ship. (2/mo x 12 mo x \$30/ea)		\$72
	• • •		Ψ, Σ
	Supplies Install and Dispose of GAC (assume 2 replacements)	ents of 2 000 lb. of GAC/year)	\$9,40
	misc. (lump sum)	me of Epoca in the control of the co	\$1,00
	Travel (52 trips from Midwest City OK per year @ \$100/	ea)	\$5,20
	Electricity	,	\$5,00
	Site Management (Included in LTM portion of Alternative	e 1)	\$

Remedial Alternative 3 Cost Estimate Site FTA-2 RNA TS Tinker AFB, Oklahoma 729691.3405

Author: MLP Date: 3/24/99 Checked by: Date:

Summary of Present Worth Capital Costs (assuming system installation in 1999)

Installation of Groundwater Extraction System

\$708,000 P/A i=7%, n=0

\$708,000

Total

\$708,000

Summary of Present Worth Annual + Future Costs (assuming 10-yr operating period 1999-2008)

O&M for Pump and Treat System (10 years)

Annual Cost	\$55,520
1999 n=1	\$51,887.85
2000 n=2	\$48,493.32
2001 n=3	\$45,320.86
2002 n=4	\$42,355.94
2003 n=5	\$39,584.99
2004 n=6	\$36,995.32
2005 n=7	\$34,575.07
2006 n=8	\$32,313.15
2007 n=9	\$30,199.20
2008 n=10	\$28,223.55

Total Present Worth Cost (O&M)

\$389,949

Total Present Worth Cost of LTM Program/ Site Characterization/ Source Removal - Alternative 1 and 2 (includes site management)

\$1,057,850

Total Capital + Annual + Future Costs for Alternative 3

\$2,155,800

APPENDIX E

SUPPORTING MODEL INPUT AND CALCULATIONS FOR GROUNDWATER/FATE AND TRANSPORT MODELING

TABLE E.1 SUMMARY OF CHLORINATED SOLVENT DECAY RATES AUGUST 1997 FTA-2 RNA TS TINKER AFB, OKLAHOMA

Contaminant	Sample	Flow Path	Analysis	Decay Rate	HalfLife
	Date		Method	(day ⁻¹)	(years)
TCE	August-97	2-62B > CG39B97-43S > 2-355B	Buscheck and Alcantara'	3.06E-04	6.21
cis-1,2-DCE	August-97	2-62B > CG39B97-43S > 2-355B	Buscheck and Alcantar	4.08E-04	4.65
trans-1,2-DCE	August-97	2-62B > CG39B97-43S	Buscheck and Alcantar	7.58E-04	2.51
Total Chlorinated Ethenes	August-97	2-62B > CG39B97-43S > 2-355B	Buscheck and Alcantar	2.96E-04	6.42
Total Chlorinated Ethenes	August-97	2-62B > CG39B97-43S > 2-355B	Reductive Dechlorination	1.88E-07	10,091
1,2-DCA	August-97	2-62B > CG39B97-43S > 2-355B	Buscheck and Alcantar	1.53E-03	1.24
Total Chlorinated Ethanes	August-97	2-62B > CG39B97-43S > 2-355B	Buscheck and Alcantar	1.28E-03	1.49
Total Chlorinated Ethanes	August-97	2-62B > CG39B97-43S > 2-355B	Reductive Dechlorination	5.23E-07	3,630
Chlorobenzene	August-97	2-62B > 2-65B > CG39B97-43S	Buscheck and Alcantar	9.28E-04	2.05
1,2-Dichlorobenzene	August-97	2-62B > 2-65B	Buscheck and Alcantar	3.36E-03	0.57
1,3-Dichlorobenzene	August-97	2-62B > CG39B97-43S	Buscheck and Alcantar	2.97E-04	6.39
1,4-Dichlorobenzene	August-97	2-62B > 2-65B	Buscheck and Alcantar	2.33E-03	0.82
Total Chlorinated Benzenes	August-97	2-62B > 2-65B > CG39B97-43S	Buscheck and Alcantar	1.05E-03	1.81
Total Chlorinated Benzenes	August-97	2-62B > 2-65B	Reductive Dechlorination	3.41E-04	5.56

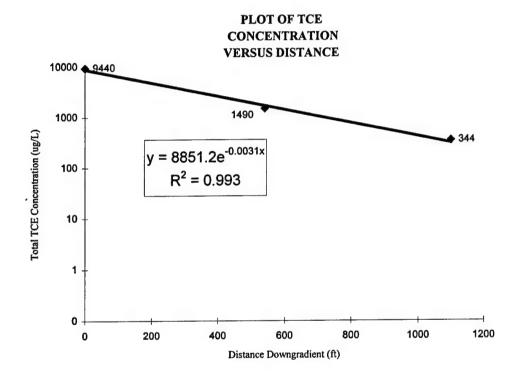
Method of Buscheck and Alcantar (1995) for steady-state plumes.

^{b/} Reductive Dechlorination by method of Moutoux, et al., (1996), which gives decay rate attributable to reductive dechlorination.

FIRST-ORDER DECAY RATE CALCULATION USING THE METHOD OF BUSCHECK AND ALCANTAR (1995) SITE FTA-2 RNA TS

TINKER AFB, OKLAHOMA

	Distance (ft)	TCE (µg/L)
Point	Downgradient	Aug-97
2-62B	0	9440
CG39B9743s	540	1490
2-355B	1100	344



$$\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$$
 where $v_c = 0.066$ ft/day $\alpha_x = 160$ feet $k/v = 0.0031$ therefore $\lambda = 3.06E-04$ days⁻¹ half life = 6.20 years

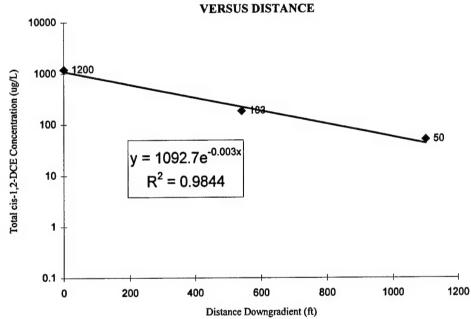
FIRST-ORDER DECAY RATE CALCULATION USING THE METHOD OF BUSCHECK AND ALCANTAR (1995)

SITE FTA-2 RNA TS

TINKER AFB, OKLAHOMA

Point	Distance (ft) Downgradient	cis-1,2-DCE (μg/L) Aug-97
2-62B	0	1200
CG39B9743s	540	183
2-355B	1100	50

PLOT OF cis-1,2-DCE CONCENTRATION VERSUS DISTANCE

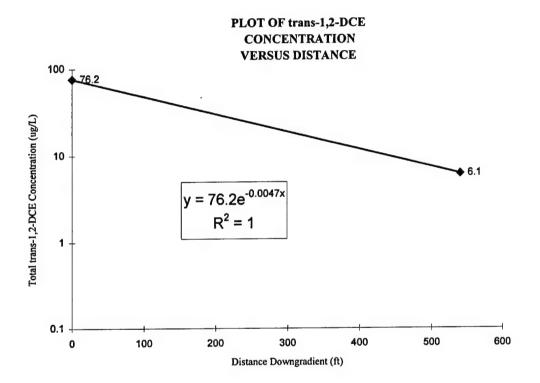


$$\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$$

where $v_c =$	0.092	ft/day
$\alpha_x =$	160	feet
k/v =	0.003	
therefore $\lambda =$	4.08E-04	days ⁻¹
half life =	4.65	years

SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA

	Distance (ft)	trans-1,2-DCE (µg/L)
Point	Downgradient	Aug-97
2-62B	0	76.2
CG39B9743s	540	6.1



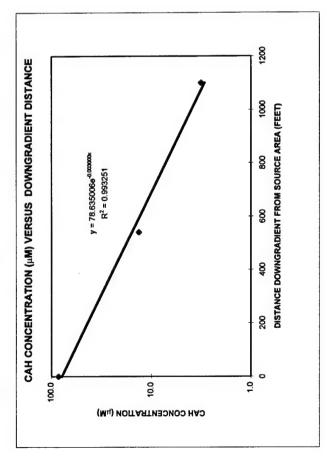
where
$$v_c = 0.092$$
 ft/day $\alpha_x = 160$ feet $k/v = 0.0047$ therefore $\lambda = 7.58\text{E}-04$ days⁻¹ half life = 2.51 years

 $\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$

ESTIMATED FIRST-ORDER RATE CONSTANT CALCULATION FOR TOTAL CHLORINATED ETHENES USING THE METHOD OF BUSCHECK AND ALCANTAR (1995) AUGUST, 1997

TINKER AFB, OKLAHOMA SITE FTA 2 RNA TS

	Distance	PCE (µg/L)	PCE (µM)	TCE (µg/L)	TCE (µM) 1	Total DCE (µg/L	Total DCE (µM)	VC (µg/L)	VC (µM)	ľ	.) Total CAH (µM)
Well	Downgradient	Aug-97	Aug-97	Aug-97	Aug-97	Aug-97	Aug-97 Aug	Aug-97	Aug-97	Aug-97	Aug-97
2-62B	0	6.1	0.04	9440	71.9		13.3	1.7	0.0	10733.5	85.2
CG39B9743s	540	0.0	00.0	1490	11.3	190.2	2.0	0.0	0.0	1680.2	13.3
2-355B	1100	0.0	00.0	344	5.6		0.5	0.0	0.0	394.0	3.1



1200 CAH CONCENTRATION (µG/L) VERSUS DOWNGRADIENT DISTANCE DOWNGRADIENTFROM SOURCE AREA (FEET) $y = 9920.472476e^{-0.000002x}$ $R^2 = 0.993456$ 800 DISTANCE 8 8 1.0 10000 100.0 10.0 1000001 100001 САН СОИСЕИТВАТІОИ (µG/L)

Method of Buscheck & Alcantar (1995) $\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$ ft/day 990.0 160 k/v = **^**° ≡ where

 $\lambda = 2.9304E-02 \%day$ therefore $\lambda = 2.9304E-04$ 1/day half life =

years

6.48

1/day %/day ft/day ft Method of Buscheck & Alcantar (1995) $\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$ 2.9563E-04 2.9563E-02 160 990.0 k/v = ر د اا $\alpha_x =$ therefore $\lambda =$ where

years

6.42

half life =

REDUCTIVE DECHLORINATION RATE FOR CHLORINATED ETHENES 2-62B > CG39B97-43S > 2-355B

SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA

The following procedures outlined by Moutoux et al. (1996)

Step 1: Enter Observed Contaminant Concentration (mg/L)

Well	PCE	TCE	DCE	VC	Ethene	Total CAHs
2-62B	6.1	9440.0	1285.7	1.7	0	10733.5
CG39B97-43S	0.0	1490.0	190.2	0.0	0	1680.2
2-355B	0.0	344.0	50.0	0.0	0	394.0

Step 2: Compute Molar Concentrations (micromoles/Liter)

Well	M _{PCE}	M_{TCE}	M _{DCE}	M _{VC}	M _{Ethene}	Total M _{CAHs}
2-62B	0.04	71.85	13.26	0.03	0.00	85.18
CG39B97-43S	0.00	11.34	1.96	0.00	0.00	13.30
2-355B	0.00	2.62	0.52	0.00	0.00	3.13

Step 3: Compute Carbon Equivalents

Well	Total M _{CAHs}	x 2	= Ceqi	
2-62B	85.18		170.36	
CG39B97-43S	13.30		26.61	
2-355B	3.13		6.27	

Step 4: Compute Chlorine Equivalents

Well	M _{PCE} x 4	M _{TCE} x 3	M _{DCE} x 2	M_{VC}	$S = Cleq_i$
2-62B	0.15	215.56	26.53	0.03	242.26
CG39B97-43S	0.00	34.02	3.92	0.00	37.95
2-355B	0.00	7.86	1.03	0.00	8.89

Step 5: Compute Corrected CAH Concentrations

 $C_{i,corr} = C_{i-1,corr} \times (Cleq_i / Cleq_{i-1}) \times (Ceq_{i-1} / Ceq_i)$

Well	C _{i-1,corr}	Cleq _i / Cleq _{i-1}	Ceq _{i-1} / Ceq _i	$C_{i,corr}$
2-62B	10733.500300			10733.500300
CG39B97-43S	10733.500300	0.156641	6.402853	10765.153575
2-355B	10765.153575	0.234182	4.244514	10700.462928

Step 6: Plot Exponential Trendline of Contaminant Concentration vs. Time

Well	Distance From Source (ft)	CAH avg. velocity (ft/day)*	CAH Travel Time (day)	$C_i,corr$
2-62B	0	0.066	0	10733.500300
CG39B97-43S	540	0.066	8182	10765.153575
2-355B	1100	0.066	16667	10700.462928

^{*}Assumes an unretarded velocity of 0.33 feet/day

and a retardation coefficient of 5.0.

Reductive Dechlorination Rate

 $C = C_0 e^{-kt}$

where:

C = Corrected Contaminant Concentration (mg/L) at time t (days)

C₀ = Initial Contaminant Concentration (mg/L)

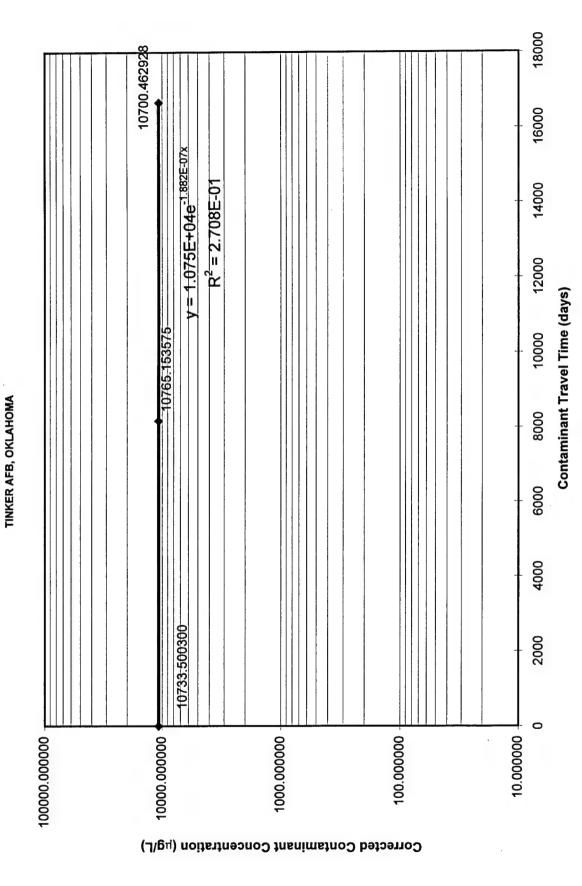
k = Reductive Dechlorination Rate (days⁻¹)

from plot:

 $y = 10,750e^{-1.882E-07x}$

k = 1.882 E-07 days⁻¹

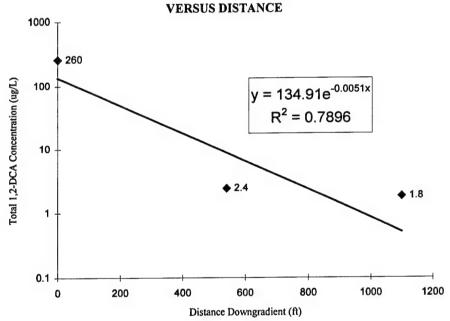
REDUCTIVE DECHLORINATION RATE FOR CHLORINATED ETHENES 2-62B > CG39B97-43S > 2-355B AUGUST, 1997 SITE FTA-2 RNA TS



SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA

	Distance (ft)	1,2-DCA (μg/L)
Point	Downgradient	Aug-97
2-62B	0	260
CG39B9743s	540	2.4
2-355B	1100	1.8

PLOT OF 1,2-DCA CONCENTRATION VERSUS DISTANCE



$$\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$$

where $v_c =$	0.165	ft/day
$\alpha_x =$	160	feet
k/v =	0.0051	
therefore $\lambda =$	1.53E-03	days ⁻¹
half life =	1.24	vears

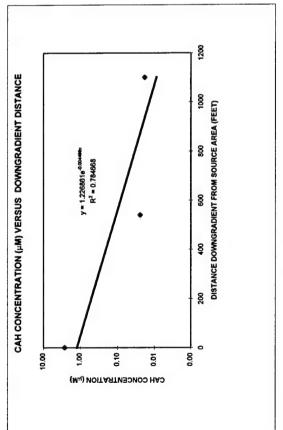
ESTIMATED FIRST-ORDER RATE CONSTANT CALCULATION FOR TOTAL CHLORINATED ETHANES USING THE METHOD OF BUSCHECK AND ALCANTAR (1995)

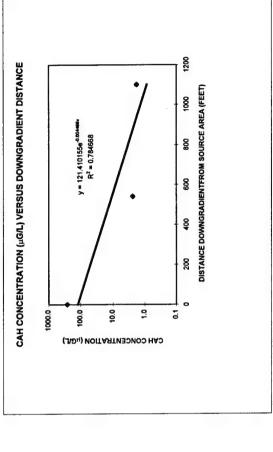
SITE FTA2 RNA TS TINKER AFB, OKLAHOMA

Total CAH (µM) Aug-97 2.6 0.02 0.02 Total CAH (μg/L) Aug-97 260.0 2.4 1.8 1,2-DCA (µM) 1,2-DCA (µM) Aug-97 Aug-97 Aug-97 0.00 260.0 2.4 0.02 0.00 1.8 0.02 1,2-DCA (μg/L) Aug-97 0.00 0.00 0.00 1,1,1-TCA (μg/L) 1,1,1-TCA (μM) Aug-97 Aug-97 0.00 0.00 0.00 0.00 . 2-62B CG39B9743S TMP-3

540 1100

Well





Method of Buscheck & Alcantar (1995) $\lambda = v_c 4\alpha_x ([1+2\alpha_x(k/v_c)]^2 - 1)$	ft/day ft	1/day %/day	years
of Buscheck & Alcantar (199 $\lambda = v_{\nu}/4\alpha_{\kappa}([1+2\alpha_{\kappa}(k/v_{\kappa})]^{2}-1)$	0.165 160 0.004496	1.2755E-03 1/day 1.2755E-01 %/day	1.49
Method of B ₁ $\lambda = v$	where $v_c = \alpha_x = K/v = K$	therefore $\lambda = \lambda$	half life =

where $v_e = 0.165$ fl/ $(1+2\alpha_e(k/v_a))^{2-1}$ where $v_e = 0.165$ fl/ $\alpha_a = 160$ fl $k/v = 0.00496$ therefore $\lambda = 1.2755E-01$ $9/\omega$
--

REDUCTIVE DECHLORINATION RATE FOR CHLORINATED ETHANES 2-62B > CG39B97-435 > 2-355B SITE FTA-2 NA TA THISTER AFF, ORCAHOMA

The following procedures	outlined by Mo	utoux et al. (1996)
--------------------------	----------------	---------------------

Step 1:		taminant Concentration (µg/L)			CA	Total CAHs
	Well	1,1,1-TCA	1,1-DCA	1,2-DCA	0.0	260.0
	2-62B	0.0	0.0	260.0	0.0	2.4
	CG39B97-43S	0.0	0.0	2.4	0.0	1.8
	2-355B	0.0	0.0	1.8	0.0	1.0
tep 2:	Compute Molar Con	centrations (micromoles/Liter)				
	Well	M _{1,1,1-7CA}	M _{1 1-DCA}	M _{1,2-DCA}	Ma	Total Mcu-
	2-62B	0.0000	0.0000	2.6273	0.0000	2.6273
	CG39B97-43S	0.0000	0.0000	0.0243	0.0000	0.0243
	2-355B	0.0000	0.0000	0.0182	0.0000	0.0182
tep 3:	Compute Carbon Eq	julvalents				
	Well	Total Mouse	x2	= Ceqì		
	2-62B	2.6273241714		5.2546483428		
	CG39B97-43S	0.0242522231		0.0485044482		
	2-355B	0.0181891673		0.0383783347		
itep 4:	Compute Chlorine E	quivalents				
	Well	M _{1.1.1-TCA} x3	M _{1,1-0CA} x 2	M _{1,2-DCA} × 2	Moa	$\Sigma = Cleq_i$
	2-62B	0.000000000	0.0000000000	5.2546483428	0.000000000	5.2548483428
	CG39B97-43S	0.000000000	0.000000000	0.0485044482	0.000000000	0.0485044462
	2-355B	0.000000000	0.000000000	0.0363763347	0.000000000	0.0363783347
Step 5:	Compute Corrected	CAH Concentrations				
	•	C _{Loor} =	C _{i-1,corr} x (Cleq, / Cleq, 1) x (Ceq,	. ₁ / Ceq.)		
	Well	C _{b1 corr}	Cleq. / Cleq. 1	Ceq., / Ceq.	CLOOP	
	2-62B	260.0000000000			260.0000000000	
	CG39B97-43S	260.0000000000	0.0092307692	108.333333333	260.0000000000	
	2-355B	260.0000000000	0.7500000000	1,333333333	260.0000000000	
tep 5:	Plot Exponential Tre	endline of Contaminant Concen	tration vs. Time			
•		Distance	CAH	CAH		
		From	avg. velocity	Travel		
	Well	Source (ft)	(ft/day)*	Time (day)	C _{Looff}	
	2-62B	0.000000000	0.185000000	0.000000000	260.0000000000	
	CG39B97-43S	540,0000000000	0.1650000000	3272.7272727273	260.0000000000	
	2-355B	1100.0000000000	0.1850000000	6666.6666666667	260.0000000000	
		(3 feet/day and an average retardation coef	Micient of 2.0			

Distance From Source (ft) 0.0000000000 540.0000000000 1100.0000000000 2-62B CG39B97-43S 2-355B mes an unretarded velocity of 0.33 feet/day and an average retards

Reductive Dechlorination Rate $C = C_0 e^{4\epsilon}$

C = Corrected Contaminant Concentration ($\mu Q/L$) at time t (days) C₀ = Initial Contaminant Concentration ($\mu Q/L$)

k = Reductive Dechlorination Rate (days⁻¹)

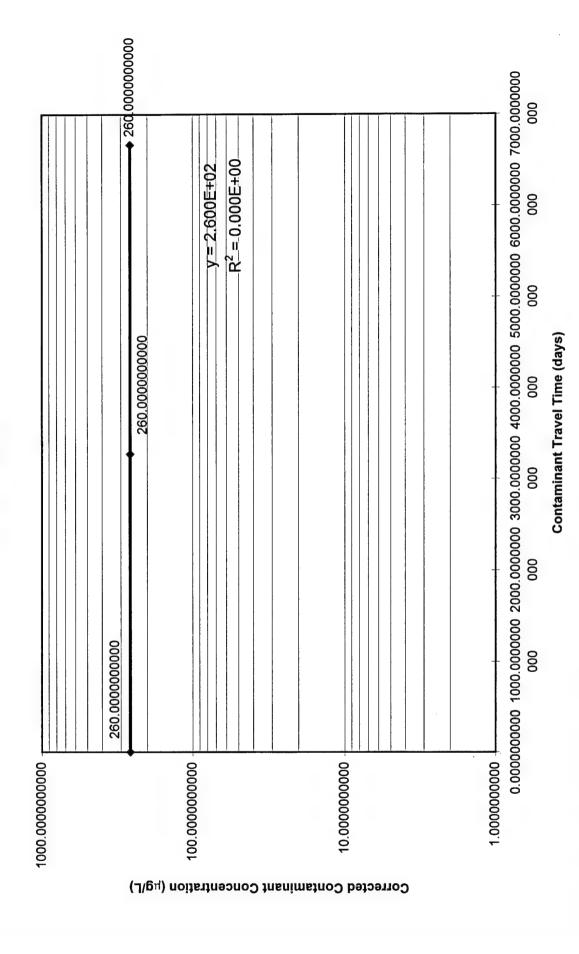
from plot:

y = 2673e^{-5.7324E-07x}

where:

k = 5.232E-07 days-1

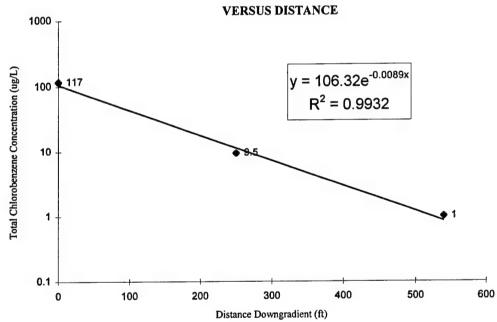
REDUCTIVE DECHLORINATION RATE FOR CHLORINATED ETHANES 2-62B > 2-65B SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA



SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA

	Distance (ft)	Chlorobenzene (µg/L)
Point	Downgradient	Aug-97
2-62B	0	117
2-65B	250	9.5
CG39B9743s	540	1

PLOT OF CHLOROBENZENE CONCENTRATION



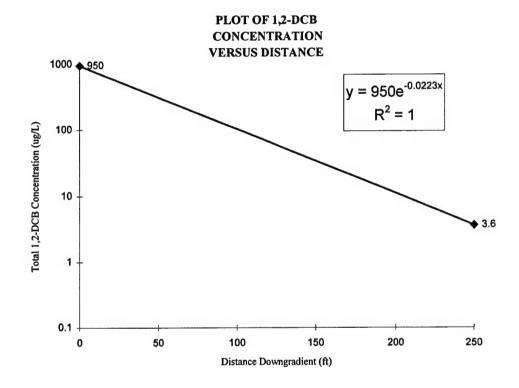
$$\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$$

where $v_c =$	0.043	ft/day
$\alpha_x =$	160	feet
k/v =	0.0089	
therefore $\lambda =$	9.28E-04	days ⁻¹
half life =	2.05	years

SITE FTA-2 RNA TS

TINKER AFB, OKLAHOMA

	Distance (ft)	1,2-DCB (μg/L)
Point	Downgradient	Aug-97
2-62B	0	950
2-65B	250	3.6

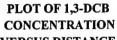


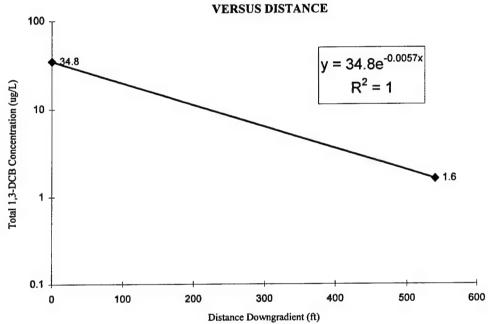
$$\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$$

where
$$v_c = 0.033$$
 ft/day $\alpha_x = 160$ feet $k/v = 0.0223$ therefore $\lambda = 3.36\text{E}{-}03$ days⁻¹ half life = 0.56 years

SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA

	Distance (ft)	1,3-DCB (μg/L)
Point	Downgradient	Aug-97
2-62B	0	34.8
CG39B9743s	540	1.6



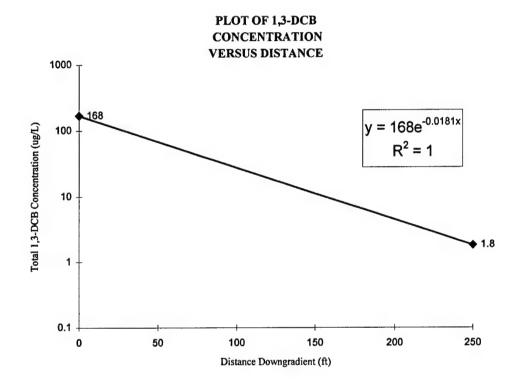


$$\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$$

where
$$v_c$$
 = 0.033 ft/day α_x = 160 feet k/v = 0.005 therefore λ = 2.97E-04 days⁻¹ half life = 6.39 years

TINKER AFB, OKLAHOMA

	Distance (ft)	1,4-DCB (μg/L)
Point	Downgradient	Aug-97
2-62B	0	168
2-65B	250	1.8



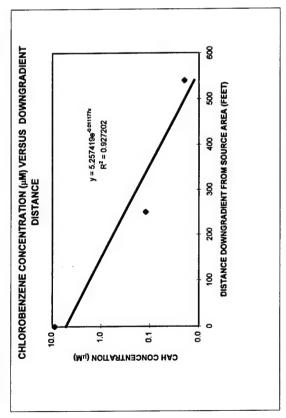
$$\lambda = v_c/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$$

where
$$v_c = 0.033$$
 ft/day $\alpha_x = 160$ feet $k/v = 0.0181$ therefore $\lambda = 2.33E-03$ days⁻¹ half life = 0.82 years

ESTIMATED FIRST-ORDER RATE CONSTANT CALCULATION FOR TOTAL CHLORINATED BENZENES USING THE METHOD OF BUSCHECK AND ALCANTAR (1995) AUGUST 1997

SITE FTA- 2 RNA TS TINKER AFB, OKLAHOMA

									ı		
	Distance	1,2-DCB (µg/L)	1,2-DCB (µM)	1,3-DCB (µg/L)	5	1,4-DCB (µg/L)	1,4-DCB(µM)	CB (µg/L)		Total C-Benz. (µg/L)	Total C-Benz. (µM)
Well	Downgradient	Aug-97	Aug-97		Aug-97	Aug-97	Aug-97	Aug-97	Aug-97	Aug-97	Aug-97
2-62B	0	950	6.5	34.8	0.2	168	1.1	117	1.0	1269.8	8.9
2-65B	250	3.6	0.0	0.0	0.0	1.8	0.0	9.5	0.1	14.9	0.1
39B97-43S	540	0.0	0.0	9.1	0.0	0.0	0.0	1.0	0.0	2.6	0.0



CHLOROBENZENE CONCENTRATION (µG/L) VERSUS DOWNGRADIENT DISTANCE	y = 722.216585.8 ^{0.0113112} R ² = 0.918588	100 200 300 400 500 600 DISTANCE DOWNGRADIENTFROM SOURCE AREA (FEET)
HLOROBENZEI DOV	100.00	1.0 to 10
Ö	(אפת) (אפת) (אפת) 현 수	

ft/day ft	1/day %/day	years
160	1.0284E-03	1.85
where $v_c = \alpha_x = \alpha_x = 1$	therefore $\lambda = \lambda$	half life =
	$v_c = 0.033$ $\alpha_x = 160$ $v_{x,t} = 0.01177$	$v_{s} = 0.033$ $\alpha_{x} = 160$ $k/v = 0.011177$ $\lambda = 1.0284E-03$

1995) -1)	ft/day ft	1/day %/day	years
Method of Buscheck & Alcantar (1995) $\lambda = v_s/4\alpha_x([1+2\alpha_x(k/v_x)]^2-1)$	0.033 160 0.011321	1.0503E-03 1.0503E-01	1.81
Method of Buscl $\lambda = v_o l$	where $v_c = \alpha_x = k/v = k/v = k$	therefore $\lambda = \lambda = \lambda$	half life =

REDUCTIVE DECHLORINATION RATE FOR CHLORINATED BENZENES 2-62B > 2-65B

SITE FTA-2 RNA TS TINKER AFB, OKLAHOMA

The following procedures outlined by Moutoux et al. (1996)

Step 1: Enter Observed Contaminant Concentration (µg/L)

Well	1,2-DCB	1,3-DCB	1,4-DCB	СВ	Total CB's
2-62B	950.0	34.8	168.0	117.0	1269.8
2-65B	3.6	0.0	1.8	9.5	14.9

Step 2: Compute Molar Concentrations (micromoles/Liter)

Well	M _{1,2-DCB}	M _{1.3-DCB}	M _{1,4-DCB}	M_{CB}	Total M _{CBs}
2-62B	6.46	0.24	1.14	1.04	8.88
2-65B	0.02	0.00	0.01	0.08	0.12

Step 3: Compute Carbon Equivalents

Well	Total M _{CBs}	x 6	= Ceqi
2-62B	8.88		53.29
2-65B	0.12		0.73

Step 4: Compute Chlorine Equivalents

	Well	$M_{1,2\text{-DCB}} \times 2$	$M_{1,3-DCB} \times 2$	M _{1,3-DCB} x 2	M_{CB}	$\Sigma = Cleq_i$
_	2-62B	12.93	0.47	2.29	1.04	16.72
	2-65B	0.05	0.00	0.02	0.08	0.16

Step 5: Compute Corrected CAH Concentrations

 $C_{i,corr} = C_{i-1,corr} \times (Cleq_i / Cleq_{i-1}) \times (Ceq_{i-1} / Ceq_i)$

	Well	$C_{i-1,corr}$	Cleq _i / Cleq _{i-1}	Ceq _{i-1} / Ceq _i	$C_{i,corr}$	
_	2-62B	1269.80			1269.80	
	2-65B	1269.80	0.01	73.32	931.02	878.87

Step 6: Plot Exponential Trendline of Contaminant Concentration vs. Time

	Distance	СВ	СВ	
	From	avg. velocity	Travel	
Well	Source (ft)	(ft/day)*	Time (day)	C _{i,corr}
2-62B	0	0.033	0	1269.80
2-65B	30	0.033	909	931.02

^{*}Assumes an unretarded velocity of 0.33 feet/day and an average retardation coefficient of 10.

Reductive Dechlorination Rate

 $C = C_0 e^{-kt}$ where:

C = Corrected Contaminant Concentration (μ g/L) at time t (days)

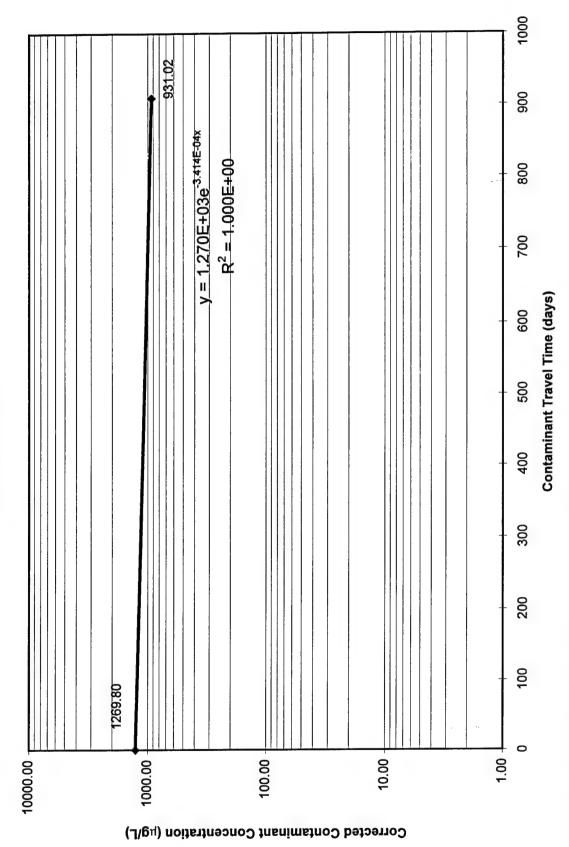
C₀ = Initial Contaminant Concentration (μg/L)

k = Reductive Dechlorination Rate (days⁻¹)

from plot: $y = 1270e^{-3.414E-04x}$ $k = 3.414E-04 days^{-1}$

REDUCTIVE DECHLORINATION RATE FOR DICHLOROBENZENES TO MONOCHLOROBENZENE

2-62B > 2-65B
AUGUST, 1997
SITE FTA-2 RNA TS
TINKER AFB, OKLAHOMA



MODEL RUN:	FTA2-17	MODFLOW Inp		
FLOW RMS:	0.657717	SOLVER:	WHS	
TCE RMS:	314.83	ADVECTION:	UPSTREAM FD	
Head Calib Inpu	t File:		Initial Head File:	FTA-SS4.HDS
TCE Calib Input	File:		-	@ 12775
GRID (see map)				
	# rows	63	# layers	3
	# columns	69	top of layer elev.	1253 ft. msl
	source area cell size	50 ft X 50 ft	bottom of layer elev.	1161 ft. msl
	total area of model	30,000,000 sq ft		
	# cells	4347	_	
CONDUCTIVITY		V	V	Onlaw
Zone #	K _x	K _y	K _z	Color
1	1.5	2.25	0.015	White
2	1.00E-02	1.00E-02	1.00E-20	Blue
3	1.5	2.25	0.15	Green
4	10	15	0.1	Teal
5	24	36	0.24	Red
6	24	36	36	Purple
7	24	36	36	Olive
8	18	27	27	Grey
STORAGE	_	_		
	S_s	S_y	n _e	n_{total}
	.00001 ft-1	0.2	0.2	0.3
	NDITIONS (see map)			
	- constant head bounds	aries up- and dowr	n-gradient	
	243' to 1234'	404446 4040		
West: L		1214 to 1213		
	ayer 2	1212 to 1211		
L	ayer 3	1210 to 1209		
		MT3D Input	Form: FTA-2	
SOURCE CONC	ENTRATION CELL (s			
	ROW	COLUMN		
Cell A	43	52		
Cells B	44	51		
	45	50		
Cell C	42	52		

CONCENTRATIO	N SCHEDULE (ug/l):	Cell A	Cells B	Cell C
0 to	1825	75,000	0	9500
1825	3650	65,000	5000	9500
3650	5475	55,000	7500	9500
5475	7300	50,000	9500	9500
7300	9125	40,000	9500	9500
9125	10950	30,000	9500	9500
10950	12775	20,000	9500	9500

ANIOSOTROPY RATIO

Tx/Ty

6-6-3

DISPERSION COEFFICIENTS	DISPERSION COEFFICI	ENTS
-------------------------	----------------------------	-------------

Longitudinal	Dispersion	(D.)	- all lavers	
Longitudina	Dispersion	(UL)	- an layers	,

- 3	\ <u>L</u>					
		Zone 1	Zone 2	Zone 3	Zone 4	Zone 5
		20	20	20	2	5
Transverse Horizont	al Dispersion (D _{TH})					
Ratio D _L /	D _{TH}	Zone 1	Zone 2	Zone 3	Zone 4	Zone 5
Layer 1	0.1	2	2	2	0.2	0.5
Layer 2	0.2	4	4	4	0.4	1
Layer 3	0.22	4.4	4.4	4.4	0.44	1.1
Transverse Vertical I	Dispersion (D_{TV})					
Ratio D _L /	'D _{TV}	Zone 1	Zone 2	Zone 3	Zone 4	Zone 5
Layer 1	0.01	0.2	0.2	0.2	0.02	0.05
Layer 2	0.02	0.4	0.4	0.4	0.04	0.1
Laver 3	0.05	1	1	1	0.1	0.25

CHEMICAL REACTION PACKAGE (see map)

Zone #	Kd (ft3/kg)	Bulk Density	Decay Rate (1/yr)	Sorbed (1/yr)
1 - white	0.004	46.73	0.0001	0.0005
2 - blue	0.004	46.73	0.00003	0.000015
3 - green	0.003	46.73	0.00003	0.000015
4 - teal	0.004	46.73	0.00034	0.00017
5 - red	0.004	46.73	0.0002	0.0001

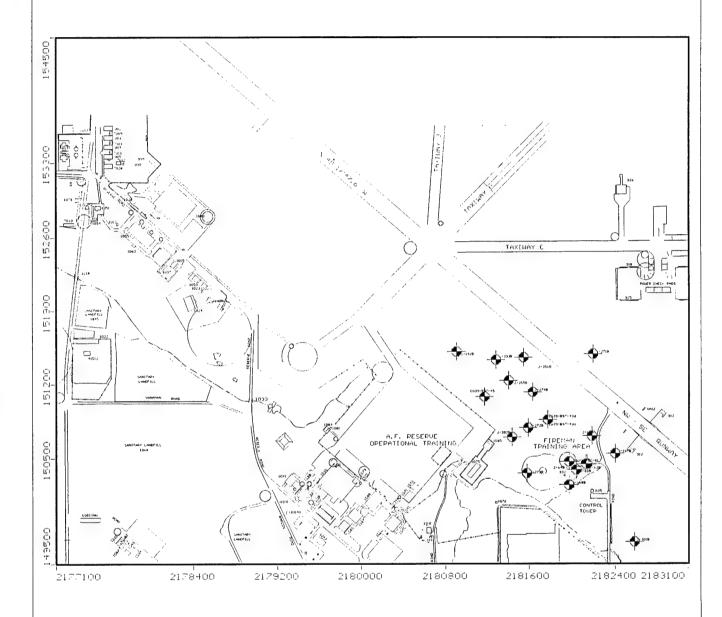
HEAD CALIBRATION

DEE	P	
1231.43	41	1232.87
1232.52	43D	1232.35
1231.48	2-274B	1231.65
1230.19	2-393B	1230.11
1231.22	45	1229.71
1231.12	2-65B	1233.71
1230.45	2-271B	1233.78
1234.67	2-142B	1228.78
1230.16	2-392B	1229.56
1231.06		
1233.86		
	1231.43 1232.52 1231.48 1230.19 1231.22 1231.12 1230.45 1234.67 1230.16 1231.06	1232.52 43D 1231.48 2-274B 1230.19 2-393B 1231.22 45 1231.12 2-65B 1230.45 2-271B 1234.67 2-142B 1230.16 2-392B 1231.06

TCE CALIBRATION

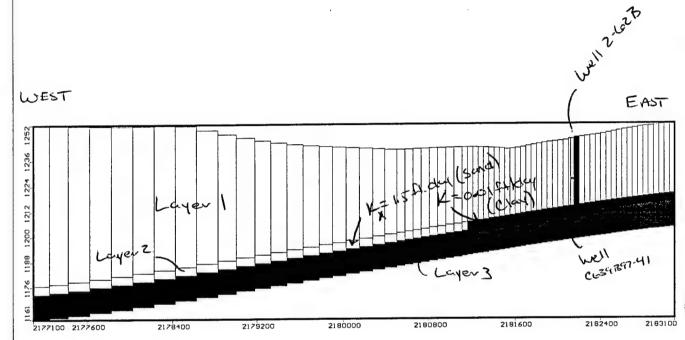
DEEP		
9440	2-274B	518
1490	2-393B	406
914	43D	169
344	45	153
75	2-65B	90
28	41	4
1	2-271B	0
0	2-142B	0
0	2-392B	0
0		
O		
	9440 1490 914 344 75 28 1 0	9440 2-274B 1490 2-393B 914 43D 344 45 75 2-65B 28 41 1 2-271B 0 2-142B 0 2-392B 0

Well Locations



Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Well IDs 23 Mar 99

MODEL CROSS SECTION



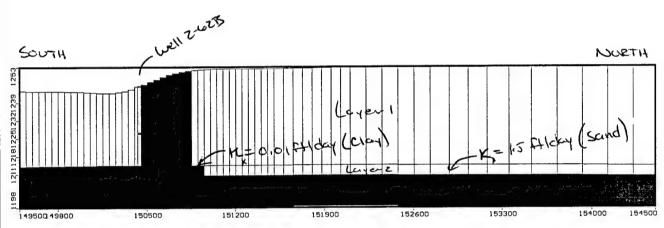
Vertical Exaggeration = 20x

Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: East to West Cross-Sect

20 Dec 99

MODEL CROSS-SECTION



Vertical Exasquation = 20x

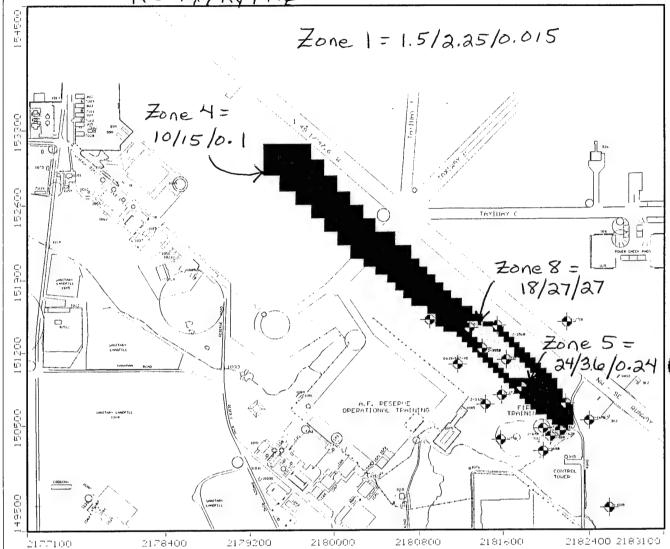
Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: South to North Cross-Sect

20 Dec 99

Visual MODFLOW v.2.50, (C) 1995-1997 Waterloo Hydrogeologic Software NC: 69 NR: 63 NL: 3

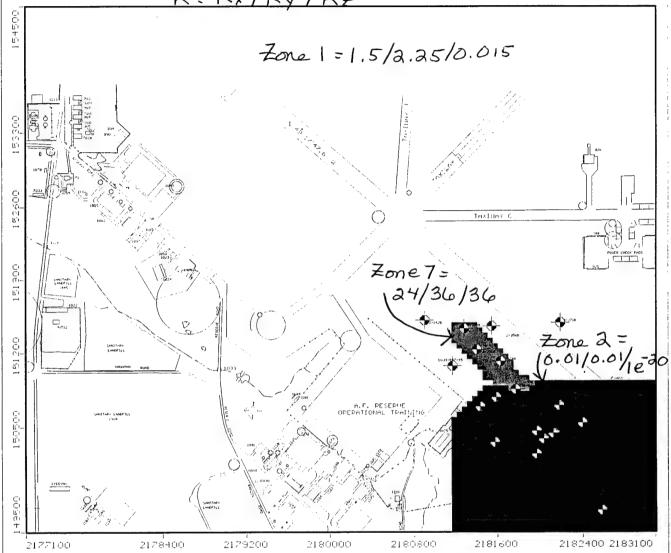
Current Column: 50

Hydraulic Conductivity (ft/day) Layer 1 K = Kx/Ky/Kz



Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Hyd. Cond. Layer 1 23 Mar 99

Hydraulic Conductivity (ft/day) Layer 2 K-Kx/Ky/Kz

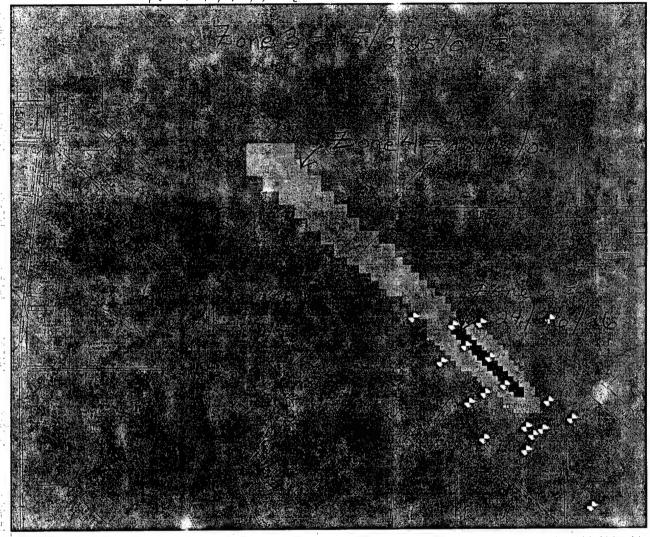


Parsons Eng Science, Inc - Denver, CO

Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Hyd. Cond. Layer 2

23 Mar 99

Hydraulic Conductivity (f+1day) ** Layer 3 K= Kx/Ky/Kz

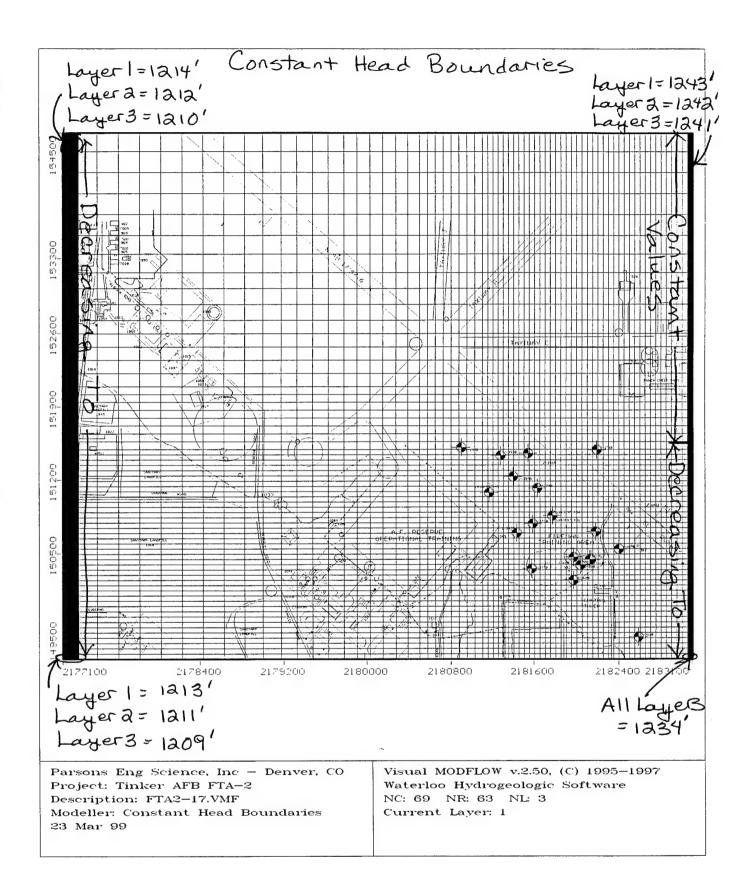


.177100 217-400 2179.1 2180000 2180900 2491000 .198.400 2181400

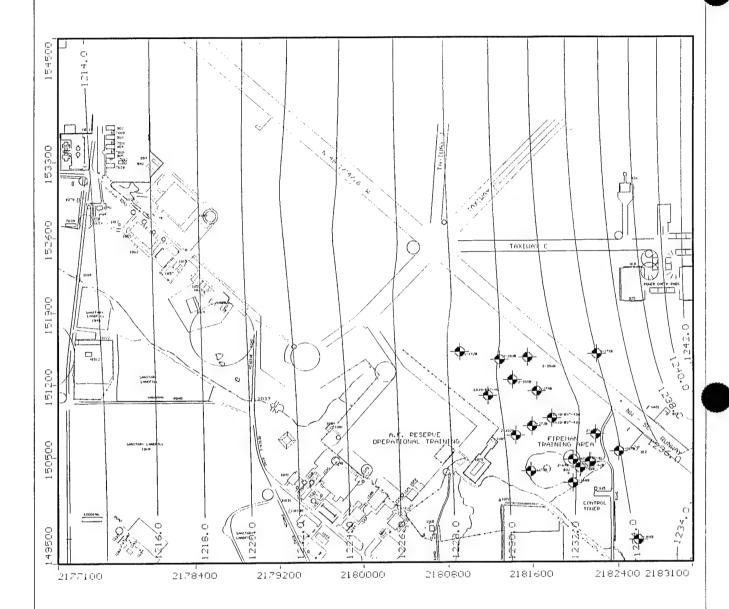
Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2

Description: FTA2-17.VMF Modeller: Hyd. Cond. Layer 3

23 Mar 99



Flow Calibration Layer 1 35 years



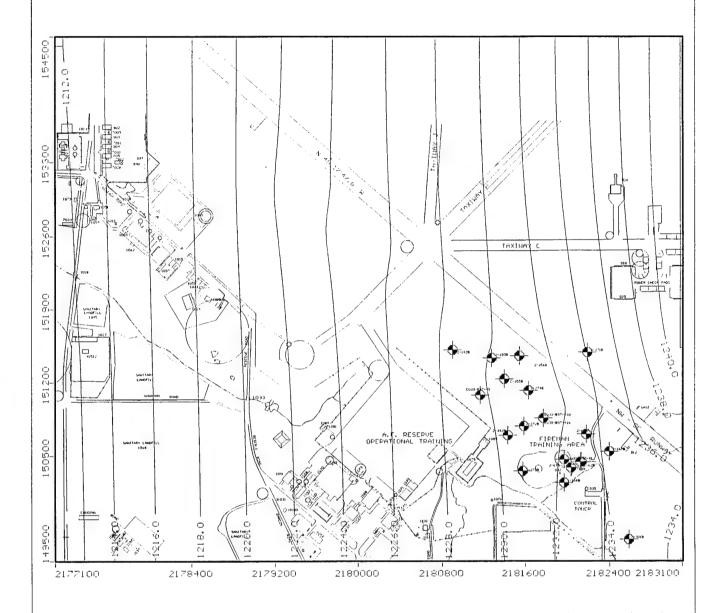
Parsons Eng Science, Inc - Denver, CO

Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Calibrated Heads Lay 1

23 Mar 99

Flow Calibration Layer 2 35 years

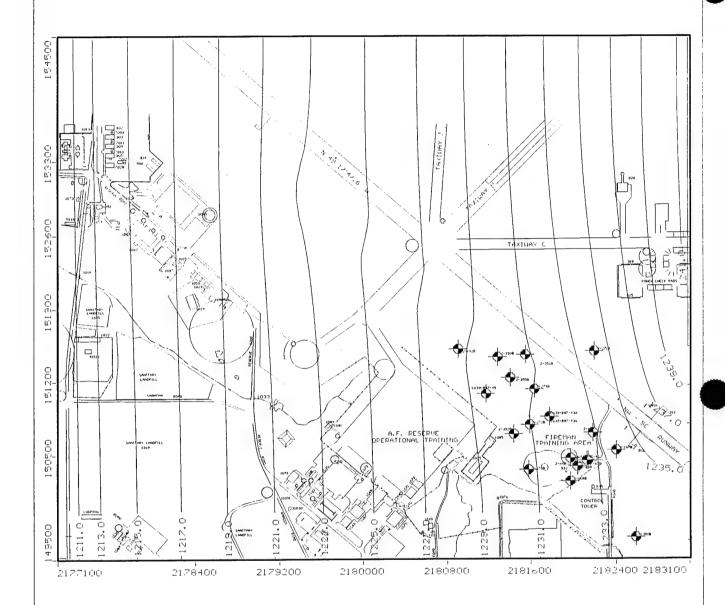


Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Calibrated Heads Lay 2

23 Mar 99

Flow Calibration Layer 3 35 years

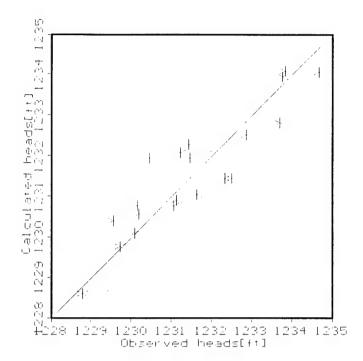


Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Calibrated Heads Lay 3

23 Mar 99

Flow Calibration 35 years



Period: 7 Step: 10

Mean error: 0.00233468

Mean abs. err: 0.558848

RMS error: 0.670091

Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2

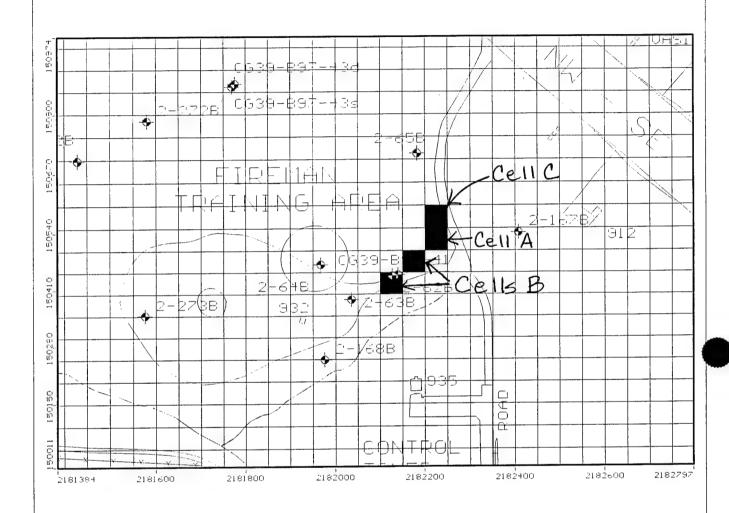
Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Obs. vs. Calc. Heads

23 Mar 99

Visual MODFLOW v.2.50, (C) 1995—1997 Waterloo Hydrogeologic Software NC: 69 NR: 63 NL: 3

Current Layer: 1

Constant Concentration Cells



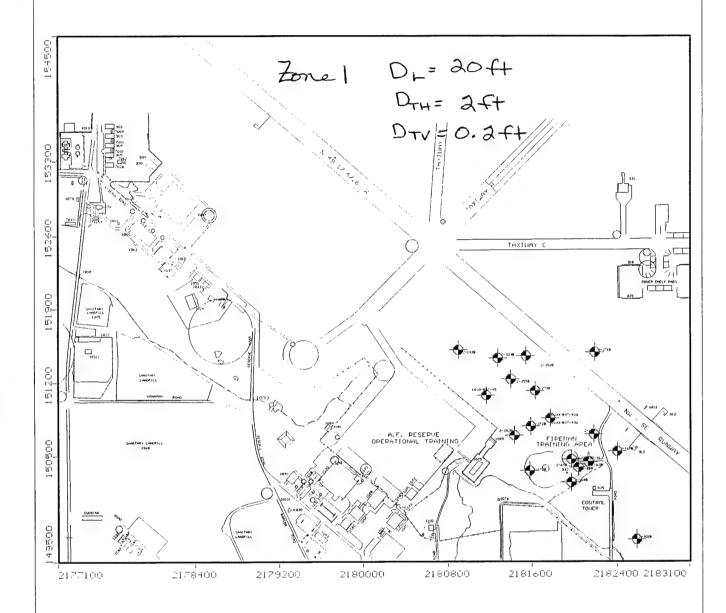
Parsons Eng Science, Inc - Denver, CO

Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Constant Concentration

23 Mar 99

Dispersion Layer 1

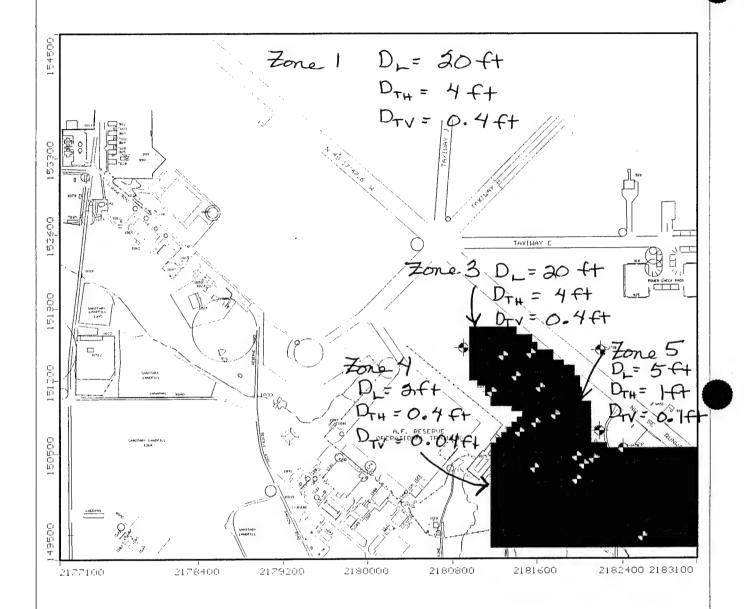


Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Dispersion Layer 1

23 Mar 99

Dispersion Layer 2

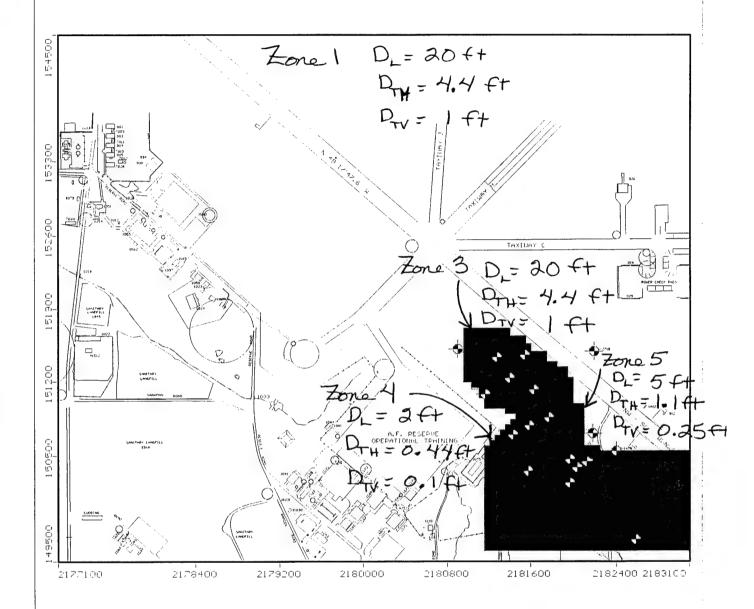


Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Dispersion Layer 2

23 Mar 99

Dispersion Layer 3

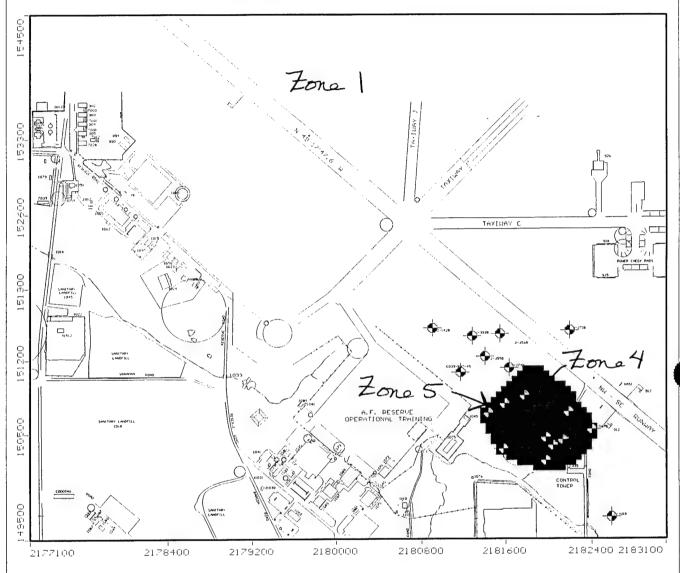


Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Dispersion Layer 3

23 Mar 99

Chemical Reactions Layer 1



Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Chem. Reaction Layer 1

23 Mar 99

Chemical Reactions Layer 2 Zone 2



2177100 2178400 2179200 2180000 2180900 2181600 2182400 2183100

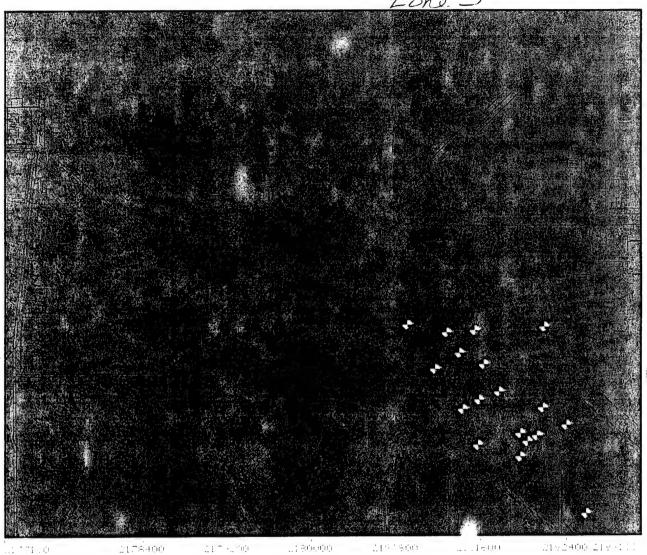
Parsons Eng Science, Inc - Denver, CO

Project: Tinker AFB FTA=2 Description: FTA2=17.VMF

Modeller: Chem. Reaction Layer 2

23 Mar 99

Chemical Reactions Layer 3 Zone 3



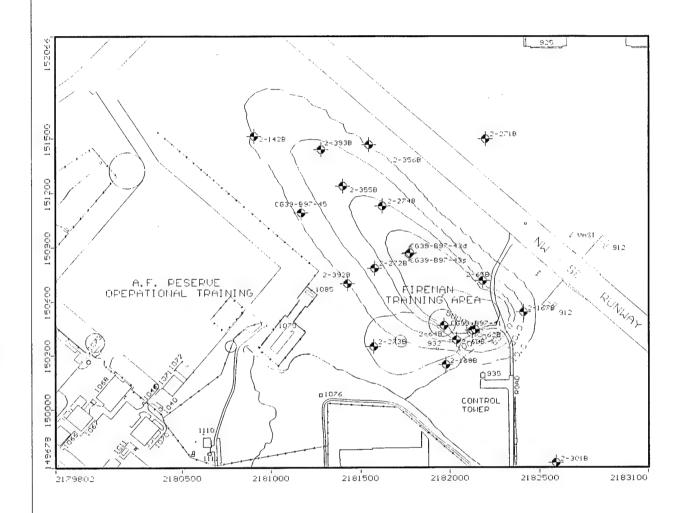
Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2

Project: Tinker AFB FTA-2 Description: FTA2-17.VMF

Modeller: Chem. Reaction Layer 3

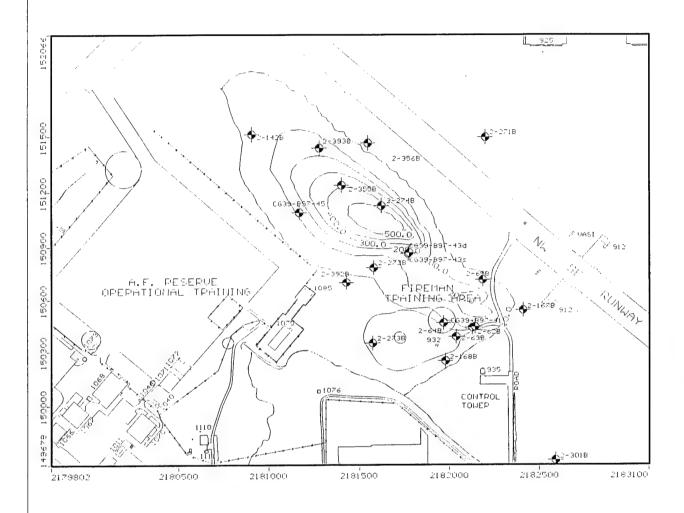
23 Mar 99

TCE Calibration Layer 1 35 years



Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Calibrated TCE Lay 1 23 Mar 99

TCE Calibration Layer 3 35 years

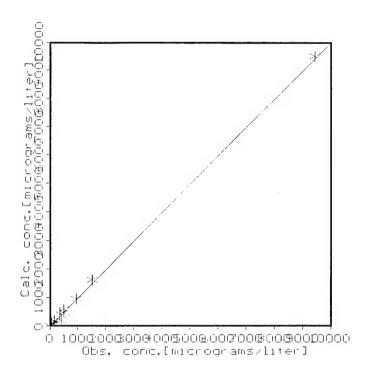


Parsons Eng Science, Inc - Denver, CO Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Calibrated TCE Lay 3

23 Mar 99

Visual MODFLOW v.2.50, (C) 1995-1997 Waterloo Hydrogeologic Software NC: 69 NR: 63 NL: 3 Current Layer: 3

TCE Calibration 35 years



Period: 7 Step: 10

Mean error: 6.84024

Mean abs. err: 44.2538 RMS error: 74.4457

Parsons Eng Science, Inc - Denver, CO

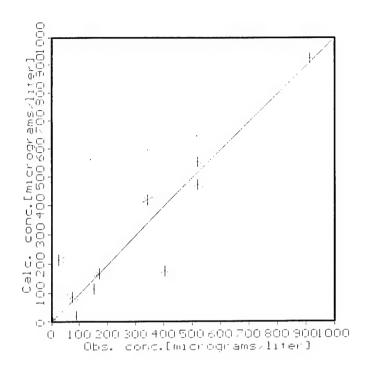
Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Obs. vs. Calc. TCE

23 Mar 99

Visual MODFLOW v.2.50, (C) 1995-1997 Waterloo Hydrogeologic Software NC: 69 NR: 63 NL: 3

Current Layer: 1

TCF Calibration 35 years



Period: 7 Step: 10

Mean error: 6.84024

Mean abs. err: 44.2538

PMS error: 74.4457

Parsons Eng Science, Inc - Denver, CO

Project: Tinker AFB FTA-2 Description: FTA2-17.VMF Modeller: Obs. vs. Calc. TCE

23 Mar 99

Visual MODFLOW v.2.50, (C) 1995-1997 Waterloo Hydrogeologic Software NC: 69 NR: 63 NL 3 Current Layer: 1

APPENDIX F MODEL INPUT AND OUTPUT FILES

APPENDIX F MODFLOW/MT3D MODEL INPUT AND OUTPUT FILES

Filename	Description
TITL 01 SD 1 STD	PRIOR LIGHT IN THE PRI
FTA217IN.ZIP	FTA2 Model Calibration Input Files
FTA217OU.ZIP	FTA2 Model Calibration Output Files
FTA2A1IN.ZIP	FTA2 Remedial Alternative One Input Files
FTA2A1OU.ZIP	FTA2 Remedial Alternative One Output Files
FTA2A2IN.ZIP	FTA2 Remedial Alternative Two Input Files
FTA2A2OU.ZIP	FTA2 Remedial Alternative Two Output Files
FTA2A3IN.ZIP	FTA2 Remedial Alternative Three Input Files
FTA2A3OU.ZIP	FTA2 Remedial Alternative Three Output Files
PKUNZIP.EXE	PKUNZIP utility executable

To decompress these file, type the following at the c:\> prompt:

A:\pkunzip a:*.zip c:\

This will create the input and output files for each model run in uncompressed ASCII format. The model files generated and the data contained therein are listed on the attached pages. All applicable input files are included. On the general MODFLOW *.LST and the MT3D *.OT and *.MAS output files are included.

Translated Inputs for Numerical Models

The following files are generated by Visual MODFLOW during translation:

Translated MODFLOW Files

MODFLOW.IN List of translated files that Visual MODFLOW creates for MODFLOW. filename.BAS Translated MODFLOW file containing data for the BASIC Package. Translated MODFLOW file containing data for the Block-Centred filename.BCF Flow Package. Translated MODFLOW file containing data for the transient filename.CH constant head package. Translated MODFLOW file containing data for the Drain Package. filename.DRN Translated MODFLOW file containing data for the filename.EVP Evapotranspiration Package. Translated MODFLOW file containing data for the General Head filename.GHB Boundary Package. Translated MODFLOW file containing data for the output control filename.OC options. Translated MODFLOW file containing data for the PCG2 solver. filename.PCG filename.RCH Translated MODFLOW file containing data for the Recharge Package. Translated MODFLOW file containing data for the River Package. filename.RIV Translated MODFLOW file containing data for the SIP solver. filename.SIP Translated MODFLOW file containing data for the SOR solver. filename.SOR Translated MODFLOW file containing data for the Horizontal Flow filename.WAL Boundary Package. Translated MODFLOW file containing data for the Well Package. filename.WEL filename.WHS Translated MODFLOW file containing data for the WHS Solver.

Translated MT3D Files

filename.AD3	Translated MT3D file containing Advection data
filename.BT3	Translated MT3D file containing Basic Transport data
filename.DP3	Translated MT3D file containing Dispersion data
filename.RC3	Translated MT3D file containing Chemical Reaction data
filename.SS3	Translated MT3D file containing Source / Sink data
MT3D.IN	Translated MT3D file containing the list of files that Visual MODFLOW creates for use in MT3D

Visual MODFLOW Output Files

After running the model, a number of result files will be generated. Some of the files generated by Visual MODFLOW may be very large (more than 100 Mbytes) especially the .BGT and the .UGN files. These files are typically in ASCII format, but some are in binary format to save disk space. These files are described below. The files marked with an asterisk (*) can get quite large especially with a transient simulation.

General

filename.ASC Visual MODFLOW output file containing data to be used by plotting programs, such as Surfer (Golden Software) - ASCII format

filename.LST Visual MODFLOW output file containing the listing information

and messages from MODFLOW - ASCII format

filename.PS Visual MODFLOW output file containing the POSTSCRIPT

graphics file - ASCII format

filename.DXF Visual MODFLOW output file containing the DXF graphic file -

ASCII format

MODFLOW

filename.DDN MODFLOW output file containing drawdown X, Y, Z heads for

each node - Binary format

filename.DVT MODFLOW output file containing drawdown versus time results-

Binary format

* filename.FLO MODFLOW output file containing output for input to MT3D cell-

by-cell flow terms (See MT3D manual for format) - Binary format

* filename.HDS MODFLOW output file containing equipotential results - Binary

format;

filename.HVT Contains MODFLOW head versus time results - Binary format;

MT3D

 filename.OT
 MT3D output file containing listing information and messages from MT3D - ASCII format;

filename.UCN MT3D output file containing unformatted concentration information

* filename.UCN MT3D output file containing unformation concentuation information
- Binary format;

* filename.MAS MT3D output file containing mass balance file - ASCII format

* filename.CNF MT3D output file containing model grid configuration file - ASCII format

APPENDIX G RESPONSES TO COMMENTS

PARSONS

Parsons Engineering Science, Inc. • A Unit of Parsons Infrastructure & Technology Group Inc.

1700 Broadway, Suite 900 • Denver, Colorado 80290 • (303) 831-8100 • Fax: (303) 831-8208 • www.parsons.com

07 December 1999

Mr. Jerry Hansen AFCEE/ERT Brooks AFB, TX

Subject: Responses to Comments on the Draft Treatability Study in Support of

Remediation by Natural Attenuation for Groundwater at Site FTA-2, Tinker

Air Force Base, Oklahoma (Contract F4164-92-D-8036-0025)

Dear Mr. Hansen:

This letter provides responses to comments received on the Draft Treatability Study (TS) in Support of Remediation by Natural Attenuation (RNA) for Groundwater at Site FTA-2, Tinker Air Force Base (AFB), Oklahoma. The draft TS was prepared by Parsons Engineering Science, Inc. (Parsons ES) for the Air Force Center for Environmental Excellence Technology Transfer Division (AFCEE/ERT) and Tinker AFB.

The intent of the TS was to determine the role of natural attenuation in remediating chlorinated solvent and hydrocarbon fuel contamination in groundwater at Fire Training Area No. 2 (FTA-2). The draft TS was submitted to AFCEE and Tinker AFB in April 1999. Comments on the draft TS were received from Sara Sayler of Tinker AFB on 17 May 1999; and from AFCEE as reviewed by Jon Atkinson, dated 08 June 1999. Responses to these comments were prepared by Parsons ES and are presented herein.

Responses to Tinker AFB Comments: 17 May 1999

Comment 1) Page 3-13, Figure 3.7, There is a well labeled 2-G2B/, I think you mean 2-62B.

<u>Parsons ES Response:</u> The well label will be corrected to 2-62B.

Comment 2) Page 4-12/4-13, since you are including the soil as possible source, it would be good to also include soil contaminant maps also.

Parsons ES Response: A description of historical and TS soil analytical results is provided in Section 4.2.1. For all previous site investigations, the only chlorinated aliphatic hydrocarbon (CAH) detected in soil was 1,1,1-trichloroethane (1,1,1-TCA) at concentrations less than 10 micrograms per kilogram (ug/kg). For this TS report, only TCE was detected, at a concentration of 19 ug/kg at well

location CG39-B97-41. Therefore, the nature and extent of soil

contamination has not been determined at FTA-2, and a soil contaminant map would not provide any further insight. Further site characterization to address this data deficiency is recommended in this TS.

- Comment 3) Page 4-15, fourth paragraph mentions TCE contamination in the Lower Sat. Zone. This is the first evidence of contamination in the LSZ, no samples have showed this previously, any clues as to how the contamination got down there? Is there a 'hole' in the confining layer, if so, where do you think it is? Maybe 2-62 or a well upgradient, created the pathway down to the LSZ?
- Parsons ES Response: TCE was detected at monitoring well 2-62A for the first time at concentration of 10.9 µg/L. Parsons ES interprets this to be a result of contaminant migration across the USZ/LSZ aquitard. While the permeability (hydraulic conductivity) of the USZ/LSZ aquitard is low relative to sediments of the USZ, it is not absolutely impermeable. Contaminant migration may occur across the USZ/LSZ aquitard given sufficient time (i.e., the approximate 30 year age of the contaminant source).
- Comment 4) Page 4-47, under the Temperature heading, the second sentence reads "Temperature can affects the types and growth...". Need to change "affects" to 'affect'.

<u>Parsons ES Response:</u> The text will be changed as indicated.

Comment 5) Page 4-52, fifth paragraph, last sentence, need to add a comma after the word "site".

<u>Parsons ES Response:</u> The text will be changed as indicated.

Comment 6) Page 5-1, just curious if you had considered using a modeling algorithm that accounts for viscosity and density differences of the contamination (such as SWANFLOW). When modeling contaminant transport over a large chunk of time, such as 35 years, the densities and viscosities can have a large effect (depending on the contaminant being considered) on contaminant migration.

Parsons ES Response: Modeling algorithms were not considered to account for viscosity and density differences of TCE relative to groundwater. Density and viscosity differences were not considered to be significant due to the dilute concentrations of TCE in groundwater (relative to the solubility limit of TCE).

Comment 7) Page 5-3, under the Grid Design heading, first sentence, "The model area for FTA-2 includes the FTA-1 source area...", I think you mean FTA-3, not FTA-1.

<u>Parsons ES Response:</u> The text will be changed to indicate the model area includes the FTA-3 fire training area.

Comment 8) Page 5-5, second paragraph, fourth sentence states "Layer 2 was defined using borehole data, and was specified as sand with a", according to the model description in the previous paragraph, "The upper layer...., a middle layer represents the intra-USZ clay horizon,....". Shouldn't Layer 2 be specified as a clay?

> It would also be helpful to include a simplified cross-section of the 3 Layers used for the model. The model layers could have pertinent model information relating to that layer printed directly on the model crosssection.

> Near the end of the second paragraph in the sentence "All three layers were assumed to be hydraulically connected, with the presence of vertical hydraulic gradients largely ignored". The wording on this statement confused me a little. Does is mean to say that the vertical hydraulic gradients were ignored because this was accounted for by hydraulically connected the layers. Would it be correct (and maybe more clear) to sav 'All three layers were assumed to be hydraulically connected, therefore, including vertical hydraulic gradients in the model was not necessary'.

Parsons ES Response: Layer 2 will be specified as intra-USZ clay where present (based on borehole logs), and as sand elsewhere. A simplified crosssection figure of the model to illustrate this will be added to Appendix E.

> Assuming all three layers to be hydraulically connected indicates that groundwater flow and contaminant transport across layer boundaries is allowed in the model. While vertical gradients have been observed at the site, the model does not attempt to duplicate those gradients. Note that this does not necessarily

preclude simulation of vertical gradients in the model, only that this was not a criteria for model calibration. contaminant transport is accomplished by simulated vertical flow and/or dispersion.

Comment 9) Page 5-7, second paragraph under the heading Aquifer Properties, last sentence, "The flow model developed for this exercise contains both a preferred pathway (a channel of sand more permeable than surrounding materials) and horizontal hydraulic conductivity anisotropy". How was the location of this channel inferred and what hydraulic conductivity was used for this channel versus non-channel in the model?

Parsons ES Response: The location of the sand channel was inferred from hydraulic conductivity measurements (Table 3.2) and the contaminant plume configuration. It should be noted this reflects the conceptual model of a single source area with a preferential flowpath oblique to the direction of maximum hydraulic gradient. This represents only one possible interpretation for the observed plume configuration. Additional data is necessary to determine if this is a correct conceptual model. hydraulic conductivity were initially selected from Table 3.2, and the values used in the calibrated model are described in Section 5.5.1.

Comment 10) Page 4-29. Table 4.10 shows some erratic behavior for the Trans DCE and the VC plots but there is no discussion about this in the text. Can you discuss why these plots appear this way?

Parsons ES Response: VC has only been detected at well 2-62B in 1997 at a concentration of 1.7 ug/L, only slightly higher that the detection limit (1 ug/L). This is not sufficient to determine trends in VC behavior.

> Non-detects for trans-1,2-DCE are incorrectly plotted at 1 ug/L for 1995 and 1996. The detection limit for trans-1,2-DCE in 1995 was not provided with the historical data table, and the detection limit in 1996 was 250 ug/L. Therefore, the figure will be corrected delete the non-detection for 1995, and to plot the non-detection for 1996 as one-half the detection limit (125 ug/L). This does not preclude erratic behavior in trans-1,2-DCE concentration, but will aid the reader in interpreting the limited data.

Comment 11) Page 5-15, under section 5.5.2.3, middle of paragraph. Mention is made of a "Section 4.6". There is no section 4.6. My copy only goes to section 4.5.

<u>Parsons ES Response:</u> The reference to "Section 4.6" will be corrected to "Section 4.5."

Responses to AFCEE Comments: 08 June 1999

General

Comment 1) The computer model grid design is flawed as described in comment eight below. Consequently, consideration should be given to reorienting the MODFLOW and MT3D grids and re-executing the modeling portion of this study.

<u>Parsons ES Response:</u> The MODFLOW and MT3D grid will not be re-oriented as indicated in the response to Specific Comment 8.

Specific

Comment 2) Page ES-2, Para 1, Line 8: To correct verb-subject agreement, recommend changing "has" to "have."

<u>Parsons ES Response:</u> The sentence will be changed as recommended.

Comment 3) Page 1-4, Sec 1.2, Para 3, Line 8: Suggest changing "air" to "aerial."

<u>Parsons ES Response:</u> The sentence will be changed as suggested.

Comment 4) Page 2-4, Sec 2.1.2, Sent 3: Suggest expanding by adding SP (Spontaneous Potential) to the list geophysical logging methods used. Additionally, recommend adding the rationale for running these downhole tests.

Parsons ES Response: The text will be expanded to include the SP logging method and the rationale for running downhole logs (i.e., required Base protocol and enhanced resolution of lithologic contacts).

Comment 5) Pages 3-9 and 3-10, Figs 3.4 and 3.5: Suggest adding a bar scale for horizontal distance.

<u>Parsons ES Response:</u> Bar scales for horizontal distance will be added to Figures 3.4 and 3.5 as suggested.

Comment 6) Page 3-16, Sec 3.3.2.1, Para 3 and Page 3-19, Para 3: To better reflect the problematic accuracy/certainty of hydraulic conductivity (K) values calculated from slug tests, recommend reporting K values to two significant figures.

<u>Parsons ES Response:</u> Values for hydraulic conductivity (K) will be reported to two significant figures throughout the final TS.

Comment 7) Page 4-28, Sec 4.3.2, Para 2, Sent 1: Recommend changing "would expected" to "would be expected."

<u>Parsons ES Response:</u> The text will be changed as recommended.

Comment 8) Page 5-3, Sec 5.4.1, Sent 3: Although the direction of maximum hydraulic gradient is the rationale given for orienting the model grid, the long axis of the grid should be parallel to features or conditions (e.g., faults, sand channels) controlling predominant flow direction. Anderson and Woessner (Applied Groundwater Modeling, 1992) note (p. 5-1) that: "The horizontal plane of the grid should be aligned so that the x and y coordinate axes are colinear with K_x and K_y." Because the longitudinal axis of the TCE plume and the sand channel unit have an approximate bearing of N 45 W, the long axis of the grid should also have this approximate compass direction.

Parsons ES Response: A preferential flowpath has not been delineated, but was inferred based on relative changes in hydraulic conductivity and a conceptual model that included a singular source area. As recommended in the draft TS, additional source and site characterization should be performed to verify the conceptual model for the site. Parsons ES suspects multiple sources may be present at FTA-2. Parsons ES requests that the model not be reoriented, as the work required to modify the model is not warranted until the conceptual model can be validated.

Comment 9) Page 5-5, Sec 5.4.2, Para 1, Line 12: Recommend deleting "and."

<u>Parsons ES Response:</u> The text will be changed as recommended.

- Comment 10) Page 5-11, Sec 5.5.1, Para 3, Sent 4: The stated K_y to K_x ratio is 2:1; however, a work sheet containing MODFLOW input in Appendix E annotates a K_y to K_x ratio of 1.5. Based on the MODFLOW BCF file for Alternative 3 provided in Appendix F, horizontal anisotropy ratios were 6.0 and 3.0 for layer 1 (Upper USZ) and layer 3 (Lower USZ), respectively. These apparent discrepancies should be addressed in the text and resolved appropriately.
- Parsons ES Response: The stated Ky to Kx ratio will be corrected to 1.5:1, and the text will be modified to indicate anisotropy under the Block-Centered-Flow option in MODFLOW was set to 6.0 for layer 1 and 3.0 for layer 3.
- Comment 11) Page 5-13, Sec 5.5.1, Para 1, Sent 1: A RMS error of 0.67 feet and the total relief of the water table over the modeled area of 20 feet (based on Fig 5.2) results in a calibration error of 3.4 percent, not 11.1 percent. Recommend the text state the relief of the water table used to calculate calibrated hydraulic head errors.
- Parsons ES Response: The RMS error of 0.67 feet is calculated over the relief of the monitoring wells used for calibration, not over the entire model domain. The relief in observed elevations for the area of the available calibration data is approximately 7 feet, as shown on Figure 5.3. Therefore the calculated calibration error is 11.1 percent. The text will be changed to indicate that the calibration error is calculated using the change in relief in the water table for the calibration wells only.
- Comment 12) Page 5-15, Sec 5.5.2.4, Sent 1: Suggest reporting porosity to two significant figures because it reflects better the accuracy/certainty of site-specific values.
- Parsons ES Response: Page 5-15, Section 5.5.2.4, Sentence 1 refers to retardation coefficients. Retardation values will be reported to two significant figures throughout the text to better reflect the accuracy/certainty of site-specific values (i.e., organic carbon).

- Comment 13) Page 5-16, Sec 5.5.2.6: Recommend stating here, or in Section 5.7, whether the modeling results and other data document a steady-state or nonsteady-state plume.
- <u>Parsons ES Response:</u> The text will be changed in Sections 5.5.2.6 and 5.7 to indicate that the modeling results represent a nonsteady-state contaminant plume.
- Comment 14) Page 5-20, Sec 5.6.3, Para 2, Last Sent: To better reflect accuracy/certainty of simulated values, recommend rounding concentrations to no more than four significant figures.
- <u>Parsons ES Response:</u> Simulated concentrations will be rounded to four figures on Table 5.3 and in Section 5.6.3 as suggested.
- Comment 15) Page 6-4, Sec 6.2.3.1, Para 1, Last Sent: To better reflect accuracy/certainty of the groundwater velocities, recommend rounding these values to no more than three significant figures.
- <u>Parsons ES Response:</u> Values for groundwater velocity will be rounded to no more than three significant figures as suggested.
- Comment 16) Page 6-13, Sec 6.3.3: Suggest adding to the heading "Source Characterization and Removal" to better reflect content of sentence one.
- <u>Parsons ES Response:</u> The heading will be changed as suggested.
- Comment 17) Page 6-17, Sec 6.4.1.3 and Table 6.3: Here and throughout the remainder of Section 6.4, recommend rounding total cost estimates to no more than four significant figures to better reflect accuracy/certainty of these estimates.
- <u>Parsons ES Response:</u> Total costs will be rounded to no more than four significant figures throughout Section 6.4 as suggested.
- Comment 18) Page 6-21, Sec 6.4.3.1, Para 2, Sent 1:

- a. The stated pumping rate for the seven extraction wells is given as 5 gpm. The MODFLOW well package file for Alternative 3 contains two wells with a pumping rate of 3.1 gpm and five wells with a 1.8-gpm pumping rate. These discrepancies should be discussed in the text and resolved.
- b. Two of the proposed extraction wells for the upper USZ sand layer are located outside of the relatively high-K channel where K is assigned a low 1.5 ft/day value. Because of this low K, these two wells cannot pump at a very high rate, cannot develop large capture zones, and cannot extract significant TCE mass. Consequently, recommend that these wells be deleted from Alternative 3 or that they be repositioned in the buried sand-filled channel. Additionally, suggest that an optimization computer code (e.g., MODMAN, REMAX) be used to help identify optimal well number, locations and pumping rates.

Parsons ES Response:

- a. Parsons ES verified the pumping rates for the seven extraction wells specified in Alternative 3 is 5.0 gallons per minute. New MODFLOW input files will be generated for the Final TS to ensure they are correct.
- b. Parsons ES concurs. The two proposed extraction wells should be moved to the model area with higher K-values. However, it should be noted that the assignment of hydraulic conductivity in the model is limited by the available K data (Table 3.2), and that the exact location of high hydraulic conductivity sand channels is not known with certainty. Therefore, further site characterization is recommended before exact locations are selected for extraction wells and before trying to optimize the collection system.
- Comment 19) Page 6-22, Fig 6.3: Suggest depicting and labeling the proposed extraction wells as described on page 6-21, paragraph two.
- <u>Parsons ES Response:</u> The proposed extraction wells for Alternative 3 will be added to Figure 6.3 as suggested.
- Comment 20) Page 6-24, Sec 6.5, Para 2, Sent 1: Recommend stating whether Figures 5.4 and 6.1 substantiate the assertion of an expanding plume.

Mr. Jerry Hansen 07 December 1999 Page 10

Parsons ES Response: The following sentence will be added after Page 6-24, Section 6.5, Paragraph 2, Sentence 1: "Under all three remedial alternatives the TCE plume is predicted to expand, but without impact to potential receptors (Figures 6.1, 6.2, and 6.3)."

Comment 21) Pages 7-2 and 7-3, Sec 7.2, Line 10: Suggest changing "will be sampled" to "would be sampled" because the LTM is a proposed plan, not an approved, funded or implemented plan.

Parsons ES Response: The text will be changed as suggested.

Comment 22) Page 7-4, Sec 7.6 and Page 7-7, Table 7.2: Recommend rounding total cost estimates to no more than four significant figures to better reflect accuracy/certainty of these estimates.

Parsons ES Response: The cost estimates in Section 7 will be rounded to no more than four significant figures as suggested.

Comment 23) Page 8-2, Sec 8, Para 4, Line 3: Recommend correcting the typo "wit."

Parsons ES Response: The text will be corrected as recommended.

If you have any questions, have additional comments, or require additional information, please call me at (303) 831-8100.

Sincerely,

PARSONS ENGINEERING SCIENCE, INC.

Brue M Henry Bruce M. Henry, P.G. Project Manager

File 729691.34000 cc:

APPENDIX H

FINAL ADDENDUM TO THE TREATABILITY STUDY IN SUPPORT OF REMEDIATION BY NATURAL ATTENUATION FOR GROUNDWATER AT SITE FTA-2, TINKER AIR FORCE BASE, OKLAHOMA

FINAL ADDENDUM TO THE TREATABILITY STUDY IN SUPPORT OF REMEDIATION BY NATURAL ATTENUATION FOR GROUNDWATER AT SITE FTA-2

at

TINKER AIR FORCE BASE OKLAHOMA

December 1999

Prepared for:

AIR FORCE CENTER FOR ENVIRONMENTAL EXCELLENCE BROOKS AIR FORCE BASE SAN ANTONIO, TEXAS

and

TINKER AIR FORCE BASE OKLAHOMA

Prepared by:

Parsons Engineering Science, Inc. 1700 Broadway, Suite 900 Denver, Colorado 80290

TABLE OF CONTENTS

rag	e
LIST OF ACRONYMS AND ABBREVIATIONSi	V
1.0 INTRODUCTION	1
1.1 Scope and Objectives 1.2 Site Background 1.3 Groundwater Monitoring	1
2.0 HYDROGEOLOGY	2
2.1 Upper Saturated Zone	6 8
3.0 NATURE AND EXTENT OF CONTAMINATION	8
3.1 Fuel Hydrocarbons In Groundwater	2 7 7 7
4.0 NATURAL ATTENUATION ANALYSIS2	9
4.1 Field-Scale Contaminant Mass Loss	0
4.2Presence of Daughter Products and CAH Molar Fractions34.3Electron Donors34.4Chloride as an Indicator of Dehalogenation34.5ORP as an Indicator of Redox Processes44.6Alternate Electron Acceptors and Metabolic Byproducts44.6.1Dissolved Oxygen44.6.2Nitrate + Nitrite44.6.3Ferrous Iron4	2 7 7 1 1 4 4
4.6.4 Sulfate .4 4.6.5 Methane, Ethane and Ethene .4 4.7 Other Geochemical Parameters .4 4.7.1 Ammonia .4 4.7.2 Alkalinity .4	4 5 5

TABLE OF CONTENTS (Continued)

		Page
4.8	Summary of Biodegradation of Chlorinated Solvents	45
5.0	CONCLUSIONS AND RECOMMENDATIONS	46
6.0	REFERENCES	47
	ment A - Analytical Data ment B - Responses to Comments	
	LIST OF TABLES	
No. 1 2 3 4 5	Title Groundwater Analytical Protocols	5 10 13
	LIST OF FIGURES	
No. 1 2 3 4 5 6 7 8 9	Title Sampling Locations – Fire Training Area 2 Upper Saturated Zone Potentiometric Surface August 1997 and April 1999. Lower Saturated Zone Potentiometric Surface August !997 and April 1999. Concentrations of PCE, 1,1-DCE, and VC in USZ Groundwater August 1997 and April 1999 Concentration of TCE in Upper USZ Groundwater August 1997 and April 1999. Concentration of TCE in Lower USZ Groundwater August 1997 and April 1999. TCE Concentration versus Time for Selected Wells – Upper USZ Sand TCE Concentration versus Time for Selected Wells – Lower USZ Sand Concentration of cis-1,2-DCE in Upper USZ Groundwater August 1997	7 15 17 18
10	and April 1999	22
11	and April 1999	

TABLE OF CONTENTS (Continued)

LIST OF FIGURES (Continued)

No.	Title	Page
13	Concentration of 1,2-DCA in USZ Groundwater August 1997 and April	
	1999	28
14	Total Chlorinated Ethenes versus Time	31
15	Total Chlorinated Ethanes versus Time	33
16	Total Chlorobenzenes versus Time	34
17	Molar Fraction of Chlorinated Ethenes Over Time – Well 2-62B	35
18	Molar Fraction of Chlorinated Ethenes versus Distance August 1997	36
19	Molar Fraction of Chlorinated Ethenes Over Time Well 2-65B	
20	Redox Potential of USZ Groundwater August 1997 and April 1999	42
21	Concentration of Dissolved Oxygen in USZ Groundwater August 1997	
	and April 1999	43

LIST OF ACRONYMS AND ABBREVIATIONS

 $^{\circ}C$ degrees Celsius ٥F degrees Fahrenheit 1.1-dichloroethane 1.1-DCA 1.2-DCA 1,2-dichloroethane 1,2-DCB 1,2-dichlorobenzene 1.3-DCB 1.3-dichlorobenzene 1,4-dichlorobenzene 1,4-DCB 1,1-DCE 1,1-dichloroethene micrograms per liter μg/L Air Force Base AFB

AFCEE Air Force Center for Environmental Excellence

AOC Area of Concern

BTEX benzene, toluene, ethylbenzene, and xylenes

CAHs chlorinated aliphatic hydrocarbons

CBs chlorobenzenes
cis-1,2-DCE cis-1,2-dichloroethene
DO dissolved oxygen
FTA-2 Fire Training Area 2

ft/day feet per day ft/ft foot per foot

LSZ lower saturated zone
LTM long-term monitoring
MCL maximum contaminant level

mg/L milligrams per liter

MTBE methyl tertiary-butyl ether

mV millivolts N nitrogen

NAPL non-aqueous phase liquid

NRMRL National Risk Management Research Laboratory

ORP oxidation-reduction potential Parsons ES Parsons Engineering Science, Inc.

PCE tetrachloroethene

RCRA Resource Conservation and Recovery Act RNA Remediation by Natural Attenuation

SWMU solid waste management unit

TCE trichloroethene
TMBs trimethylbenzenes

trans-1,2-DCE trans-1,2-dichloroethene

TS Treatability Study

USEPA United States Environmental Protection Agency

USZ upper saturated zone VC vinyl chloride

1.0 INTRODUCTION

This document was prepared for the Air Force Center for Environmental Excellence (AFCEE) by Parsons Engineering Science, Inc. (Parsons ES) as an addendum to the Draft Treatability Study (TS) in Support of Remediation by Natural Attenuation (RNA) for Fire Training Area No. 2 (FTA-2), Tinker Air Force Base (AFB), Oklahoma (Parsons ES, 1999), under Air Force contract number F41624-92-D-8036. The TS was completed to evaluate the use of intrinsic remediation (natural attenuation) with long-term monitoring (LTM) for remediation of groundwater contaminated with chlorinated solvents and fuel hydrocarbons at FTA-2.

1.1 Scope and Objectives

This addendum documents the effectiveness of natural attenuation at Site FTA-2 by summarizing results of a groundwater sampling event conducted by researchers from the United States Environmental Protection Agency (USEPA) National Risk Management Research Laboratory (NRMRL) Subsurface Protection and Remediation Division in April 1999. Groundwater was initially sampled in July and August 1997 for the TS. The focus of this addendum is to evaluate changes in concentrations of dissolved chlorinated aliphatic hydrocarbons (CAHs), or chlorinated solvents; CAH plume extent; and natural attenuation mechanisms through time. Concentrations of dissolved benzene, toluene, ethylbenzene, and xylenes (BTEX) also are evaluated. Results, calculations, and predictions presented in the TS are used as the basis for comparison. Analytical data for the April 1999 sampling event is presented in Attachment A.

1.2 Site Background

Tinker AFB covers approximately 5,000 acres in the southeastern Oklahoma City metropolitan area. The Base has supported air operations since its founding as the Midwest Air Depot in July 1941. The Base was formally activated by the Air Force in March 1942 and serves as an international repair depot for a variety of aircraft, weapons, and engines.

Site FTA-2 is located in the south-central portion of the Base, and was established as a temporary, unlined pit used infrequently for fire training exercises between 1962 and 1966. Standard operating procedures reportedly included adding water to the pit to saturate the soil and reduce fuel infiltration. Fuel was then brought in by tank truck, placed on top of the water, ignited, and extinguished using water and foam. Any residues were left in the pit to evaporate and infiltrate until the next fire training exercise. Available aerial photographs indicate that the site is currently a gently sloping, open grassy area with no visible evidence of former training practices. Records for facility operation do not exist. Therefore, data on composition, frequency, and quantity of fuel or extinguishing agents are not available.

Tinker AFB currently operates under a Resource Conservation Recovery Act (RCRA) Hazardous Waste Management Permit issued by the USEPA (July 1, 1991). This permit requires Tinker AFB to investigate all solid waste management units (SWMUs) and

Areas of Concern (AOCs), including FTA-2, and to perform corrective action at those sites identified as posing a threat to human health or the environment.

Sources of dissolved groundwater contamination at FTA-2 are not clearly defined. It is likely that residues from fuel or extinguishing agents have leached into soil and groundwater at the fire training pit and are a primary source of groundwater contamination. However, the exact locations and nature of the releases is not clear from the available data. Secondary sources also may be present that are not associated with the fire training pit. Residual non-aqueous phase liquid (NAPL) has not been identified during either previous or current subsurface investigations.

Engineered remediation has not been implemented at the site. Natural attenuation is the only process acting to reduce dissolved CAH concentrations and mass at the site. Additional site information, including site background, geology, and hydrogeology, is available in the TS (Parsons ES, 1999).

1.3 Groundwater Monitoring

Since the July/August 1997 groundwater sampling event performed for the TS (Parsons ES, 1999), researchers from the USEPA NRMRL collected an additional round of groundwater samples in April 1999. Sampling locations for the site are shown on Figure 1. Nineteen monitoring wells/points were evaluated at FTA-2 in April 1999. Monitoring wells/points sampled in 1997 but not in 1999 include CG39-B97-45, 2-142B, 2-167B, 2-271B, 2-356B, and 2-142A. Monitoring wells 2-271B and 2-142A were not recommended for LTM in the TS. The USEPA NRMRL elected not to sample the other four locations.

Samples collected in April 1999 were analyzed in the field for temperature, pH, conductivity, dissolved oxygen (DO), oxidation reduction potential (ORP), alkalinity, sulfide, and ferrous iron. Additionally, samples were analyzed at the USEPA NRMRL in Ada, Oklahoma for BTEX, trimethylbenzenes (TMBs), methyl tertiary-butyl ether (MTBE), CAHs, chlorobenzenes (CBs), chloroform, sulfate, chloride, nitrate + nitrite as nitrogen (N), ammonia, methane, ethene, and ethane. Analytical methods for April 1999 are summarized in Table 1.

2.0 HYDROGEOLOGY

Prior to purging and collecting samples from each monitoring location, depth to groundwater was measured to the nearest 0.1 foot in April 1999. Groundwater elevation data and monitoring well/point completion information are summarized in Table 2. The hydrogeology at FTA-2 is characterized by groundwater flow within an upper saturated zone (USZ) and a lower saturated zone (LSZ) as described in the TS (Parsons ES, 1999). The low permeability USZ/LSZ aquitard separates the two zones, and extends across the site. The USZ is generally considered an unconfined aquifer. However, because the underlying LSZ may exist as an unconfined aquifer in the area, the USZ may exhibit characteristics of a perched aquifer in the vicinity of FTA-2.

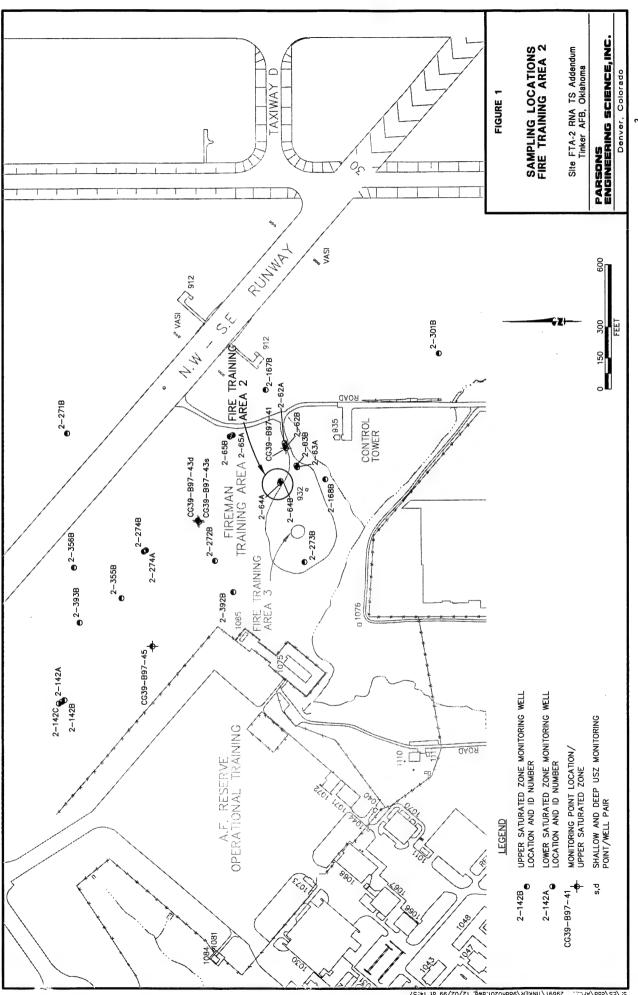


TABLE 1 GROUNDWATER ANALYTICAL PROTOCOLS APRIL 1999

SITE FTA-2 RNA TS ADDENDUM TINKER AFB, OKLAHOMA

MATRIX Analyte	метнор	FIELD (F) OR ANALYTICAL LABORATORY (L)
WATER		
Redox Potential	Direct-reading meter	F
Dissolved Oxygen	Direct-reading meter	F
pН	Direct-reading meter	F
Conductivity	Direct-reading meter	F
Temperature	Direct reading meter	F
Ferrous Iron (Fe ⁺²)	Titrimetric, Chemetrics Method or equivalent	F
Alkalinity (Carbonate [CO ₃ ⁻²] and Bicarbonate [HCO ₃ ⁻¹])	Titrimetric, Hach Method 8221 or equivalent	F
Sulfide	Chemetrics Method or equivalent	F
Ammonia (NH ₃)	Lachat FIA Method 10-107-06-1	L
Nitrate + Nitrite	Lachat FIA Method 10-107-04-2-A	L
Chloride	Waters Capillary Electrophoresis Method N-601	L
Sulfate	Waters Capillary Electrophoresis Method N-601	L
Methane, Ethane, Ethene VOCs (BTEX, TMBs,	RSKSOP-194 ^{a/} /RSKSOP-175	L
and MTBE) ^{b/} VOCs (CAHs, CBs, and	RSKSOP-122	L
chloroform) ^{c/}	RSKSOP-148	L

^a/RSKSOP = Robert S. Kerr Laboratory standard operating procedure.

b/VOCs = volatile organic compounds; BTEX = benzene, toluene, ethylbenzene, and xylenes;

TMBs = trimethylbenzenes; MTBE = methyl tertiary-butyl ether.

c/CAHs = chlorinated aliphatic hydrocarbons; CBs = chlorobenzenes.

MONITORING WELL SUMMARY AND GROUNDWATER ELEVATIONS SITE FTA-2 RNA TS ADDENDUM TABLE 2

TINKER AFB, OKLAHOMA

	Screened	Elevation	Ground	Survey	Survey	Depth to Water	Elevation of Water	Depth to Water	Elevation of Water
Well/Borehole	Interval	Datum	Surface	Northing	Easting	August-1997	Table - August 1997	April-1999	Table - April 1999
Identification	(Feet bgs)2'	(Feet msl) ^{b'}	Elevation	(State Plane)	(State Plane)	(Feet btoc)	(Feet msl)	(Feet btoc)	(Feet msl)
Monitoring Points and Monitoring Wells - Upper Saturated Zone	and Monitoring	Wells - Upper	Saturated Zone						
CG39-B97-41	34 - 39	1245.20	1245.57	150440.90	2182126.70	12.33	1232.87	11.2	1234.0
CG39-B97-43s	30 - 35	1252.42	1252.72	150861.52	2181770.08	19.90	1232.52	19.4	1233.0
CG39-B97-43d	44.5 - 49.5	1252.54	1252.56	150866.94	2181777.04	20.19	1232.35	19.4	1233.1
CG39-B97-45	36.5 - 41.5	1241.82	1242.24	151085.60	2181167.63	12.11	1229.71	NM ^{d/}	NM
2-62B	14-24	1245.94	1245.94	150444.52	2182140.56	14.51	1231.43	13.9	1232.0
2-63B	18-23	1243.28	1243.28	150386.11	2182035.36	12.06	1231.22	10.9	1232.4
2-64B	16-26	1245.59	1245.59	150467.40	2181967.76	14.11	1231.48	13.0	1232.6
2-65B	37-47	1250.81	1250.81	150712.88	2182183.20	17.10	1233.71	17.0	1233.8
2-142B	32-42	1242.09	1242.45	NA ^e ′	NA	13.31	1228.78	MN	NM
2-167B	19-29	1250.70	1247.70	150539.81	2182408.35	16.03	1234.67	NM	NM
2-168B	10-20	1243.00	1240.10	150249.00	2181976.69	12.55	1230.45	11.2	1231.8
2-271B	36-46	1252.13	1252.41	151494.42	2182196.43	18.35	1233.78	NM	NM
2-272B	25-35	1248.91	1249.38	150784.70	2181581.74	17.79	1231.12	17.2	1231.7
2-273B	17-27	1238.42	1238.92	150351.11	2181576.36	8.26	1230.16	7.2	1231.2
2-274B	37-42	1248.32	1248.77	151126.23	2181625.92	16.67	1231.65	16.9	1231.4
2-301B	16-26	1244.00	1244.00	149701.00	2182586.00	10.14	1233.86	NM	NM
2-355B	19.9-29.9	1245.92	1246.61	151233.46	2181400.66	15.73	1230.19	15.9	1230.0
2-356B	27.8-37.8	1249.30	1249.75	151462.29	2181548.00	18.24	1231.06	NM	MM
2-392B	36 - 46	1244.99	1245.30	150696.27	2181430.89	15.43	1229.56	15.1	1229.9
2-393B	35 - 45	1245.97	1246.19	151434.08	2181280.87	15.86	1230.11	16.2	1229.8
Monitoring Wells - Lower Saturated Zone	- Lower Saturate	ed Zone							
2-62A	54-64	1246.21	1246.21	150447.89	2182152.17	57.32	1188.89	57.0	1189.2
2-63A	53-63	1243.39	1243.39	150389.96	2182043.02	54.70	1188.69	54.0	1189.4
2-64A	99-95	1246.05	1246.05	150470.21	2181959.37	57.34	1188.71	56.6	1189.5
2-65A	92-99	1250.98	1250.98	150698.28	2182189.88	61.48	1189.50	61.2	1189.8
2-274A	76-81	1248.75	1249.10	151115.88	2181631.94	59.20	1189.55	58.9	1189.9
2-142A	72-82	1241.76	1242.30	NA	NA	53.03	1188.73	NM	NM
2/ Foot has indies	Heet has indicates elevation in feet below around surface	Feet helow aroun	d enreace		beauseau toN = MN /p	escured			

^{a/} Feet bgs indicates elevation in feet below ground surface.
^{b/} Feet msl indicates elevation in feet above mean sea level.

c/ Feet btoc indicates feet below top of casing.

2.1 Upper Saturated Zone

Figure 2 is a groundwater potentiometric surface map for the USZ based on data from August 1997 and April 1999. The horizontal hydraulic gradient in the USZ in August 1997 was approximately 0.003 to 0.01 foot per foot (ft/ft), while in April 1999 the hydraulic gradient was similar at 0.004 to 0.008 ft/ft. The direction of maximum horizontal hydraulic gradient in the USZ for 1999 is consistent with gradient data observed in 1997 (Parsons ES, 1999). Assuming isotropic horizontal hydraulic conductivity, horizontal flow direction in the vicinity of FTA-2 was generally towards the west-southwest in both 1997 and 1999.

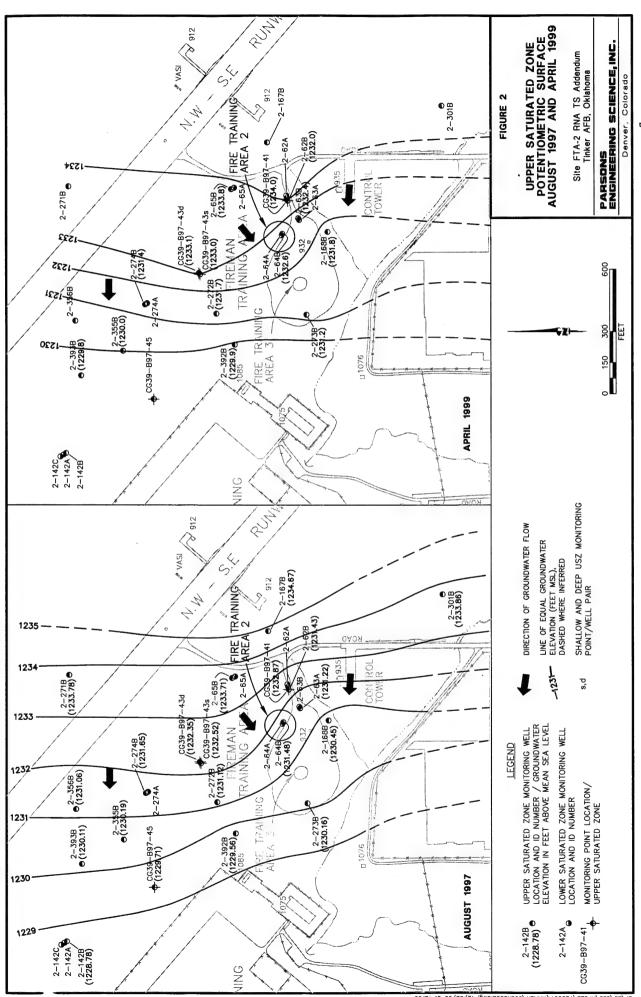
The USZ can be further divided into the upper USZ sand and the lower USZ sand based on a persistent clay layer that separates the two sands in the southern portion of the site. Vertical hydraulic gradients between the upper and lower USZ sand intervals were calculated for monitoring well pair CG39-B97-43s and CG39-B97-43d, and well pair 2-62B and CG39-B97-41. The vertical gradients were calculated by dividing the water level elevation difference between two clustered wells by the distance between the midpoints of the saturated portions of the well screens. The vertical groundwater gradient at well pair CG39-B97-43s and CG39-B97-43d was 0.012 ft/ft in a downward direction in 1997 and 0.007 ft/ft in an upward direction in 1999. The vertical groundwater gradient at well pair 2-62B and CG39-B97-41 was 0.081 ft/ft in an upward direction in 1997, and 0.11 ft/ft in a likewise upward direction in 1999.

The upward vertical gradients observed in 1999 are typical to aquifers that are semi-confined or confined. While the USZ is regionally considered an unconfined aquifer, semi-confined to confined conditions may exist locally within the lower USZ sand interval. At FTA-2, an upward USZ groundwater gradient within the source area at well pair 2-62B and CG39-B97-41, and along the contaminant plume axis at well pair 2-62B and CG39-B97-41, may impact contaminant transport at the site.

While the upper and lower USZ sand intervals are hydraulically connected, where it is developed, the intermediate clay and silt interval may act as a significant vertical permeability barrier to local groundwater flow and contaminant transport.

Hydraulic conductivity in the USZ was estimated by performing slug tests in seven monitoring wells (CG39-B97-41, CG39-B97-43s, CG39-B97-43d, CG39-B97-45, 2-392B, 2-393B, and 3-335B) in August 1997. Well 2-392B, located on the western edge of the site, had the lowest hydraulic conductivity of 1.1 feet/day (ft/day). The remainder of the wells, located to the east of 2-392B, had hydraulic conductivities that averaged 15 ft/day. The difference in hydraulic conductivities from east to the west suggest the presence of low permeability conditions to the west (downgradient) of the site, which may impact contaminant transport.

Using an estimated effective porosity of 0.20, an average hydraulic conductivity of 14 ft/day, and an average hydraulic gradient of 0.006 ft/ft, the average advective groundwater flow velocity in the upper USZ sand interval in 1997 was calculated to be 0.42 ft/day (152 ft/yr) (Parsons ES, 1999). Similarly, using an average hydraulic gradient of 15 ft/day, the average advective groundwater flow velocity in the lower USZ sand



interval in 1997 was calculated to be 0.46 ft/day (167 ft/yr). Advective groundwater velocities in preferential or anisotropic flow directions other than the direction of maximum hydraulic gradient would be less than these computed values.

2.2 Lower Saturated Zone

The top of the LSZ coincides with the base of the USZ/LSZ aquitard. The LSZ potentiometric surface locally lies beneath the base of the USZ/LSZ aquitard under unconfined conditions. Locally the LSZ may be confined where the LSZ potentiometric surface rises above the USZ/LSZ aquitard. Figure 3 is a groundwater potentiometric surface map for the LSZ based on data from August 1997 and April 1999. The direction of maximum hydraulic gradient and groundwater flow in the LSZ is towards the southwest. The horizontal hydraulic gradient in April 1999 was approximately 0.002 ft/ft. The flow direction and gradient in the LSZ for 1999 are consistent with the southwesterly groundwater flow direction and a hydraulic gradient of 0.003 ft/ft observed in August 1997 (Parsons ES, 1999).

3.0 NATURE AND EXTENT OF CONTAMINATION

Both fuel hydrocarbons and CAHs have been detected in groundwater at FTA-2. Because CAHs have been detected at significantly higher concentrations than the BTEX compounds, the TS and this addendum focus on the natural attenuation potential for CAHs at the site.

3.1 Fuel Hydrocarbons In Groundwater

Results of analysis for dissolved fuel hydrocarbon compounds in groundwater at FTA-2 are presented in Table 3. Compound specific fuel hydrocarbons detected in groundwater at FTA-2 are BTEX; 1,2,3-TMB; and 1,2,5-TMB. In 1997, benzene was detected at two locations, 2-62B and 2-168B, at concentrations of 6.0 and 2.5 micrograms per liter (μ g/L), respectively. In 1999, benzene was detected only at location 2-62B at a concentration of 3.9 μ g/L, a decrease from the 1997 concentration at this location. Toluene was only detected at location 2-62B in 1997 at a concentration of 1.2 μ g/L. In 1999, toluene was detected at concentrations of 13.6 μ g/L and 7.5 μ g/L at locations 2-64A and 2-392B, respectively. Ethylbenzene and xylenes were detected for the first time in 1999 at location 2-274A at concentrations of 1.6 μ g/L and 1.0 μ g/L, respectively. 1,2,3-TMB was detected in 1999 at a low concentration of 2.7 μ g/L at well location 2-168B; and 1,3,5-TMB was detected at a low concentration of 3.7 μ g/L at well location 2-392B.

Only benzene exceeded its respective USEPA maximum contaminant level (MCL) for drinking water (USEPA, 1996). Benzene exceeded the MCL of 5 μ g/L at a concentration of 6.0 μ g/L at well location 2-62B in 1997. However, no fuel hydrocarbons exceeded respective MCLs in 1999. Because of the limited occurrence and low concentrations of fuel hydrocarbons detected at the site, natural attenuation of BTEX is not further analyzed. However, the presence of low levels of anthropogenic carbon (fuel hydrocarbons) may be beneficial to biodegradation of CAHs as discussed in Section 3.2

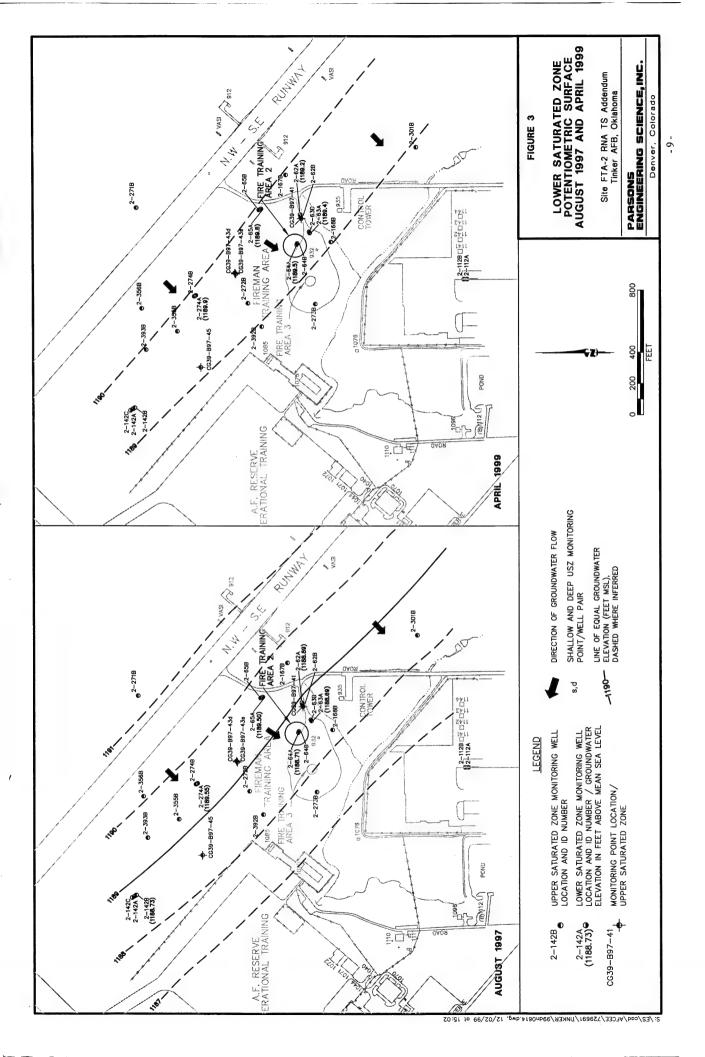


TABLE 3
SUMMARY OF FUEL HYDROCARBONS IN GROUNDWATER
SITE FTA-2 RNA TS ADDENDUM
TINKER AFB, OKLAHOMA

										3		
		Total Fuel			Ethyl-				Total	:		
Sample	Sample	Carbon	Benzene	Toluene	penzene	p-Xylene	m-Xylene	o-Xylene	BTEX ^{a/}	1,3,5-TMB ^{b'}	1,2,4-TMB	1,2,3-TMB
Location	Date	(μg/L) ^{c/}	(µg/L)	(µg/L)	(μg/L)	(µg/L)	(μg/L)	(μg/L)	(hg/L)	(μg/L)	(μg/L)	(µg/L)
	MCL	- و	5	1000	700	10,000	10,000	10,000	1	1	:	:
Upper Saturated Zone	Zone											
CG39-B97-41	Aug-97	ND ^{g/}	ND	QN	QN	ND	ND	ND	ND	ND	ND	ND
	Apr-99	NA	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
CG39-B97-43s	Aug-97	ND	ND	ND	QN	ND	ND	ND	ND	ND ND	ND	ND
	Apr-99	NA	ND	ND	ND	ND	ND	ND	ND	ND	QN	ND
CG39-B97-43d	Aug-97	QN	ND	ND	QN	ND	ND	ND	ND	ND	ND	ND
	Apr-99	NA	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
CG39-B97-45	Aug-97	ND .	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-62B	Jul-97	1,140	6.0	1.2	ND	ND	ND	$BLQ^{i'}$	7.2	BĽQ	ND	ND
	Apr-99	NA	3.9	ND	ND	ND	ND	ND	3.9	ND	ND	ND
2-63B	Jul-97	QN	ND	ND	QN	QN	ND	ND	QN	ND	ND	ND
	Apr-99	NA	BLQ	ND	ND	ND	ND	ND	ND	ΩN	QN	QN
2-64B	Aug-97	QN	ND	ND	ND	ND	ND	ND	ND	QN	ND	QN
	Apr-99	NA	BLQ	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-65B	Aug-97	QN	QN	QN	ND	ND	ND	ND	ND	NO	ND	QN
	Apr-99	NA	BLQ	BLQ	ND	ND	BLQ	ND	ND	ND	ND	ND
2-142B	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-167B	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-168B	76-Inf	5.8	2.5	ND	ND	ND	ND	ND	ND	ND	ND	ND
	Apr-99	NA	BLQ	ND	ND	ND	1.3	ND	1.3	ND	BLQ	2.7
2-271B	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-272B	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	QN	ND	ND
	Apr-99	NA	ND	ND	ND	BLQ	BLQ	ND	ND	ND	ND	ND
2-273B	Aug-97	ND	ND	ND	ND	QN	ND	ND	ND	ND	ND	NO
	Apr-99	NA	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-274B	Aug-97	ND	ND	N	QN	ND	ND	ND	ND	ND	ND	ND
	Apr-99	NA	BLQ	ND	ND	ND	ND	ND	ND	ND	ND	ND

- 10 -

TABLE 3 (Continued) SUMMARY OF FUEL HYDROCARBONS IN GROUNDWATER SITE FTA-2 RNA TS ADDENDUM TINKER AFB, OKLAHOMA

	0	Total Fuel			Ethyl-				Total			
Sample	Sample	Carbon	Benzene	Toluene	benzene	p-Xylene	m-Xylene	o-Xylene	$BTEX^{a'}$	1,3,5-TMB [™]	1,2,4-TMB	1,2,3-TMB
Location	Date	(μg/L) ^{c/}	(µg/L)	(µg/L)	(μg/L)	(μg/L)	(μg/L)	(μg/L)	(μg/L)	(µg/L)	(μg/L)	(μg/L)
	MCL	- د	5	1000	700	10,000	10,000	10,000	1	-	1	:
Upper Saturated Zone	ed Zone											
2-355B	Aug-97	ND	ND	ND	ND	ND	QN	ND	QN	ND	ND	ND
	Apr-99	NAN	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
3-356B	Sep-97	NA	ND	ND	ND	ND	ND	ND	ND	ND	ND	NA
2-392B	Aug-97	QN	ND	ND	ND	QN	QN	ND	QN	ND	ND	ND
	Apr-99	Y Z	ND	7.5	ND	ND	ND	ND	7.5	3.7	ND	ND
2-393B	Aug-97	BLQ	BLQ	QN	ND	ND	ND	ND	QN	ND	ND	QN
	Apr-99	NA	ND	BLQ	ND	ND	ND	ND	ND	ND	ND	ND
Lower Saturated Zone	ed Zone											
2-62A	Jul-97	ND	ND	QN	ND	QN	ND	ND	ΩN	QN	QN	ΩN
	Apr-99	NA	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-63A	Jul-97	ND	ND	ON	ND	ND	ND	ND	ND	ND	ND	ND
	Apr-99	Ϋ́	N Q	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-64A	Aug-97	QN	QN	ND	QN	ND	ND	QN	ND	ND	QN	ND
	Apr-99	NA	QN	13.6	ND	ND	ND	ND	13.6	ND	ND	ND
2-65A	Aug-97	QN	QN	ND	ND	ND	ND	ND	QN	ND	ND	ND
	Apr-99	AN	QN	ND	ND	ND	BLQ	ND	ND	ND	ND	ND
2-274A	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
	Apr-99	NA	ND	ND	1.6	ND	1.0	ND	2.6	ND	ND	ND
2-142A	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
a' BTEX = ber	BTEX = benzene, toluene, ethylbenzene, and xylenes.	ethylbenzene,	and xylenes.			J	e' = no standard listed.	fard listed.				
b' TMB = trimethlbenzene.	rethlbenzene.					-	" Standard liste	" Standard listed is for total xylenes.	lenes.			
$_{\rm c'}$ $_{\rm \mu g/L} = {\rm micr}$	$\mu g/L = micrograms per liter.$	ï				, Line	g' ND = Not Detected.	betected.				
d' MCL = Maximum Contaminant Level, State of Oklahoma OAC 165:25-3-65+A21, or	cimum Contami	nant Level, St	ate of Oklaho	oma OAC 16:	5:25-3-65+		h' NA = not analyzed	nalyzed.				

^{or} MCL = Maximum Contaminant Level, State of Oklahoma OAC 165:25-3-65+A21, or U.S. EPA Primary Drinking Water Regulations, May 14, 1996.

¹⁷ BLQ = below limit of quantitation (1 microgram per liter).

3.2 Chlorinated Aliphatic Hydrocarbons

As discussed in the TS, chlorinated solvents can be transformed, directly or indirectly, by biological processes. Biodegradation of CAHs and CBs results from the same general processes that bring about biodegradation of fuel hydrocarbons. However, a more complex series of processes often is involved, and CAHs and CBs may act as either substrates (electron donors) or electron acceptors depending upon prevailing geochemical conditions.

Whereas fuel hydrocarbons are biodegraded in essentially one step by acting as an electron donor/carbon source, CAHs and CBs may undergo several types of biodegradation involving multiple steps. CAHs may undergo biodegradation through four different pathways: use as an electron acceptor, use as an electron donor, abiotic transformation, or cometabolism, which is degradation resulting from exposure to a catalytic enzyme fortuitously produced during an unrelated process. At a given site, one or all of these processes may be operating, although at many sites the use of CAHs as electron acceptors appears to be the most likely.

In a pristine aquifer, native organic carbon is utilized as an electron donor and DO is utilized first as the prime electron acceptor. Where anthropogenic carbon (e.g., fuel hydrocarbons, less-chlorinated CAHs, or CBs with four or fewer chlorines) is present, it may also be utilized as an electron donor. Most chlorinated solvents that can act as electron donors have thus far only been demonstrated to do so under aerobic conditions, with the notable exception of vinyl chloride (VC) (Bradley and Chapelle, 1996). After the DO is consumed, anaerobic microorganisms typically use native electron acceptors (as available) in the following order of preference: nitrate, manganese, ferric iron oxyhydroxide, sulfate, and finally carbon dioxide. Evaluation of the distribution of these electron acceptors can provide evidence of where and how fuel hydrocarbon or CAH biodegradation is occurring. Because CAHs may be used as electron acceptors or electron donors (in competition with other acceptors or donors), maps showing the distribution of these compounds also will provide evidence of the types of biodegradation processes acting at a site.

3.2.1 Chlorinated Ethenes

Chlorinated ethenes detected in groundwater at FTA-2 include tetrachloroethene (PCE), trichloroethene (TCE), 1,1-dichloroethene (1,1-DCE), cis-1,2-dichloroethene (cis-1,2-DCE), trans-1,2-dichloroethene (trans-1,2-DCE), and VC. Data for chlorinated ethenes and other chlorinated compounds are presented in Table 4.

PCE

PCE was detected in 4 of the 14 groundwater samples collected from the USZ at Site FTA-2 in April 1999 (Figure 4). Concentrations of PCE in groundwater ranged from less than the limit of quantitation (<1.0 μ g/L) up to 154 μ g/L at monitoring location 2-65B. Compared to August 1997 sampling results, PCE concentrations decreased at locations 2-62B (from 6.1 to 3.7 μ g/L) and 2-63B (from 1.9 to 1.0 μ g/L). PCE concentrations increased at locations 2-64B (from <1.0 μ g/L to 1.3 μ g/L) and 2-65B (from 52 μ g/L to

-13-

TABLE 4 CHLORINATED ALIPHATIC HYDROCARBONS IN GROUNDWATER SITE FTA-2 RNA TS ADDENDUM TINKER AFB, OKLAHOMA

								IINKEKA	IINKEK AFB, UKLAHUMA	Y.						
Sample	Sample	PCE ^{a/}	TCE®	1,1-DCE ^{8/}	cis-1,2-DCE	cis-1,2-DCE trans-1,2-DCE	VC ^{a/}	1,1,1-TCA*/	1,1-DCA"	1,2-DCA	Chlorobenzene	1,2-DCB"	1,3-DCB	1,4-DCB	Carbon Tetrachloride	Chloroform
Location	Date	(μg/L) ^{b/}	(μg/L)	(µg/L)	(μg/L)	(μg/L)	(μg/L)	(μg/L)	(μg/L)	(μg/L)	(µg/L)	(μg/L)	(μg/L)	(μg/L)	(μg/L)	(μg/L)
	MCL°'	5	5	7	70	100	2	200	/P	5	100	009	009	75	5	-
Surface Water										,						
Creek Grab	Apr-99	ND _{e/}	1.7	ND	1.0	ND	ND	ND	ND	ND	ND	ND	ND	ND	QN	ND
Upper Saturated Zone	Zone															
CG39-B97-41	Aug-97	ΩN	3.6	ND	<i>y</i>	ND	ND	ND	ND		QN	:	ND	QN	QV Q	N Q
	Apr-99	N	6.4	ND	***	ND	ND	ND	ND	ND	ND		ND	ND	ND	ND
CG39-B97-43s	Aug-97		1,490	Ξ	183	6.1	Ω	QN	ND	2.4		ND	1.6	QN	QN	1
	Apr-99	ı	751	1	9.86	3.3	ND	ND	ND	1.5	-	ND	1.1	ND	ND	-
CG39-B97-43d	Aug-97	QV	691	QN	31.8	2.4	ND	ND	ND	1	ND	ND	:	QN	QN	ND
	Apr-99	ND	95.6	QN	19.1	1.2	ND	ND	ND		ND	ND	ND	ND	ND	ND
CG39-B97-45	Aug-97	ND	153	ND	22.6		ND	ND	ND	1.4	ND	ND		ND	ND	ND
2-62B	Jul-97	6.1	>2000	9.5	1,200	76.2	1.7	ND	ND	260	117	950	34.8	168	QN	5.3
	^{/8} (dnb) 26-Inf	ı	9,440	i	1,110	77	N Q	ΩN	Ω	239	104	696	25	123	QN	ı
	Apr-99	3.7	6,200	5.1	650	47.5	ND	ND	ND	37.2	19.5	101	8.0	15.5	ND	5.0
2-63B	Jul-97	1.9	75.4	-	191	3.6	9.9	ND	4.6	1	ND	-	ND	ND	QN	ND
	Apr-99	1.0	53.7	ł	126	3.2	3.6	ND	3.4	:	ND	ND	-	ND	ND	ND
2-64B	Aug-97	!	914	-	154	17.6	1	ND	ND	1.3	1	:	1.2	ł	QN	-
	Apr-99	1.3	2,130	1.2	328	31.1	ND	QN	ΩN	2.0	1.5	ND	2.8	ND	QN	1.5
	Apr-99(dup)		1,870	1.1	302	26.7	ND	ND	ND	2.0	1.3	ND	2.2	ND	QN	1.1
2-65B	Aug-97	52	06		37	2.5	1.3	ND	1	3.2	9.5	3.6	i	8.1	QN	QN
	Apr-99	154	250	1.4	147	8.9	5.1	ND	1.7	18.3	51.6	23.8	1.4	9.6	ND	-
2-167B	Aug-97	ΩN	ND	ND	QN	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-168B	Jul-97	1	1.4	:	27.6	ND	ND	QN.	1.2	ND	QN	ND	Q	ΩN	Q	QN
	Apr-99	ND	1.0	ND	6.3	ND	ND	ND	:	ND	ND	ND	ND	ND	QN	ND
2-271B	Aug-97	ΩN	ΩN	QN	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-272B	Aug-97	QN	28.1	ND	2.1	ND	ND	ND	N Q	1	QN	QN QN	ND	ND	ΩN	ND
	Apr-99	***	455		34.9	1.9	ND	ND	ND	1	:	QN	1	QN	QN	:
2-273B	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	ND	ΩN	ND	ND	ND	QN	ND
	Apr-99	ND	ND	ND	ND	ND	ND	ND	ND	ND	QN	ND	ND	QN	QN	ND
2-274B	Aug-97		818	i	51.7	2.2	ND	QN	ND	Ξ	1	ND	ı	QN	QN	i
	Aug-97(dup)	ł	206	ł	53.4	2.1	QN	ND	ND	1.4	ı	ΩN	1	ND	QN	i
	Apr-99	1	942	ł	99.2	4.0	ND	QN	ND	2.1	ł	ΩN	1.6	ND	QN	I
	Apr-99(dup)	:	1,060	ND	9.06	4.0	ND	ND	ND	2.0	1	ND	1.6	ND	QN	1
2-355B	Aug-97	QN	344	ND	50	i	ND	ND	ND	1.8	I	ND	i	Q	QN	į
	Apr-99	ND	186	ND	28.5		ND	ND	ND	1	ND	QN	ı	ND	QN	ND
2-356B ^{IV}	Sep-97	QN	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND

022/733939/30.xls, Table 4

TABLE 4 CHLORINATED ALIPHATIC HYDROCARBONS IN GROUNDWATER SITE FTA-2 RNA TS ADDENDUM TINKER AFB, OKLAHOMA

Sample	Sample	PCE*	TCE^{u}	1,1-DCE*	cis-1,2-DCE trans-1,2-DCE	trans-1,2-DCE	$VC^{a'}$	1,1,1-TCA*/	1,1-DCA*	1,2-DCA	Chlorobenzene	1,2-DCB*/	1,3-DCB	1,4-DCB	Carbon Tetrachloride	Chloroform
Location	Date	(μg/L) ^{b/}	(µg/L)	(µg/L)	(μg/L)	(µg/L)	(μg/L)	(μg/L)	(μg/L)	(µg/L)	(µg/L)	(μg/L)	(μg/L)	(µg/L)	(μg/L)	(μg/L)
	MCL	5	5	7	70	100	2	200	/p	5	100	009	009	75	5	;
Upper Saturated Zone (Continued)	1 Zone (Contin	ned)														
2-392B	Aug-97	QN	1	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ΩN	QN	ΩN
	Apr-99	ND	-	ND	ND	ND	ND	ND	ND	ND	ND	QN	QN	QN	QN	ND
2-393B	Aug-97	QN	406	1	54.3	1	ND	ND	ND	45.4	1	ND	l	QN	QN	1
	Aug-97(dup)	ND	425	1	54.5	ŀ	N	ND	ND	49.2	i	ND	I	QN	QN	1
	Apr-99	ND	114	ND	13.3	ND	ND	ND	ND		ND			ND	ND	ND
2-142B	Aug-97	QN	1	QN	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
Lower Saturated Zone	d Zone															
2-62A	Jul-97	Q	10.9	ND	;	ND	QN	ND	QN	QN	ND	1	ND	ND	QN	ΩN
	Apr-99	ND	i	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-63A	Jul-97	Q	1	ND	ND	ND	ND	ND	ND	QN	ND	ł	ND	ND	QN	ND
	Apr-99	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-64A	Aug-97	Q	ΩN	ND	ND	ND	ND	QN	ND	QN	ND	ND	ND	QN	ND	ND
	Apr-99	Q.	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-65A	Aug-97	QN	QN	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
	Apr-99	ND	l	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-274A	Aug-97	QN	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	QN	ND	ND	ND
	Apr-99	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2-142A	Aug-97	ND	ND	ND	ND	ND	ND	ND	ND	Ω	ND	ND	ND	QN	ND	ND
^{a/} PCE = tetrachloroethene, TCE = trichloroethene, DCE = dichloroethene, VC = vinyl chloride, TCA	oroethene, TCE	= trichloroe	thene, DCE =	= dichloroethe.	ne, VC = vinyl ch	hloride, TCA = tri	= trichloroethane	e, d	= no standard listed	l listed.						

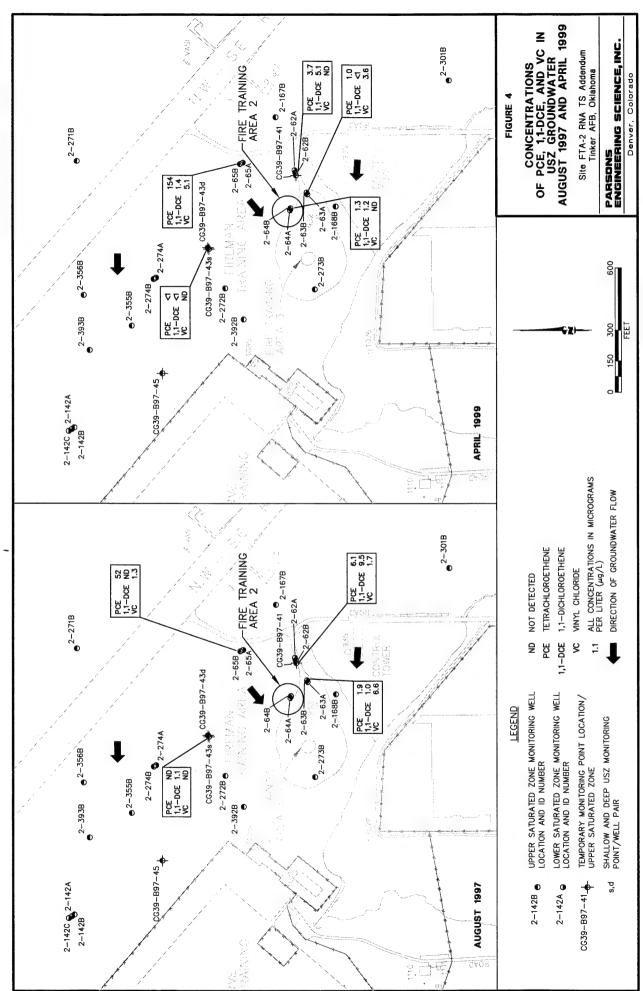
 $DCA = dichloroethane, \ and \ DCB = dichlorobenzene.$

b µg/L = micrograms per liter.
d MCL = Maximum Contaminant Level, U.S. EPA Primary Drinking Water Regulations, May 14, 1996.

e' ND = not detected.

 $^{\it ff}$ --- = below calibration limit (1 microgram per liter).

 $^{B^{\prime}}$ Dup = field duplicate of preceding sample.



154 μ g/L), respectively. The concentration of 154 μ g/L at monitoring well 2-65B is above the USEPA MCL for PCE of 5 μ g/L.

Monitoring well 2-64B is located in the former FTA-2 fire pit area. The other six locations correspond to an area upgradient (east) and crossgradient of the FTA-2 fire pit area. It is unclear to what extent the presence of PCE is associated with the former FTA-2 fire pit or other unknown sources at the site. The elevated concentration of PCE at well location 2-65 does suggest a source other than the FTA-2 fire pit.

TCE

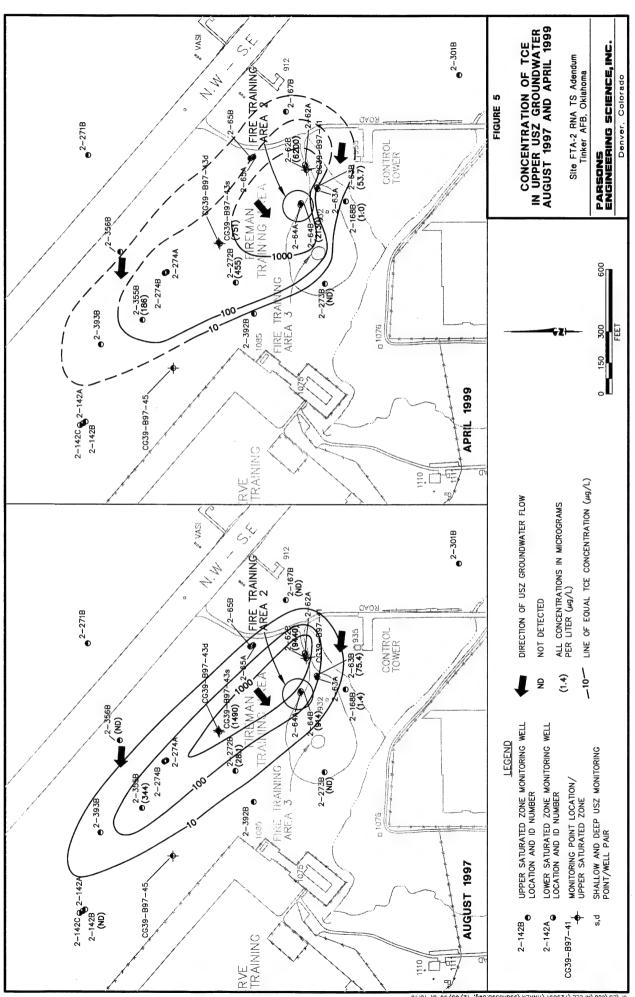
As measured in April 1999, TCE was detected in 12 of the 14 groundwater samples collected from the USZ aquifer at FTA-2 (Table 4). TCE concentrations detected in USZ groundwater in 1999 ranged from <1.0 μ g/L to 6,200 μ g/L at 2-62B, with concentrations at 11 locations exceeding the USEPA MCL of 5 μ g/L. Figures 5 and 6 show the distribution of TCE in the upper USZ sand and the lower USZ sand, respectively. Comparison of Figures 5 and 6 indicates that the areal distribution and concentration of TCE in groundwater differs between the upper and lower USZ sands.

Concentrations of TCE in the upper USZ sand increased at two locations (2-64B and 2-272B) and decreased at five locations (CG39-B97-43s, 2-62B, 2-63B, 2-168B, and 2-355B). Figure 5 suggests that the contaminant plume is migrating toward the west in the direction of maximum hydraulic gradient, based on an increase in concentration of TCE at well locations 2-64B and 2-272B. TCE concentration versus time is plotted for select monitoring wells in Figure 7. The concentration of TCE in 1999 at 2-62B (6,200 μ g/L) remained within the range of historical values for this location (4,300 μ g/L in 1996 to 9,440 μ g/L in 1997). Based on the elevated TCE concentrations at well 2-62B, this location is suspected to be located near the primary contaminant source for the site. Downgradient from 2-62B, well 2-64B has experienced a historical increase in TCE concentration from 96 μ g/L in 1993 to a maximum of 2,130 μ g/L in 1999. This is likely a result of downgradient transport of TCE from a source near the 2-62B well location.

The concentration of TCE also has increased over time at well 2-272B, located downgradient from well CG39-B97-43s (Figure 5). This may be a result of downgradient migration of contamination from an area near well CG39-B97-43s. While there are limited historical data for wells CG39-B97-43s and 2-355B, concentrations of TCE decreased from August 1997 to April 1999 at both these locations.

In the lower USZ sand, concentrations of TCE increased at three wells (CG39-B97-41, 2-65B, and 2-274B) and decreased at two wells (CG39-B97-43d and 2-393B). Figure 6 shows the TCE plumes for August 1997 and April 1999, and suggests that the extent of the contaminant plume in the lower USZ sand has not increased significantly. TCE concentration versus time is plotted for four of these wells on Figure 8. Concentrations of TCE did increase significantly from 1997 to 1999 for well 2-274B (from 518 μ g/L to 942 μ g/L) and for well 2-65B (from 90 μ g/L to 250 μ g/L).

The increase in the concentration of TCE at well 2-274B suggests vertical migration of contaminants from the upper USZ sand to the lower USZ sand. Alternately, elevated



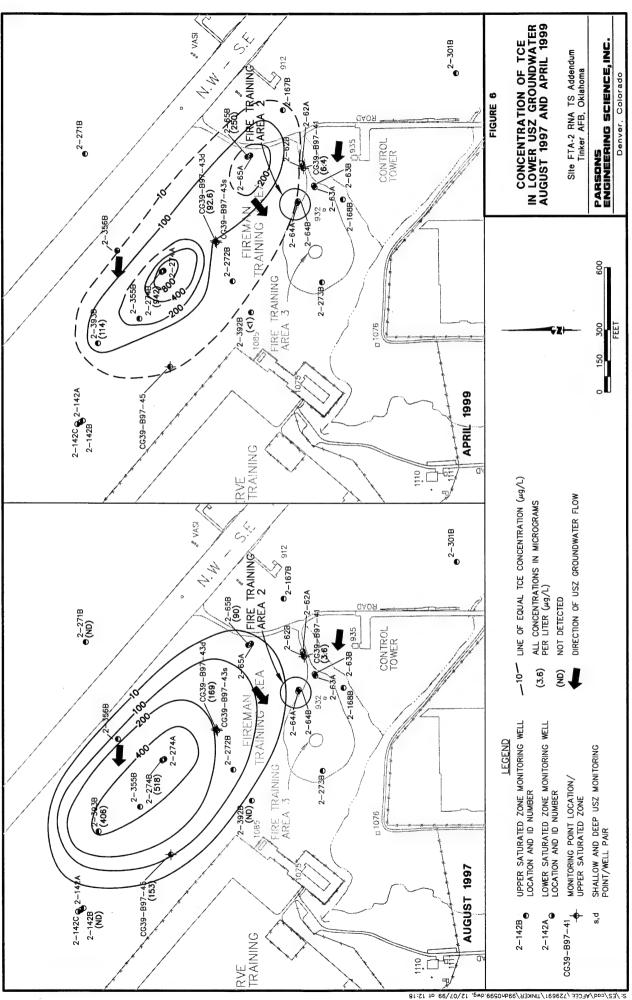


FIGURE 7

TCE CONCENTRATION VERSUS TIME FOR SELECTED WELLS
UPPER USZ SAND
SITE FTA-2 RNA TS ADDENDUM

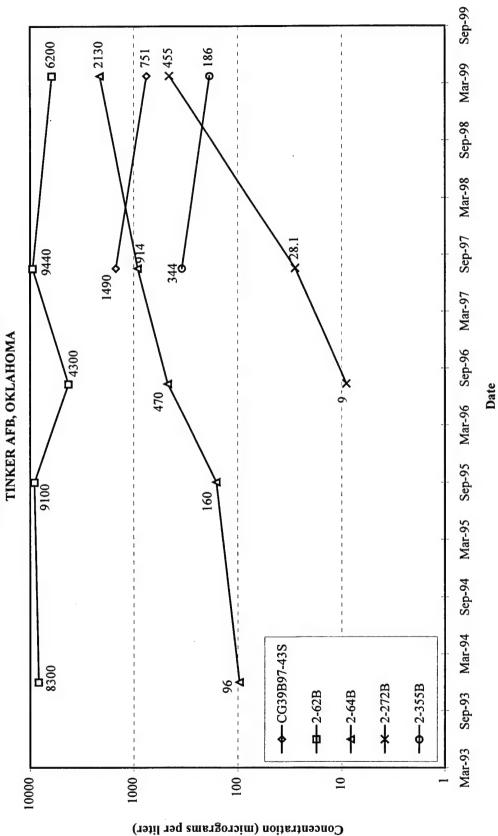
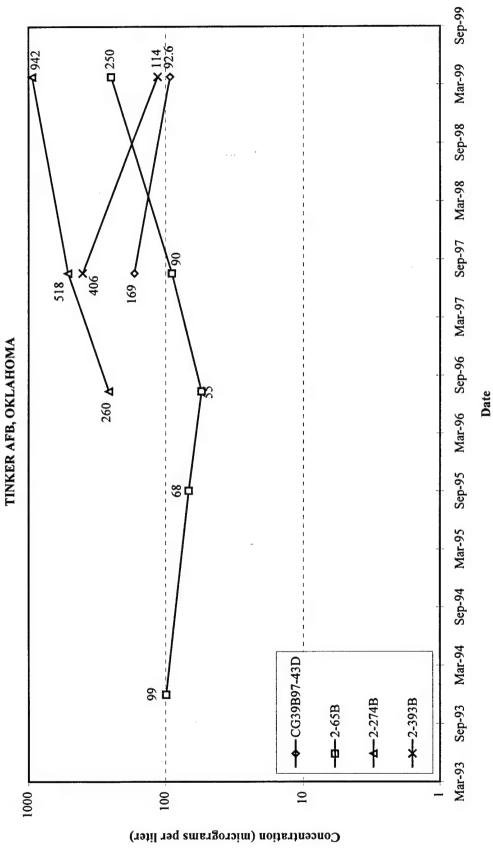


FIGURE 8

TCE CONCENTRATION VERSUS TIME FOR SELECTED WELLS

LOWER USZ SAND

SITE FTA-2 RNA TS ADDENDUM



TCE concentrations at well 2-274B suggest a secondary source near or upgradient of this location. Well 2-65B is located hydraulically up- and crossgradient from the FTA-2 fire pit location, along a former access road to the FTA-2 fire pit. Conversely, a significant decrease in TCE concentration occurred at well 2-393B (406 μ g/L to 114 μ g/L). The cause of the decrease in TCE concentration at this location is not readily apparent.

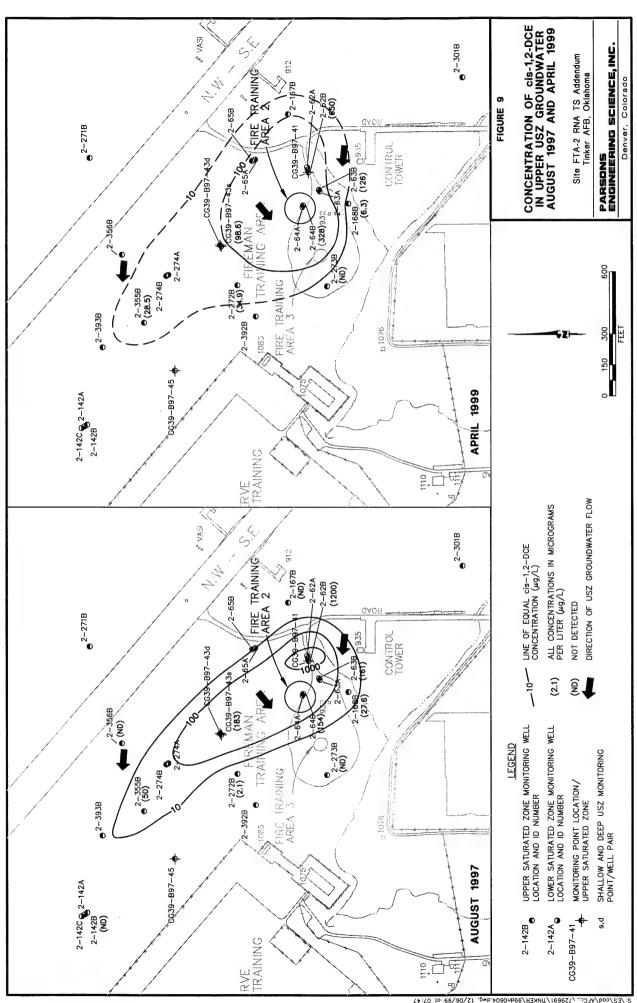
The TS presented a conceptual model where the primary source of TCE was near the FTA-2 fire pit and well 2-62B, where the TCE plume migrated toward the northwest due to a preferential flow path and anisotropy of the aquifer matrix. Alternately, Figures 5 and 6 suggest that the TCE plume may be migrating in the direction of maximum hydraulic gradient (west-southwest) from multiple sources. The presence of TCE "hotspots" at wells 2-62B, 2-65B, and 2-274B does not support a conceptual model of a single source and linear contaminant transport.

In any case, as contamination migrates downgradient it appears to be slowed by the presence of an area of low permeability near well 2-392B, which appears to trend northwest to southeast based upon the shape and location of the TCE plumes in the upper and lower USZ sands. The hydraulic conductivity measured at well 2-392B (Parsons ES, 1999) in 1997 was 1.1 ft/day, compared to an average hydraulic conductivity of 15 ft/day for wells measured to the east of well 2-392B (Section 2.1). Further discussion of this low permeability zone can be found in the TS (Parsons ES, 1998), Section 3.3 and cross-sections C-C', D-D', and E-E'.

cis-1,2-DCE

cis-1,2-DCE was detected in 11 of 14 wells sampled in the USZ during the April 1999 sampling event (Table 4). Concentrations of cis-1,2-DCE ranged up to 650 μg/L at well 2-62B. Concentrations of cis-1,2-DCE exceeded the USEPA MCL of 70 μg/L at six locations (CG39-B97-43s, 2-62B, 2-63B, 2-64B, 2-65B, and 2-274B). The two locations with the highest concentration of cis-1,2-DCE (2-62B and 2-64B) coincide with the highest TCE concentrations in the USZ. Figures 9 and 10 show the extent of cis-1,2-DCE in the upper and lower USZ sands, respectively.

cis-1,2-DCE concentrations in the upper USZ sand interval increased at two locations (2-64B and 2-272B), and decreased at five locations (CG39-B97-43s, 2-62B, 2-63B, 2-168B, and 2-355B). The contaminant plumes depicted in Figure 9 suggest that cis-1,2-DCE is migrating to the west-southwest in the direction of maximum hydraulic gradient, based on an increase in concentration of cis-1,2-DCE at well locations 2-64B and 2-272B. Concentrations of cis-1,2-DCE for the upper USZ sand over time are plotted for select monitoring wells in Figure 11. cis-1,2-DCE concentrations at 2-62B decreased from 1,200 μg/L to 650 μg/L from 1997 to 1999. Downgradient from 2-62B, well 2-64B shows an increase in cis-1,2-DCE concentration from 154 μg/L in 1997 to 328 μg/L in 1999, likely due to the downgradient transport of cis-1,2-DCE from a source area near 2-62B. The concentration of cis-1,2-DCE also increased at well 2-272B, from 2.1 μg/L in 1997 to 34.9 μg/L in 1999. Conversely, concentrations of cis-1,2-DCE decreased from August 1997 to April 1999 for wells CG39-B97-43s and 2-355B.



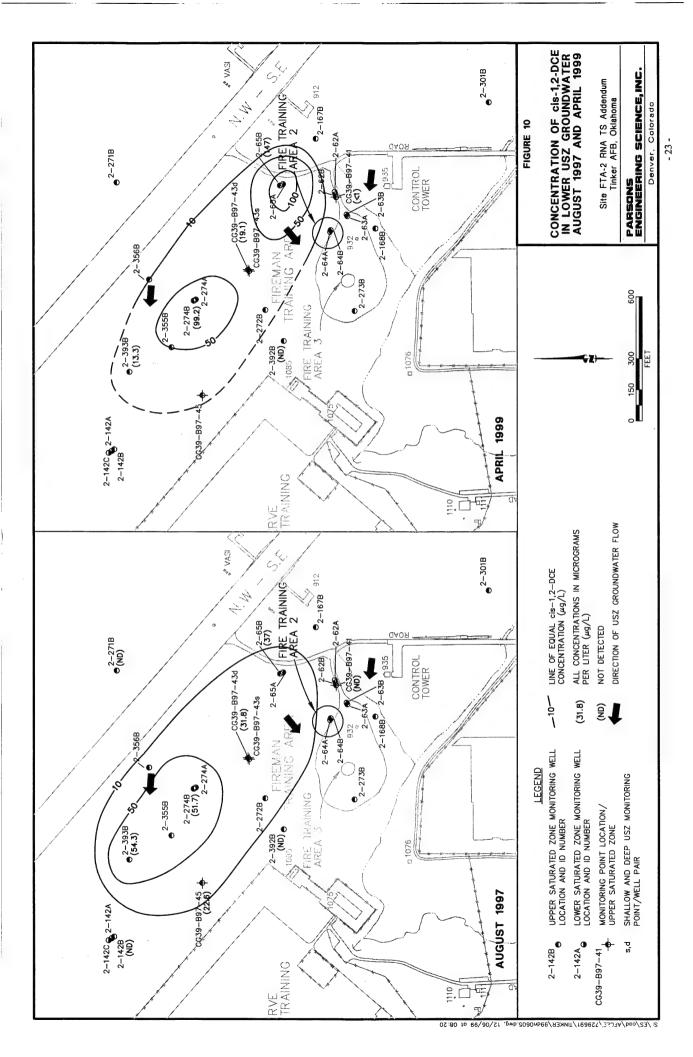
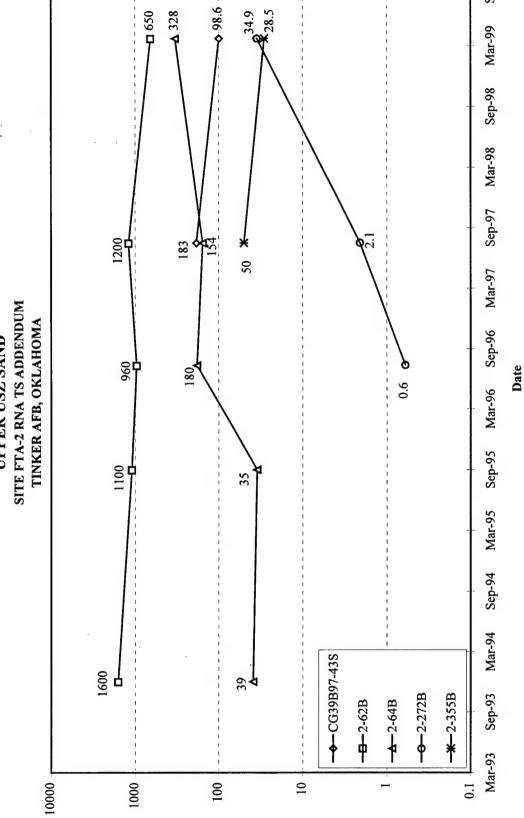


FIGURE 11
cis-1,2-DCE VERSUS TIME FOR SELECTED WELLS
UPPER USZ SAND



Concentration (micrograms per liter)

Sep-99

Concentrations of cis-1,2-DCE in the lower USZ sand interval (Figure 10) increased at two locations (2-65B and 2-274B) and decreased at two locations (CG39-B97-43d and 2-393B). Concentrations of cis-1,2-DCE versus time for these four wells are plotted on Figure 12. While the areal extent of cis-1,2-DCE plume from August 1997 to April 1999 is relatively stable, concentrations of cis-1,2-DCE increased significantly at well 2-65B (37 to 147 μ g/L) and at well 2-274B (51.7 to 99.2 μ g/L). The increase in concentration of cis-1,2-DCE at these two locations is similar to the increase in concentration of TCE in the lower USZ sand.

trans-1,2-DCE

Measurable concentrations of *trans*-1,2-DCE were detected at eight locations in 1999 at concentrations ranging up to 47.5 μg/L at well 2-62B. Concentrations of *trans*-1,2-DCE did not exceed the USEPA MCL of 100 μg/L. Locations with increases and decreases in concentrations of *trans*-1,2-DCE were similar to those for TCE and *cis*-1,2-DCE. Elevated concentrations of *trans*-1,2-DCE were encountered for wells 2-274B and 2-65B, screened in the lower USZ sand interval. This suggests downward vertical migration from the upper to lower USZ sand. Concentrations of *trans*-1,2-DCE were significantly lower than *cis*-1,2-DCE. This would be expected if TCE is being degraded to DCE via reductive dehalogenation, where under the influence of biodegradation *cis*-1,2-DCE is a more common daughter product than the other DCE isomers (USEPA, 1998). Furthermore, the *trans* isomer is more common when DCE is used for commercial applications.

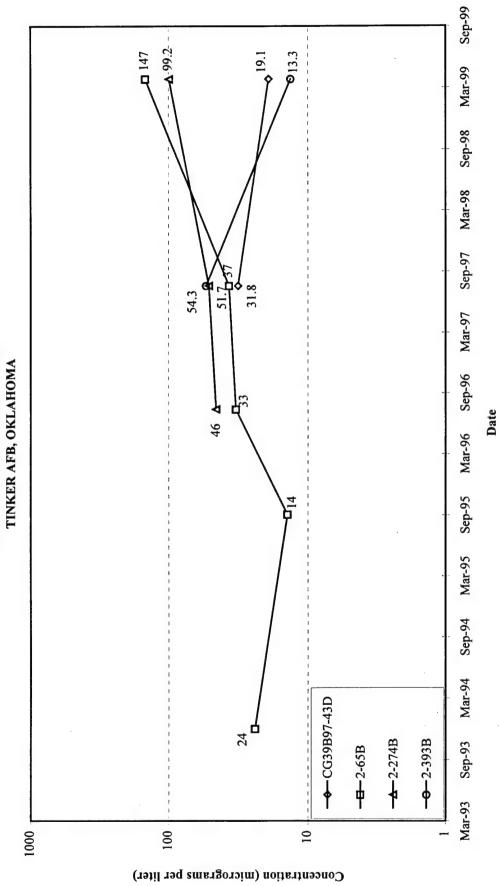
1,1-DCE

Measurable concentrations of 1,1-DCE were detected in samples from 3 of the 14 USZ monitoring wells sampled at FTA-2 in April 1999 (Table 4, Figure 4). Concentrations of 5.1 μ g/L, 1.2 μ g/L and 1.4 μ g/L were detected in samples from wells 2-62B, 2-64B, and 2-65B, respectively. Concentrations of 1,1-DCE did not exceed the USEPA MCL of 7 μ g/L. TCE and the other isomers of DCE also were detected in these wells suggesting that 1,1-DCE may have been produced by reductive dehalogenation of TCE.

Vinyl Chloride

VC was detected in samples from 2 of the 14 USZ groundwater monitoring wells sampled at FTA-2 in April 1999 (Table 4, Figure 4). Concentrations of 3.6 μ g/L and 5.1 μ g/L were detected in samples from wells 2-63B and 2-65B, respectively. Both concentrations exceed the USEPA MCL of 2 μ g/L for VC. The distribution of VC is similar to that for PCE, with VC detected at two of the four locations where PCE was detected. The presence of VC in groundwater is an indicator of reductive dehalogenation of DCE if VC was not released as a contaminant compound. Because VC is typically a gas at most working temperatures [the boiling point of VC is about -13.5 degrees Celsius (°C) or 7 degrees Fahrenheit (°F) (Montgomery, 1990)], VC is not typically found in chemical spills. VC is generally used for plastic manufacturing instead of being used as a solvent. In addition, when liquid VC is spilled, most of the mass is likely to volatilize before reaching groundwater unless there is a large-volume release.

FIGURE 12
cis-1,2-DCE VERSUS TIME FOR SELECTED WELLS
LOWER USZ SAND
SITE FTA-2 RNA TS ADDENDUM



3.2.2 Chlorinated Ethanes

Chlorinated ethanes detected in groundwater at FTA-2 include 1,1-dichloroethane (1,1-DCA) and 1,2-dichloroethane (1,2-DCA) (Table 4). 1,1-DCA was detected in 2 of the 14 USZ wells sampled in 1999. Relatively low concentrations of 1,1-DCA were detected, with the maximum 1,1-DCA concentration of 3.4 μ g/L occurring at well 2-63B, a decrease from 4.6 μ g/L in 1997. There is no USEPA MCL listed for 1,1-DCA.

The distribution of 1,2-DCA in USZ groundwater is shown on Figure 13, with measurable concentrations of 1,2-DCA detected in 5 of the 14 USZ groundwater samples analyzed in 1999. The maximum concentrations of 1,2-DCA in 1999 were 37.2 µg/L at 2-62B and 18.3 µg/L at 2-65B, both exceeding the USEPA MCL of 5 µg/L. These two locations are both in the same general locations as suspected sources of the CAH contaminant plume. 1,1-DCA is not commonly used as a solvent, whereas 1,2-DCA is used as a solvent. Therefore, the presence of 1,2-DCA rather than 1,1-DCA suggests that 1,2-DCA is a result of source release, rather than a biodegradation byproduct. Furthermore, 1997 data indicate the possibility of a secondary 1,2-DCA source located northwest of the FTA-2 fire pit in the vicinity of 2-393B (Figure 13).

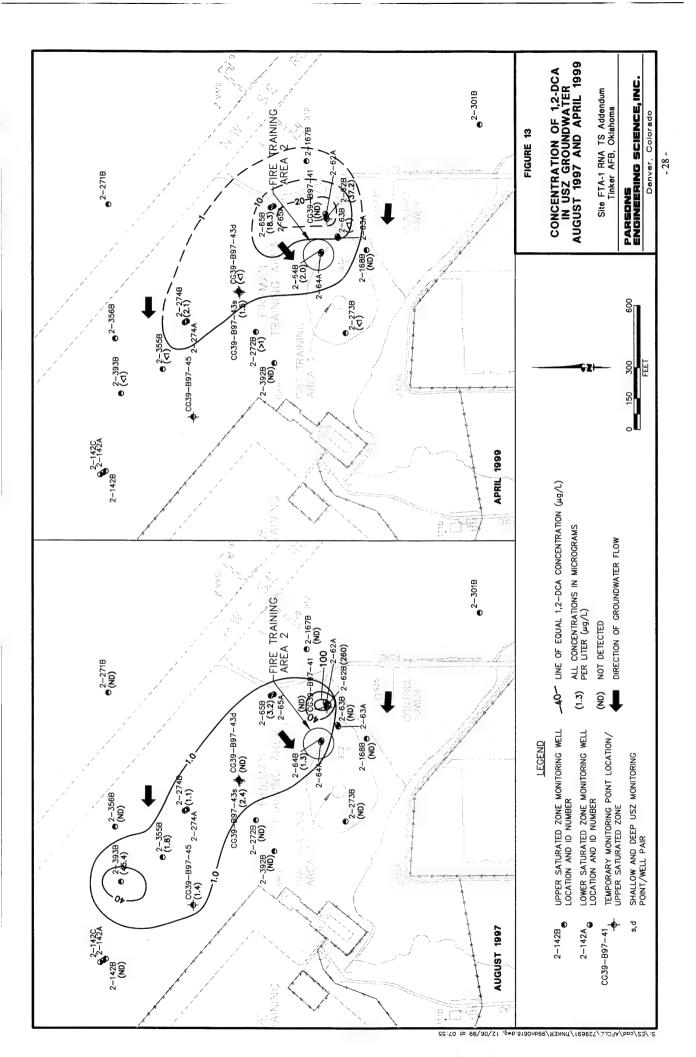
3.2.3 Chlorobenzenes

CBs detected in groundwater at FTA-2 include chlorobenzene (CB), 1,2-dichlorobenzene (1,2-DCB), 1,3-dichlorobenzene (1,3-DCB), and 1,4-dichlorobenzene (1,4-DCB) (Table 4). These compounds were detected in three wells associated within the general CAH source area(s) (2-62B, 2-64B, and 2-65B), and two wells along the apparent CAH plume axis (CG39-B9743s and 2-274B). In 1997, the CB and 1,2-DCB plume appeared to be originating from the vicinity of 2-62B, with migration toward the northwest. April 1999 data indicates that the plume has become less concentrated and reduced in size.

The concentration of CB at well 2-65B increased from 9.5 to 51.6 $\mu g/L$. CB was detected at a concentration of 19.5 $\mu g/L$ for well 2-62B, a decrease from the August 1997 concentration of 117 $\mu g/L$. Concentrations of CB did not exceed the USEPA MCL for CB of 100 $\mu g/L$. 1,2-DCB was detected in two wells. Well 2-62B exhibited a decrease in 1,2-DCB concentration from 969 $\mu g/L$ in August 1997 to 101 $\mu g/L$ in 1999. Well 2-65B had an increase in 1,2-DCB concentration from 3.6 $\mu g/L$ in 1997 to 23.8 $\mu g/L$ in 1999. Neither of these concentrations is above the USEPA MCL of 600 $\mu g/L$ for 1,2-DCB. Similarly, neither 1,3-DCB or 1,4-DCB were detected above their respective MCLs of 600 $\mu g/L$ and 75 $\mu g/L$.

3.2.4 Surface Water Quality

A single surface water sample was collected and analyzed for CAHs in 1999 from the tributary that flows northeast to southwest along the southern perimeter of the FTA-2 site (Figure 1). TCE was detected in the surface water sample at a concentration of 1.7 μ g/L, and cis-1,2-DCE at a concentration of 1.0 μ g/L. There are no background data for surface water flowing through the site along this tributary. However, due to the upward vertical gradient observed for well pair 2-62B and CG39-B97-41 in the source area near



the tributary, the detection of CAHs in surface water suggests discharge of contaminated groundwater to the tributary.

3.2.5 LSZ Groundwater Quality

Low concentrations of toluene, ethylbenzene, and m-xylene were detected in LSZ groundwater samples in April 1999. Toluene was detected at a concentration of 13.6 μ g/L at well location 2-64A, ethylbenzene at a concentration of 1.6 μ g/L at well 2-274A, and m-xylene at a concentration of 1.0 μ g/L at well 2-274A. These fuel hydrocarbons were not previously detected in LSZ groundwater in 1997, but the concentrations are below respective USEPA MCLs.

CAHs were not detected at measurable concentrations for any of the five wells sampled in the LSZ for the April 1999 sampling event (Table 4). In the August 1997 sampling event, TCE was detected at a concentration of 10.9 μ g/L at well 2-62A. The 1997 sampling event was the first time that CAHs were detected in LSZ groundwater at the site. In April 1999, TCE was not detected in well 2-62A above the reporting limit of 1.0 μ g/L.

Monitoring well 2-62A is located beneath the primary source area, immediately adjacent to well 2-62B, the USZ well with the highest detected concentration of TCE in USZ groundwater at FTA-2. TCE is the most mobile (lowest retardation) of the CAHs detected at FTA-2, and its detection in LSZ groundwater indicated that CAH contamination was starting to migrate vertically through the USZ/LSZ aquitard to the LSZ. However, the presence of an upward vertical hydraulic gradient within the USZ in this area may counteract the downward migration of TCE to the LSZ.

4.0 NATURAL ATTENUATION ANALYSIS

As discussed in the TS, available information indicates that the CAH plume primarily exhibits Type 1 behavior within a source area delimited by monitoring wells 2-63B, 2-168B, 2-62B, and 2-64B. Fuel hydrocarbons, while greatly decreased, and CBs remain in this area, and the electron donor supply may be adequate to allow continued microbial reduction of CAHs. The role of competing alternate electron acceptors (e.g., DO, nitrate, nitrate, ferric iron, and sulfate) is described later in this section. The Type 1 conditions result in significant degradation of the highly chlorinated solvents PCE and TCE in the source area. Degradation of DCE also may be occurring, as evidenced by the presence of low levels of VC dissolved in groundwater.

The downgradient (northern) portion of the CAH plume appears to primarily exhibit Type 3 behavior. There is little or no anaerobic reductive dehalogenation of TCE and DCE occurring; however, DCE may be aerobically degraded (oxidized). VC may be rapidly oxidized as it migrates into a more aerobic environment. The rapid decrease, or lack of, chlorobenzenes in the downgradient portions of the plume may indicate that CB and DCB are being utilized as electron donors in the source area of the CAH plume.

In addition to the distribution of contaminants and daughter products, other evidence can be used to support the interpretation that biodegradation of fuel hydrocarbons and chlorinated solvents is ongoing at FTA-2. Analysis of temporal changes in contaminant concentrations over time may indicate field-scale contaminant mass loss as a result of biodegradation processes. Changes in molar fractions of contaminants and daughter products also may indicate biodegradation is occurring. By comparing concentrations and distributions of electron donors, electron acceptors, and byproducts of microbially mediated reactions, types of biodegradation processes operating at a site can be identified. Geochemical parameters, including chloride, ORP, DO, and other changes in groundwater chemistry also can provide supporting evidence.

4.1 Field-Scale Contaminant Mass Loss

Groundwater quality data collected for five sampling events between December 1993 and April 1999 from monitoring wells 2-62B, 2-63B, 2-64B, and 2-65B were reviewed to assess temporal changes in contaminant concentrations over time (Parsons ES, 1999 and 1997). In the absence of source remediation, decreasing contaminant concentrations reflect natural attenuation trends. In a stable or attenuating contaminant plume, decreasing concentration trends may be used to support loss of contaminant mass through destructive processes such as biodegradation.

4.1.1 Chlorinated Ethenes Over Time

Available concentration-versus-time data for chlorinated ethenes in samples from source area wells 2-62B, 2-63B, 2-64B, and 2-65B were examined to assess trends in chlorinated ethene concentrations. Concentrations of TCE and *cis*-1,2-DCE versus time for wells 2-62B, 2-64B, and 2-65B are shown on Figures 7, 8, 11, and 12. Total chlorinated ethene concentrations for wells 2-62B, 2-63B, 2-64B, and 2-65B are shown on Figure 14.

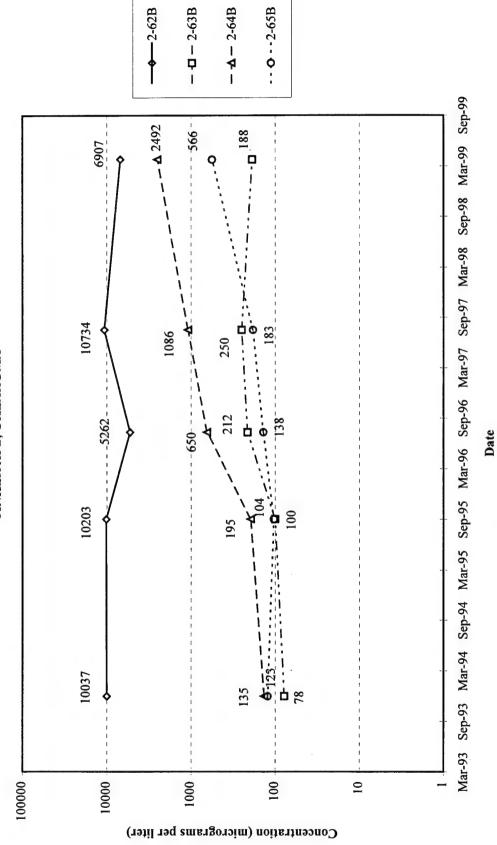
Total chlorinated ethene concentrations at these wells are relatively stable (wells 2-62B and 2-63B), or are gradually increasing over time (wells 2-64B and 2-65B). Well 2-62B has consistently exhibited the highest concentrations of CAHs at the site, and is assumed to represent dissolved contaminant concentrations in the primary source area. Concentrations at well 2-62B do not show a discernible trend indicating that the primary source of chlorinated ethenes at FTA-2 has not attenuated significantly. Well 2-64B is located in the former FTA-2 pit and concentrations at this well have shown a steady increase. This well is located along the migration pathway from the suspected source area, and increasing concentration levels could result from migration from the primary CAH source.

Well 2-65B is located upgradient from the suspected FTA-2 source area along the eastern edge of the contaminant plume. Total chlorinated ethene concentrations at this well have steadily increased with time, suggesting a secondary source of CAHs. Given the available data, it appears that the plume is expanding both upgradient and downgradient from the suspected primary FTA-2 source, based on increased concentration data for wells 2-63B, 2-64B, and 2-65B.

4.1.2 Chlorinated Ethanes Over Time

Concentration-versus-time data for total chlorinated ethanes detected in samples from wells 2-62B, 2-63B, 2-64B, and 2-65B were examined to assess trends in chlorinated

FIGURE 14
TOTAL CHLORINATED ETHENES VERSUS TIME
SITE FTA-2 RNA TS ADDENDUM
TINKER AFB, OKLAHOMA



ethanes (Figure 15). Concentrations of total chlorinated ethanes decrease through time at well 2-62B, while total chlorinated ethane concentrations increase through time at wells 2-63B, 2-64B, and 2-65B. A significant decrease in total chlorinated ethanes occurred at well 2-62B, from 509 μ g/L in 1993 to 37 μ g/L in 1999. Concentrations of total chlorinated ethanes increased over time for wells 2-63B and 2-64B, but only at low concentrations (less than 5 μ g/L). Results of this analysis indicate an attenuation of total chlorinated ethanes at well 2-62B, with some migration or dispersion of chlorinated ethanes near the primary source in the vicinity of wells 2-63B and 2-64B.

Concentrations of total chlorinated ethanes increased from non-detect to 20 μ g/L at upgradient well 2-65B, similar to increases in chlorinated ethenes and chlorobenzenes at this location. These data support the presence of a secondary source near well 2-65B.

4.1.3 Chlorobenzenes Over Time

Available concentration-versus-time data for CBs in samples from source area wells 2-62B, 2-64B, and 2-65B are shown on Figure 16. Total CB concentrations gradually increased over time for wells 2-64B and 2-65B, and decreased through time at well 2-62B. Similar to total chlorinated ethanes, an overall decrease in CBs occurs for well 2-62B, with low level migration or dispersion of CBs downgradient of the primary source area (well 2-64B). There also is an increase in total chlorobenzenes at well 2-65B, from non-detect in 1993 to 86.4 µg/L in 1999.

4.2 Presence of Daughter Products and CAH Molar Fractions

Results of the TS indicated that the presence of daughter products cis-1,2-DCE and VC provides strong evidence that parent solvents (TCE and/or PCE) are being reductively dehalogenated. Anaerobic condition conducive to reductive dehalogenation are limited to the primary source area. Concentrations of VC present at the site are low and suggest that if VC is produced by reductive dehalogenation, it is quickly degraded as it is transported downgradient into an aerobic environment. The geochemical conditions conducive to the aerobic degradation of VC do not at the same time support anaerobic reductive dechlorination of TCE or DCE. However, DCE may also be aerobically degraded, but at a rate much slower than for VC.

Changes in the molar fraction of the chlorinated ethenes can provide an indication that higher chlorinated ethenes are being degraded to lesser chlorinate ethenes (i.e., PCE to TCE to DCE to VC to ethene). Figure 17 shows molar fractions of chlorinated ethenes at well 2-62B over time. Only low levels of PCE or VC were detected at this location (Table 4), while ethene has not been detected at the site. The ratio of TCE to DCE at well 2-62B has remained relatively constant over time. While the data suggest that some TCE has degraded by reductive dehalogenation to DCE, the lack of VC indicates that the process does not occur past the initial step of TCE to DCE. Furthermore, the persistence of TCE indicates that either TCE has not continued to degrade to DCE, or that the rate of degradation of TCE to DCE is offset by the rate of mass loading to groundwater from a continuing TCE source. Figure 18 shows the molar fractions of chlorinated ethenes along the apparent CAH plume axis for August 1997. The relatively constant ratio of TCE to DCE indicates little, if any, degradation of TCE to DCE along the plume axis. Therefore,

FIGURE 15
TOTAL CHLORINATED ETHANES VERSUS TIME
SITE FTA-2 RNA TS ADDENDUM
TINKER AFB, OKLAHOMA

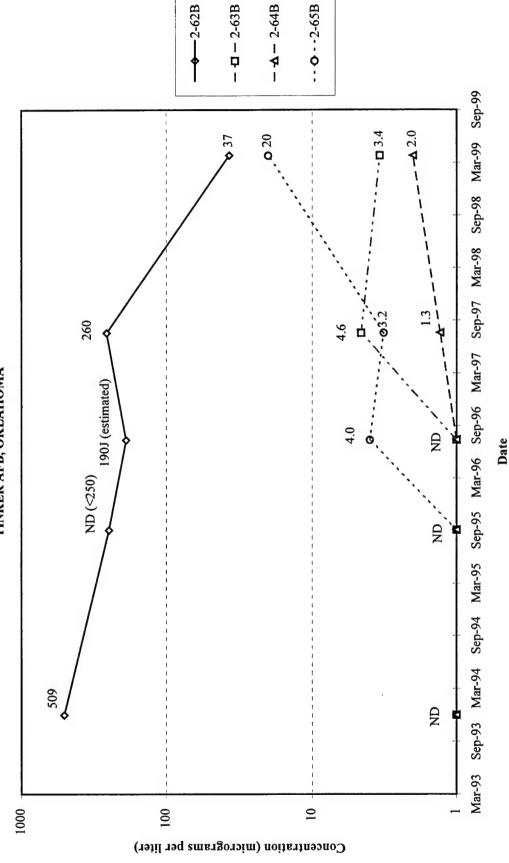
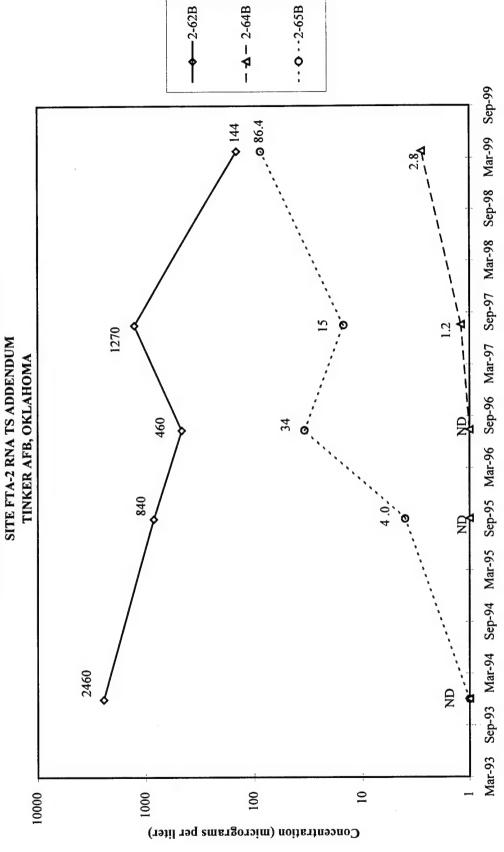
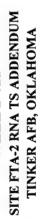


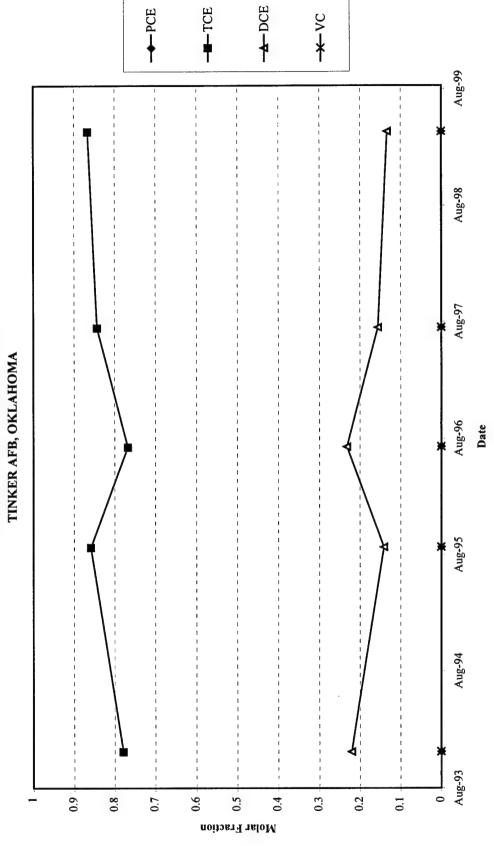
FIGURE 16
TOTAL CHLOROBENZENES VERSUS TIME
SITE FTA-2 RNA TS ADDENDUM



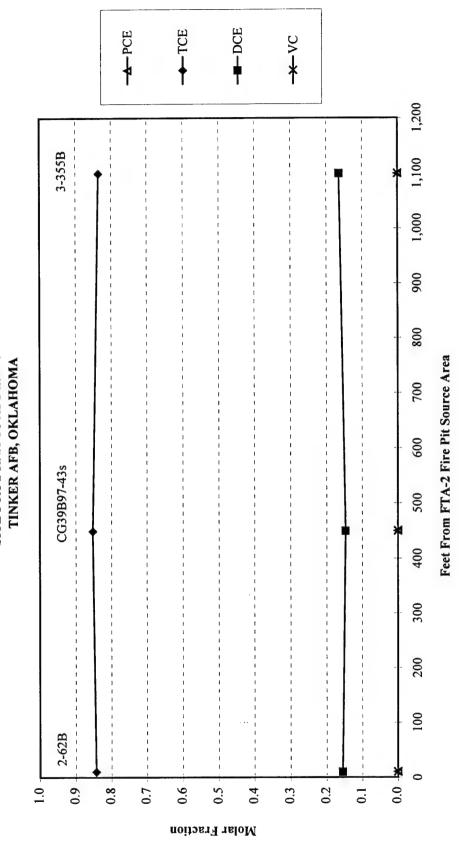
Date

MOLAR FRACTION OF CHLORINATED ETHENES OVER TIME **WELL 2-62B** FIGURE 17





MOLAR FRACTION OF CHLORINATED ETHENES VERSUS DISTANCE SITE FTA-2 RNA TS ADDENDUM AUGUST 1997 FIGURE 18



while elevated concentrations of DCE indicate some degradation of TCE to DCE has occurred, there is insufficient data to indicate this process continues or occurs outside of the primary source area.

Increasing concentrations of CAHs at upgradient well 2-65B indicate the presence of a secondary source. Molar fractions of chlorinated ethenes over time at well 2-65B are shown on Figure 19. An overall decrease in the fraction of TCE, with an increase in fraction of DCE and the presence of VC indicates that TCE is degrading to DCE, and DCE to VC. DO concentrations measured at this location in 1999 support the presence of sufficiently reducing conditions for reductive dehalogenation (Section 4.6.1). Conversely, the ORP measured at well 2-65B in 1999 does not support sufficiently reducing conditions (Section 4.5). Increasing concentrations of PCE and TCE at this location indicate a continuing source of chlorinated solvents. Because well 2-65B is screened in the lower USZ sand interval, it is possible that concentrations at this location may reflect degradation at an upgradient or more shallow (upper) USZ sand interval.

4.3 Electron Donors

The distribution of potential electron donors (carbon sources) such as BTEX or native organic carbon are useful for evaluating the feasibility of reductive dehalogenation of CAHs. The coincidence of dissolved BTEX with the highest concentrations of the reductive dehalogenation daughter products *cis*-1,2-DCE and VC is considered a favorable factor in the potential biodegradation of CAHs by reductive dehalogenation in the general source area. However, BTEX and fuel carbon concentrations are much lower than CAH concentrations (Tables 3 and 4), and the extent and concentrations of the BTEX/fuel carbon plumes are such that these compounds have not significantly migrated beyond the general source area of the contaminant plumes.

Dissolved organic carbon can also act as a source of electron donors during the reductive dehalogenation of CAHs. TOC concentrations in excess of 20 milligrams per liter (mg/L) are desirable to drive dehalogenation reactions (USEPA, 1998). However, TOC concentrations detected in shallow groundwater at FTA-2 in August 1997 were less than 12 mg/L. Native organic carbon concentrations of this magnitude are insufficient to drive reductive dehalogenation of CAHs.

4.4 Chloride as an Indicator of Dehalogenation

Chloride ions are removed from chlorinated solvents and enter solution during biodegradation, whether via reductive dehalogenation or aerobic oxidation. Therefore, chloride concentrations in groundwater may increase above background levels in areas where reductive dehalogenation or aerobic biodegradation are taking place. Chloride concentrations measured in August 1997 and April 1999 are presented in Table 5.

Chloride concentrations within the boundaries of the chlorinated solvent plume were slightly elevated above background in August 1997, with a maximum concentration of 187 mg/L at 2-355B (Parsons ES, 1999). When 1999 chloride concentrations are compared with 1997 concentrations, all locations sampled within the boundaries of the CAH plume have increased concentrations with the exception of locations 2-65B and

FIGURE 19
MOLAR FRACTION OF CHLORINATED ETHENES OVER TIME
WELL 2-65B



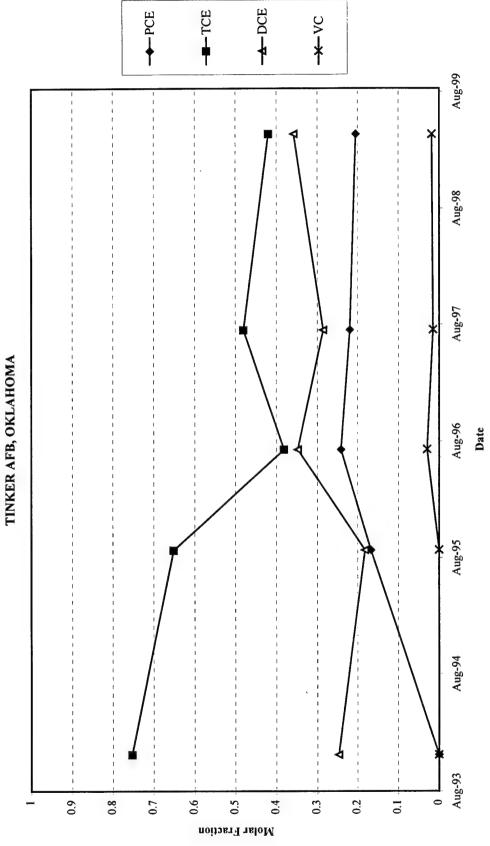


TABLE 5 GROUNDWATER GEOCHEMICAL DATA SITE FTA-2 RNA TS ADDENDUM TINKER AFB, OKLAHOMA

										,										
					Dissolved	Redox	Total			Ferrous	Nitrite+				Total	Carbon				
Sample	Sample	Temp		Conductivity	Oxygen	Potential	Alkalinity	Chloride	Ammonia	Iron	Nitrate	Sulfate	Sulfide	Manganese	Phenols	Dioxide	T0C	Methane	Ethane	Ethene
Location	Date	(°C)*/	_	(μs/cm) ^{c/}	(mg/L) ^{d/}	(mV) ^{e/}	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)
Upper Saturated Zone	d Zone																			
CG39-B97-41	Aug-97	22.0	88.9	861	3.9	136	520	5.73	<0.05	< 0.01	2.01	15.4	< 0.01	< 0.01	< 0.01	9	<0.4	<0.001	< 0.003	< 0.002
	Apr-99	17.9	7.12	848	5.4	315	380	7.17	<0.10	< 0.1	3.00	16.5	<0.1	NΑ ⁰	NA	NA	NA	< 0.001	< 0.003	< 0.002
CG39-B97-43s	Aug-97	21.5	6.80	1,280	4.8	117	330	159	< 0.05	< 0.01	4.22	49.0	< 0.01	< 0.01	0.2	48	0.539	< 0.001	< 0.003	< 0.002
	Apr-99	17.7	7.00	1,290	3.5	24	280	160	< 0.10	< 0.1	4.31	46.0	<0.1	Ϋ́	NA	NA	NA	<0.001	< 0.003	<0.002
CG39-B97-43d	Aug-97	18.4	7.07	1,110	3.3	130	450	52.0	< 0.05	< 0.01	2.55	44.1	<0.01	< 0.01	0.2	75	0.545	< 0.001	< 0.003	< 0.002
	Apr-99	17.7	7.10	1,090	3.1	313	400	55.7	<0.10	<0.1	3.60	38.3	<0.1	ΝΑ	NA	NA	NA	<0.001	<0.003	<0.002
CG39-B97-45	Aug-97	19.4	6.91	872	6.5	103	380	46.9	< 0.05	< 0.01	4.80	11.3	<0.01	< 0.01	<0.01	75	<0.4	<0.001	< 0.003	<0.002
2-62B	Aug-97	18.8	7.34	1,440	0.2	51	260	173	<0.05	< 0.01	3.92	127	< 0.01	< 0.01	< 0.01	55	3.30	1.05	< 0.003	<0.002
	Apr-99	18.2	7.20	1,440	1.2	20	360	176	< 0.10	0.1	5.52	111	<0.1	NA	NA	NA	NA	0.122	< 0.003	<0.002
2-63B	Aug-97	19.9	7.35	1,520	9.0	42	460	94.4	< 0.05	< 0.01	2.39	65.7	< 0.01	< 0.01	< 0.01	75	11.2	0.002	< 0.003	<0.002
	Apr-99	19.4	7.10	1,690	0.2	244	009	112	< 0.10	< 0.1	3.09	84.2	<0.1	. NA	NA	NA	NA	0.001	< 0.003	<0.002
2-64B	Aug-97	20.6	NA	848	1.8	124	260	49.3	<0.05	< 0.01	3.42	27.0	<0.01	< 0.01	< 0.01	40	0.795	< 0.001	< 0.003	<0.002
	Apr-99	19.8	7.30	921	1.0	177	360	53.3	<0.10	< 0.1	4.27	26.3	<0.1	NA	NA	NA	NA	<0.001	< 0.003	< 0.002
2-65B	Aug-97	18.8	7.13	1,670	2.3	91	350	78.5	< 0.05	< 0.01	3.18	44.4	< 0.01	< 0.01	<0.01	75	1.96	0.002	< 0.003	< 0.002
	Apr-99	17.5	7.10	749	0.2	347	400	56.4	<0.10	< 0.1	0.52	23.8	<0.1	NA	NA	NA	NA	0.010	< 0.003	< 0.002
2-167B	Aug-97	17.3	7.11	1,010	2.3	81	999	4.49	< 0.05	0.1	0.27	18.1	< 0.01	< 0.01	0.1	75	0.488	0.022	< 0.003	< 0.002
2-168B	Aug-97	18.3	7.17	993	0.7	-22	380	14.4	< 0.05	< 0.01	0.58	42.4	< 0.01	< 0.01	< 0.01	65	3.01	0.02	< 0.003	< 0.002
	Apr-99	17.2	9.30	1,170	6.0	283	520	3.24	< 0.10	<0.1	0.41	36.1	<0.1	NA	NA	NA	NA	< 0.001	< 0.003	< 0.002
2-271B	Aug-97	16.9	7.19	1,240	3.8	89	420	117	0.15	9.0	1.11	29.7	< 0.01	< 0.01	0.3	65	0.939	<0.001	< 0.003	< 0.002
2-272B	Aug-97	17.8	7.24	1,080	4.1	30	280	130	< 0.05	< 0.01	1.95	37.6	<0.01	< 0.01	< 0.01	45	0.448	< 0.001	< 0.003	< 0.002
	Apr-99	18.9	7.20	942	4.7	185	300	106	< 0.10	<0.1	2.83	28.9	<0.1	NA	NA	NA	NA	< 0.001	< 0.003	< 0.002
2-273B	Aug-97	18.3	7.20	761	0.2	-105	280	16.5	< 0.05	0.2	0.35	12.1	< 0.01	< 0.01	< 0.01	45	0.638	0.005	< 0.003	< 0.002
	Apr-99	19.3	7.20	477	2.0	277	320	18.6	<0.10	<0.1	09.0	9.75	<0.1	NA	NA	NA	NA	<0.001	< 0.003	< 0.002
2-274B	Aug-97	17.7	7.20	919	5.1	20	310	74.6	< 0.05	0.2	3.08	24	< 0.01	< 0.01	0.2	45	<0.4	< 0.001	< 0.003	< 0.002
	Apr-99	18.6	7.40	970	4.4	247	380	82.6	<0.10	<0.1	4.32	27.2	<0.1	NA	NA	NA	NA	< 0.001	< 0.003	<0.002
2-355B	Aug-97	17.8	7.23	1,490	7.9	117	380	187	<0.05	0.4	3.28	48.8	< 0.01	< 0.01	0.3	20	0.741	< 0.001	< 0.003	< 0.002
	Apr-99	19.0	7.00	1,320	6.1	353	340	202	< 0.10	0.1	3.32	15.5	<0.1	NA	NA	NA	NA	< 0.001	< 0.003	< 0.002
2-392B	Aug-97	18.5	7.18	269	0.9	139	350	9.01	< 0.05	< 0.01	2.57	5.91	<0.01	< 0.01	0.2	09	<0.4	< 0.001	< 0.003	< 0.002
	Apr-99	17.4	06.6	236	5.5	151	100	0.64	< 0.10	<0.1	0.43	6.39	<0.1	NA	NA	NA	NA	< 0.001	< 0.003	< 0.002
2-393B	Aug-97	18.2	7.23	1,180	7.0	111	190	148	< 0.05	< 0.01	3.70	14.1	< 0.01	< 0.01	< 0.01	110	<0.4	< 0.001	< 0.003	< 0.002
	Apr-99	17.5	7.20	620	4.7	270	320	187	< 0.10	<0.1	0.43	13.9	<0.1	NA	NA	NA	NA	<0.001	< 0.003	<0.002
2-142B	Aug-97	6.91	7.36	1,480	3.1	53	380	218	<0.05	8.0	2.00	76.1	<0.01	<0.01	0.2	55	0.929	< 0.001	<0.003	< 0.002

- 39 -

GROUNDWATER GEOCHEMICAL DATA SITE FTA-2 RNA TS ADDENDUM TINKER AFB, OKLAHOMA TABLE 5

					Dissolved	Redox	Total			Ferrous	Nitrite+	,			Total	Carbon				
Sample	Sample	Тетр	Hd	Conductivity	Oxygen	Potential	Alkalinity	Chloride	Ammonia	Iron	Nitrate	Sulfate	Sulfide	Manganese	Phenols	Dioxide	TOC	Methane	Ethane	Ethene
Location	Date	$^{p}(C)^{pq}$	(sn) _{p/}	(μs/cm) ^{c/}	(mg/L) ^{d/}	(mV) ^{e/}	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)
Lower Saturated Zone	ted Zone																			
2-62A	Aug-97	21.5	7.42	782	5.6	6	310	7.54	< 0.05	0.1	4.02	6.51	< 0.01	< 0.01	< 0.01	45	1.53	< 0.001	< 0.003	<0.002
	Apr-99	NA	Ϋ́	NA	NA A	NA	360	5.55	< 0.10	0.1	8.4	3.43	< 0.1	NA	NA	NA	NA	< 0.001	< 0.003	< 0.002
2-63A	Aug-97	18.9	7.31	654	7.2	48	380	20.5	< 0.05	1.0	1.91	8.46	< 0.01	< 0.01	< 0.01	20	0.745	< 0.001	< 0.003	< 0.002
	Apr-99	NA	A V	NA	Y'A	NA.	Ϋ́	NA	NA	NA	Y'A	NA	NA	Ϋ́Z	NA	NA	NA	NA	NA	NA
2-64A	Aug-97	20.2	69.7	999	6.1	122	270	21.5	<0.05	0.1	1.29	7.96	< 0.01	< 0.01	< 0.01	45	0.152	< 0.001	< 0.003	< 0.002
	Apr-99	Y Y	Y Y	NA	Y Z	AN	NA	NA	NA	NA	Ϋ́Z	NA	NA	NA	NA	NA	NA	NA	NA	NA
2-65A	Aug-97	19.9	79.7	576	4.2	-38	190	25.6	< 0.05	0.3	0.72	68.6	< 0.01	< 0.01	<0.01	25	<0.4	< 0.001	< 0.003	< 0.002
	Apr-99	NA A	NA A	NA	NA V	NA	NA	NA	NA	NA	ΝΑ	NA	NA	NA	NA	NA	NA	NA	NA	NA
2-274A	Aug-97	20.1	7.60	516	7.4	63	210	18.0	<0.05	< 0.01	0.62	10.4	< 0.01	< 0.01	0.4	25	<0.4	< 0.001	< 0.003	< 0.002
	Apr-99	19.4	8.00	382	3.4	264	200	10.7	< 0.10	< 0.1	0.80	10.1	< 0.1	NA	NA	NA	NA	< 0.001	< 0.003	< 0.002
2-142A	Aug-97	17.5	7.75	620	1.2	26	270	28.4	< 0.05	< 0.01	< 0.05	9.72	< 0.01	< 0.01	0.2	45	2.14	< 0.001	< 0.003	< 0.002
a Or - degrees Centionade	ve Centiorade																			

 $^{a'}$ $^{\circ}$ C = degrees Centigrade. $^{b'}$ su = standard pH units.

^{σ/} μs/cm = microsiemens per centimeter.

^{d/} mg/L = milligrams per liter.

^{e/} mV = millivolts.

^{f/} NA = not analyzed.

2-272B. Wells 2-271B and 2-148B had anomalous elevated concentrations of chloride in 1997, but were not sampled in 1999. The elevated chloride concentrations observed at FTA-2 are a moderate indication that CAHs are undergoing microbially mediated degradation when reviewed in conjunction with the distributions of CAHs.

4.5 ORP as an Indicator of Redox Processes

As discussed in the TS, redox potential is a measure of the relative tendency of a solution to accept or transfer electrons. The redox potential of a groundwater system depends on which electron acceptors are being reduced by microbes during oxidation of organic compounds. Redox potentials were measured at 15 groundwater monitoring wells in April 1999 (Table 5 and Figure 20). Redox potentials for USZ groundwater at the site ranged from -105 millivolts (mV) to 139 mV in 1997, and from 24 mV to 353 mV in April 1999. The 1999 results indicate a significant increase in ORP at the site. However, ORP is very sensitive to methods of field measurement and sampling technique. Therefore, additional ORP data are required to assess whether this is an actual trend, or a result of sampling error.

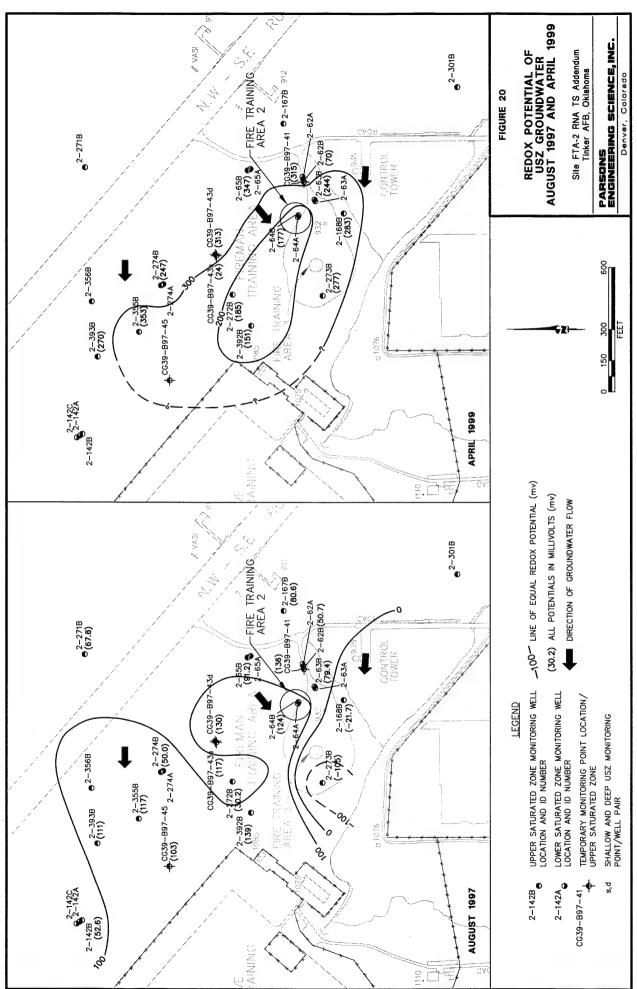
In many locations, relatively lower redox potentials continue to coincide with areas of elevated solvent concentrations; decreased DO, nitrate, and sulfate concentrations; and elevated ferrous iron and methane concentrations. As noted in the TS, this is not always the case and the correlation between redox potential and geochemical indicators is not strong.

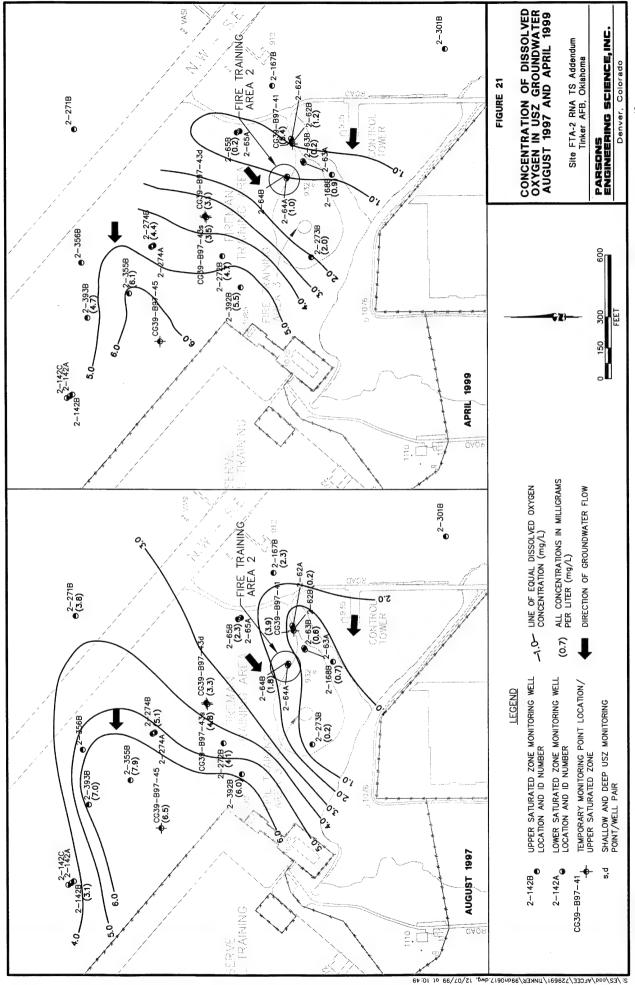
4.6 Alternate Electron Acceptors and Metabolic Byproducts

As discussed in the TS, microorganisms obtain energy for cell production and maintenance by facilitating thermodynamically advantageous oxidation/reduction reactions involving the transfer of electrons from electron donors to available electron acceptors. This results in the oxidation of the electron donor and the reduction of the electron acceptor. Electron acceptors are elements or compounds that occur in relatively oxidized stated and include DO, nitrate, manganese, ferric iron, sulfate, and carbon dioxide. DO is used first as the prime electron acceptor, after which anaerobic microorganisms use electron acceptors in the following order of preference: nitrate, manganese, ferric iron, sulfate, and finally carbon dioxide. Concentrations of compounds used as electron acceptors are depleted, and byproducts of electron acceptor reduction (e.g., ferrous iron, methane, and sulfide) are enhanced. By measuring these changes, it is possible to evaluate the importance of natural attenuation at a site. The following paragraphs discuss those parameters most useful in evaluating site biodegradation processes.

4.6.1 Dissolved Oxygen

DO concentrations were measured during the August 1997 and April 1999 sampling events (Table 5). Concentrations ranged from 0.2 mg/L to 7.9 mg/L in 1997, and from 0.2 to 6.1 mg/L in 1999. Within and immediately downgradient of the suspected FTA-2 source area (well location 2-62B) concentrations were below 2.0 mg/L in both 1997 and





1999 (Figure 21). Concentrations of DO at well location 2-65B also are depleted, decreasing from 2.3 mg/L in 1997 to 0.2 mg/L in 1999.

Depletion of DO within the CAH source area relative to upgradient and crossgradient concentrations at FTA-2 indicates that microbial activity stimulated by the presence of anthropogenic hydrocarbons (e.g., fuel hydrocarbons) is consuming DO and creating anaerobic conditions. However, outside of the suspected source areas, DO concentrations gradually increase from less than 1.0 mg/L to more than 6.0 mg/L in a north-westerly direction. Therefore, aerobic conditions exist for the majority of the CAH plume.

4.6.2 Nitrate + Nitrite

Concentrations of nitrate + nitrite (as N) were measured for groundwater samples collected in August 1997 and April 1999 (Table 5). Nitrate + nitrite concentrations generally range from less than 1 to as high as 5.52 mg/L (well 2-62B in 1999). There is no apparent correlation between depleted nitrate + nitrite and elevated CAH concentrations. However, concentrations of nitrate + nitrite did decrease at well 2-65B from 3.18 mg/L in 1997 to 0.52 mg/L in 1999, corresponding to an increase in CAHs at that location.

4.6.3 Ferrous Iron

Ferrous iron concentrations measured in groundwater samples collected in August 1997 and April 1999 are summarized in Table 5. Concentrations of ferrous iron in the USZ ranged from less than 0.1 mg/L to a maximum of 0.8 mg/L (well 2-142B in 1997). Elevated concentrations of ferrous iron were not detected at the site in the area of the contaminant plume. This suggests that ferric iron hydroxide is not being reduced to soluble ferrous iron during biodegradation of native organic carbon, fuel hydrocarbons, or less-chlorinated solvents. It is therefore unlikely that ferric iron reduction is an ongoing anaerobic biodegradation process at FTA-2.

4.6.4 Sulfate

Sulfate concentrations in USZ groundwater at the site ranged from 5.91 mg/L to 127 mg/L in 1997, and from 6.39 mg/L to 111 mg/L in 1999 (Table 5). The highest concentration of sulfate was detected in well 2-62B in both 1997 and 1999. Areas of lower sulfate concentrations do not correlate with elevated CAHs. It is therefore unlikely that sulfate reduction is an ongoing anaerobic biodegradation process at FTA-2.

4.6.5 Methane, Ethane and Ethene

Methane, ethane, and ethene concentrations measured in groundwater samples collected in August 1997 and April 1999 are listed on Table 5. Methane was detected at six locations in 1997 (2-62B, 2-63B, 2-65B, 2-167B, 2-168B, and 2-73B), and at three locations in 1999 (2-62B, 2-63B, and 2-65B). Location 2-167B was not sampled in 1999. Methane concentrations ranged up to 1.05 mg/L at well 2-62B in 1997, and up to 0.122 mg/L at well 2-62B in 1999. Ethane and ethene were not detected in either 1997 or 1999. The area of elevated methane concentrations coincides with suspected fuel hydrocarbon and chlorinated solvent source areas near wells 2-62B and 2-65B. The

presence of methane within these suspected source areas indicates highly reducing conditions favorable for oxidation of fuel hydrocarbons and native organic matter, as well as reductive dehalogenation of chlorinated solvents.

4.7 Other Geochemical Parameters

4.7.1 Ammonia

The presence of ammonia in groundwater can result from either nitrate reduction or fixing of atmospheric nitrogen. Because fixation of atmospheric nitrogen only occurs under reducing conditions [ORP less than -500 mV (Stumm and Morgan, 1981)], ammonia production via nitrate reduction is probably more common than by fixation of atmospheric nitrogen. In either case, the presence of ammonia in groundwater is a strong indication of microbial activity. Ammonia was not detected at any locations in the USZ groundwater at a concentration greater than the detection limit of 0.10 mg/L for samples collected in April 1999 (Table 5). Because nitrate is not widespread in USZ groundwater or present at any significant concentrations, it is unlikely that biodegradation by nitrate reduction is occurring at the site. Therefore, ammonia concentrations are not likely to be elevated at FTA-2.

4.7.2 Alkalinity

Total alkalinity (as calcium carbonate) was measured in groundwater samples collected in August 1997 and April 1999 (Table 5). Alkalinity is a measure of the ability of groundwater to buffer changes in pH. Total alkalinity in USZ groundwater at the site varied from 190 mg/L to 560 mg/L in 1997, and from 100 mg/L to 600 mg/L in 1999. Alkalinity within the plume area remains highly variable, but appears to be sufficient to buffer potential changes in pH caused by biologically mediated reactions.

4.8 Summary of Biodegradation of Chlorinated Solvents

Concentration versus time data indicate that the source(s) of CAHs at FTA-2 have not significantly attenuated. Concentrations of chlorinated ethenes are stable at well 2-62B, but increasing at wells 2-64B and 2-65B. This suggests downgradient migration of contaminants from the primary source area near well 2-62B to well 2-64B, and the presence of a secondary source near well 2-65B (located upgradient from well 2-62B). Concentrations of chlorinated ethanes and chlorobenzenes are likewise increasing at well locations 2-64B and 2-65B, but do appear to be gradually decreasing at well 2-62B.

The presence of daughter products cis-1,2-DCE and VC provides strong evidence that parent solvents (TCE and/or PCE) are being reductively dehalogenated. While the data suggest that TCE has been degraded by reductive dehalogenation to DCE, a relative lack of VC and ethene indicates that the process does not significantly proceed past the initial step of TCE to DCE. The persistence and strength of TCE in groundwater indicates that either TCE has not continued to degrade to DCE, or that the rate of degradation of TCE to DCE is offset by the rate of mass loading to groundwater from a continuing TCE source. Molar fractions of chlorinated ethenes along the apparent CAH plume axis indicates little, if any, degradation of TCE to DCE outside the immediate source area(s). While elevated concentrations of DCE indicate some degradation of TCE to DCE has occurred, there is

insufficient data to indicate this process continues or occurs outside of the primary source area. At well 2-65B, an overall decrease in the molar fraction of TCE, with an increase in fraction of DCE and the presence of VC indicates that TCE is degrading to DCE, and DCE to VC.

Geochemical data indicate that DO is being consumed and methane is being produced in the primary and secondary source areas near wells 2-62B and 2-65B. Nitrate reduction, ferric iron reduction, and sulfate reduction do not appear to be significant anaerobic degradation processes at the site. Away from the source areas, DO concentrations are aerobic (>1.0 mg/L) and measured redox potentials do not indicate that conditions are in the ranges that produce a sufficient environment for the reduction of CAH compounds.

5.0 CONCLUSIONS AND RECOMMENDATIONS

The FTA-2 fire pit location was identified as a potential source of fuel hydrocarbons and CAHs in the TS. Analytical data indicate that the highest concentrations of CAHs occur in the vicinity of well 2-62B, located approximately 100 to 120 feet upgradient from the former fire pit along an access road (Figure 5). The presence of a secondary source near well 2-65B (located further upgradient from well 2-62B and the former fire pit along another access road) is suggested by increasing concentrations of CAHs at that location.

Temporal data for concentrations of TCE and cis-1,2-DCE from monitoring wells/points at FTA-2 indicates that concentrations in the potential source areas are stable or gradually increasing. Given the available data, it appears that the plume is expanding downgradient from the suspected primary FTA-2 source and also at an upgradient location, based on increased concentration data for wells 2-63B, 2-64B, and 2-65B. An increase in concentrations within the center of the CAH plume at wells 2-72B and 2-274B further indicate the presence of secondary source(s) and/or expansion of the CAH plume in that area. Conversely, decreasing CAH concentration data for wells 2-355B and 2-393B along the plume axis at the most northern portion of the plume indicate the plume may be either decreasing or approaching equilibrium in this area.

The TS presented a conceptual model where the primary source of TCE was near the FTA-2 fire pit and well 2-62B, and the TCE plume migrated toward the northwest due to a preferential flow path and anisotropy of the aquifer matrix. This conceptual model was primarily based on the plume configuration. Linear transport in a northwest direction is suggested by the geometry of the CAH plume. Alternately, the TCE plume may be migrating in the direction of maximum hydraulic gradient (west-southwest) from multiple sources. The presence of TCE "hotspots" at wells 2-62B, 2-65B, and 2-274B does not support a conceptual model of a single source; rather, the "hotspots" support a conceptual model at multiple sources.

The presence of daughter products *cis*-1,2-DCE and VC provides strong evidence that parent solvents (TCE and/or PCE) are being reductively dehalogenated. However, a relative lack of VC and ethene indicates that the process does not significantly proceed past the initial step of TCE to DCE. Molar fractions of chlorinated ethenes along the apparent CAH plume axis indicates little, if any, degradation of TCE to DCE outside the

immediate source area(s). At well 2-65B, an overall decrease in the molar fraction of TCE, with an increase in molar fraction of DCE and the presence of VC indicates that TCE is degrading to DCE, and DCE to VC at that location. Geochemical data indicate that biodegradation of native or anthropogenic carbon is occurring via aerobic respiration and methanogenesis in suspected source areas, but that nitrate, ferric iron, and sulfate reduction are not significant anaerobic degradation processes at the site. Away from the suspected source areas, DO concentrations and ORPs indicate that environmental conditions are not sufficiently reducing for dehalogenation of CAH compounds.

Fate and transport modeling results presented in the TS indicated that dissolved TCE contamination present in groundwater north and west of the FTA-2 fire pit has the potential to migrate at least 1,000 feet beyond the August 1997 plume toe within 35 years. In addition, the model suggests that dissolved TCE concentrations in excess of the USEPA MCL groundwater quality standard of 5 μg/L for TCE will persist for more than 100 years. However, concentrations for wells located at the northern toe of the TCE plume (2-355B in the upper USZ sand interval and 2-393B in the lower USZ sand interval) decreased substantially from 1997 to 1999. Increases in the concentration of TCE at wells 2-64B and 2-272B in the upper USZ sand interval suggest a more westerly migration in the direction of maximum hydraulic conductivity. Increases in concentrations at wells 2-65B and 2-274B further suggest the presence of multiple CAH sources or "hotspots". Therefore, the numerical model presented in the TS should be considered as only one of multiple possible interpretations.

Continued long-term monitoring is recommended to evaluate the evolution of the CAH plume and the ability of microbially mediated biodegradation to limit and stabilize plume migration and expansion. To better evaluate the potential expansion of the plume, monitoring wells CG39-B97-45, 2-142B, 2-167B, and 2-356B are recommended to be included in future sampling events. Due to the uncertainty associated with the source(s) of CAHs, further source and site characterization is necessary to validate a conceptual fate and transport model for the FTA-2 site. Source characterization activities could include a soil gas survey, a geophysical survey, additional aquifer tests, and tracer tests as described in Section 6.4.2 of the TS (Parsons ES, 1999).

6.0 REFERENCES

- Bradley, P.M., and Chapelle, F.H., 1996, Anaerobic mineralization of vinyl chloride in Fe(III)-reducing aquifer sediments: Accepted for publication in Environmental Science and Technology, 1996.
- Montgomery, J.H. and L.M. Welkom. 1990. Groundwater Chemicals Desk Reference. Lewis Publishers, Inc. Chelsea, Michigan. 640 pp.
- Parsons ES, 1997. Work Plan for a Demonstration of Remediation by Natural Attenuation for Groundwater at Sites FTA 2 and Area A. Tinker Air Force Base, Oklahoma City, Oklahoma. March.
- Parsons ES, 1999. Final Demonstration of Remediation by Natural Attenuation for Groundwater at Sites FTA 2 and Area A. Tinker Air Force Base, Oklahoma City, Oklahoma. December.

- Stumm, W. and Morgan, J.J. 1981. Aquatic Chemistry. John Wiley & Sons, New York, NY.
- USEPA. 1998. Technical Protocol for Evaluating Natural Attenuation of Chlorinated Solvents in Ground Water. EPA/600/R-98/128. September.
- USEPA. 1996. National Primary Drinking Water Regulations.

Attachment A
Analytical Data

MEMORANDUM

MANTECH ENVIRONMENTAL RESEARCH SERVICES CORP. Environmental Science

In reply refer to: 99/JAD26

Contract # 68-c-98-138

To: Dr. Don Kampbell

Thru: Dr. Dennis Fine Hime

From: John Daniel

Subject: SF-0-68

Date: May 10, 1999

Copies: R.L Cosby

G.B. Smith

J. L. Seeley

As requested in Service Request # SF-0-68, headspace GC/MS analysis of 20 water samples from Tinker AFB for volatile organic compounds was completed. The samples were received April 11, 1999 and analyzed on May 4-5, 1999. RSKSOP-148 (Determination of Volatile Organic Compounds in Water by Automated Headspace Gas Chromatography/Mass Spectrometry (Saturn II Ion Trap Detector) was used for this analysis.

An internal standard calibration method was established for 15 compounds. The standard curves were prepared from 1.0 to 10000 ppb. The lower calibration limits were 1.0 ppb.

A quantitation report for the samples, lab duplicates, field duplicates, QCs, standards and lab blank is presented in table 1.

If you should have any questions, please feel free to contact me.

Table I. Quantitative Report for S.R. # SF-0-68 from Tinker AFB.

-5/99	2-648 Field Dup	ON	7	26.7	QN	302	÷:	Q	Q.	2.0	1870	7	1.3	2.3	2 :	2	2-392B		QV	9	QV	Q.	Q.	Q	Q :		Q	9	2 2	9 ⊊) }	2																	Printer, 3:37 PM 5/10/1999	
Date analyzed = 5/4-5/99	2-648	2	1.2	31.1	2	328	£.	Q	2	5.0	2130	1.3	1.5	9.3	2 :	Ş	2-274B	Field Dup	Q	1	4 .0	2	90.6	ı	2 :	Q (2.0 4060	200	ŀ	۲ ۲	2 5	2		BL0504A	9	2 5	Ş	2	S	QN	QN	Q	Q	Q !	2 2	2 5	2 2	Q		
Date	2-64A	Q	2	Q	Q	Q	Q	Q	Q	Q	2	Q	Q	S	2	Š	2-274B		Q	1	0.4	ᄝ	99.2	1	2	2	2.1	345	ł	1 ;	<u> </u>	2 2		QC0504F	200 ppp	550	202	205	224	193	196	201	185	180	189	206	205	205	control Std	
	2-63B	3.6	1	3.2	3.4	126	Q	Q	Q	1	53.7	1.0	Q	i	2	Q	2-274A	i	Q	Q	Q	Q	Q	Q	2	2	2 9	2 5	2 5	2 5	2 5	2 2	!	QC0504E	20 ppb	16.0	23.8	22.0	22.0	22.4	20.5	20.4	21.5	18.9	19.7	4.52	20.3	21.8	QC = Quality C	
= -	2-63A	2	2	2	Q	2	9	2	Q.	Q	2	2	9	Q	2	2	2.273B		Q	2	QN	2	Q	Q	Q	2	2 9	2 5	2 (2 5	2 5	5 5	2	QC0504D	200 ppb	156	187	200	237	213	209	212	216	213	200	228	212	204	Dup = Duplicate	
Originator ≈ G. Sewell Concentration ppb	2-62C	QV	QN	Q	2	1	Q	QN N	QN ON	Q.	6.4	Q	Q	Q.	2	ı	2,379B	27.77	Q	! 1	1.9	Q	34.9	ı	2	Q	1	455	1	ı	1 9	2 5	2	QC0504C	20 ppb	16.9	21.7	21.3	23.0	22.8	20.1	19.7	22.4	19.4	19.4	22.7	21.3	23.1	Limit(1.0 ppb)	
Orig	2-62B	QN	, 4G	47.5	S	650	5.0	QN	2	37.2	6200	3.7	19.5	8.0	15.5	101	2 4695	9901-7	Ç	2	2	1	6.3	2	Q	Q	2	0.1	2	2	2 :	2 <u>2</u>	2	Creek Grab		2	2 2	5 5	2 5	2 5	Q	2	Q	1.7	2	2 !	2 9	2 2	Below Calibration Limit(1.0 ppb) Dup = Duplicate QC = Quality Control Std	
	2-62A	Š	2	2	Q	Q	Q	Q	Q	Q	I	2	Q	S	Q.	2	9	960-7	5.1	4.	60	1.7	147	1	9	S	18.3	250	154	51.6	4.4	89. C	6.53	2-3558	Lab Dup	2	Q	1 9	787	. S	S	2	1	162	Q	Q	1 5	2 5	1	
	438	Š	} 1	3.3	Q	98.6	1	Q	Q	1.5	751	!	i	7	9	2	4	Aco-7	d CN	2	2	2	2	2	N	Q	Q	ı	2	2	2	2 9	Ž	2-355B		Q	Q Z	1 5	2 2	C 07	2 5	2	1	186	Q	Q	1 !	2 5	ž	
	43D	Ş	2 5	1.2	Q.	19.1	2	2	2	1	92.6	Q	2	2	Q	Q		X-9-7	5	2 5	2 5	Ş	2	2	Q	Q.	Q.	ı	2	Q	2	2 9	2	2-393B		2	2	2 9	2;	3.3	2 5	2 5	<u> </u>	114	Q.	2	1 !	2	l	
Date received = 4/11/99	Compound Sample Name:	VANYI CHI OBIDE	4 1-DICHLOROETHENE	T-1 2-DICHI OROETHENE	1.1-DICHLOROETHANE	C-1,2-DICHLOROETHENE	CHLOROFORM	1.1.1-TRICHLOROETHANE	CARBON TETRACHLORIDE	1.2-DICHLOROETHANE	TRICHI OROETHENE	TETRACHLOROETHENE	CHIOROBENZENE	1.3-DICHLOROBENZENE	1,4-DICHLOROBENZENE	1,2-DICHLOROBENZENE			SANS CHO SANS	4 + DICHI OBORTHENE	T-1 2-DICHI OROFITHENE	1 1-DICHI OROETHANE	C-1 2-DICHI OROFTHENE	CHIOROFORM	1,1,1-TRICHLOROETHANE	CARBON TETRACHLORIDE	1,2-DICHLOROETHANE	TRICHLOROETHENE	TETRACHLOROETHENE	CHLOROBENZENE	1,3-DICHLOROBENZENE	1,4-DICHLOROBENZENE	1,2-DICHLOROBENZENE			VINYL CHLORIDE	1,1-DICHLOROETHENE	T-1,2-DICHLOROETHENE	1,1-DICHLOROETHANE	C-1, Z-DICHLOROE I HENE	4 4 4 TBICHLOBOETUANE	CAPBON TETRACHI ORIDE	1 2-DICHI OROFTHANE	TRICHLOROETHENE	TETRACHLOROETHENE	CHLOROBENZENE	1,3-DICHLOROBENZENE	1,4-DICHLOROBENZENE	1,2-DICHLOROBENZENE	



MEMORANDUM

MANTECH ENVIRONMENTAL RESEARCH SERVICES CORP. Environmental Sciences

In reply refer to: 99-MB14

To: Dr. Don Kampbell

From: Mark Blankenship

THRU: Dr. Dennis Fine Down

Date: April 13, 1999

Copies: R.L. Cosby

G.B. Smith

J.L. Seeley

Please find attached the analytical results for Service Request SF-0-68 requesting the analysis of Tinker AFB, OK ground water samples to be analyzed for BTEXXX, TMB's and MTBE. The samples were collected April 7, 1999 and April 8, 1999. I received a total of 19 samples, in duplicate, in capped, 40 mL VOA vials on April 12, 1999. Samples were analyzed April 12,1999. The samples were acquired and processed using the Millennium data system. A 5 point (1-1000 ppb) external calibration curve was used to determine the concentration for for all compounds.

RSKSOP-122 "Analysis of Volatile Aromatic Hydrocarbons with Separation of Xylene Isomers by Purge and Trap Gas Chromatography" was used for these analyses. Autosampling was performed using a Dynatech Precision autosampler system in line with a Tekmar LSC 2000 concentrator.



Units = ng/ml Analyst M. Blankenship

10 PPB STD 20 PPB GA\GC GC LAB BLANK										
20 PPB QA\QC GC LAB BLANK		9.5	c c	6 CT	C					
GC LAB BLANK		21.4	2.7	5.0.5	9.6 9.6	9.6	9.4	10.3	9.8	9.4
The second secon		CN	: :	7. E	477	E.3.	21.8	21.2	23.2	20.1
43-D		S	9 5	9 ⊊	⊋ 5	₽!	₽	2	2	₽
43-S		Ş	2 5	2 2	⊋ :	9 :	9	9	2	2
2-62A		Ş	9 9	9 \$	⊋ 9	₽.	2	2	2	9
2-62B		, o	€	2 2	2 !	2	2	2	2	2
2-62C		SS	9 €	9 9	2 5	9 !	2	2	2	2
2-63A		S	9 €	<u>Ş</u> Ş	2 5	2 !	9	9	9	9
2-638		O IB	9 €	22	⊋ £	2	2	2	2	2
2-64A		S	- 13 G	<u> </u>	≥ 5	2	9	2	2	2
2-64B		91.0	Ç	9 5	2 5	⊋ :	₽.	2	2	2
2-64B LAB DUPLICATE		BLO	9	9 €	2 2	2 9	9 :	2	9	2
10 PPB STD		9.8	66	a ct	ŞÇ	⊋\$	⊋ ;	2	₽	2
2-65A		2	S) S	2.5	אַ כּיִצְּ ביי	70.Z	10.7	10.6	10.5
2-65B		BLO	0 18	2 €	9 \$	3 6	₽!	2	2	2
2-168B		BLQ	2	2 5	2 9	7. ·	2 9	₽!	2	9
2-272B		2	9	S C	Çă	ر د: و	⊋ £	2	BLQ	2.7
2-273B		2	2	Ş	3 2	ַבַּ בַּבַ	⊋ :	2 !	9	2
2-274A		2	Ş	- A	2 2	⊋;	⊋!	2	2	2
2-274B		BLQ	2	2 5	<u> </u>	2 9	⊋ !	9	2	2
2-355B		Q	Ç	2	9 9	⊋ 5	⊋!	₽	2	2
2-392B		9	7.5	9 €	2 2	29	2	2	9	2
2-393B		QV	81.0	2 5	2 5	⊋ 9	⊋ 5	3.7	2	2
20 PPB QA\QC	18.6	18.0	18.8	18.1	54. C&L	C at			2	9

Printed 04/15/99 SF-0-68

Originator: D. Kampbell

Site: Tinker AFB

Analyst: A. Zhao

					Comple And	wood 04/13/00
Sample Received 04/11/99			(0.)		ppm (Gas)	yzed 04/13/99 ppm (Water)
ı	opm (Gas)	ppm mg/L(Water)	ppm (Gas)	ppm mg/L(Water)	ppiii (Gas)	ppin (water)
400 mm CU4	9.35E+01	~	**	~ ·	**	~
the blant and	9.33E+U1	_	9.20E+01	~	**	~
100 ppm C2H4	**	~	**	~	9.60E+01	~
100 ppm C2H6	_	**	· ~	**	~	**
HP. Helium Blank	~	**	~	**	~	**
ab Blank	~	**	~	** .	~	**
43-D(4/8)	~	**	~	**	· ~	**
43-S(4/8)	~ .	**	~	**	~	**
2-62A(4/8)	~	0.122	_	**	~	**
2-62B(4/7)	~	U. 122 **	~	**	~	**
2-62C(4/7)	~	**	~	**	~	**
2-62C(4/7)	~			**	~	**
2-63B(4/8)	~	0.001	~	##	~	**
2-64B(4/8)	~		~	**		**
2-65B(4/8)	~	0.010	~	**	~	**
2-168B(4/8)	~		~	**	~	**
2-272B(4/8)	~	**	~	**	~	** **
2-272B Field Dup(4/8)	~	**	~	ma .	~	•
1000 PPM CH4	9.63E+02	. ~	**	~	deske	~
2-273B((4/8)	~	**	· ~	**	~	**
2-274A(4/8)	~	**	~	**	~	**
2-274B(4/8)	~	**	~	**	~	**
2-392B(4/8)	~	**	~	**	~	**
2-393B(4/8)	~	**	~	**	~	**
2-393B Lab Dup(4/8)	~	**	~	**	~	**
2-355B(4/8)	~	**	· ~	**	~	**
10 PPM CH4	9.10E+00	~	**	~	**	. ~
10PPM C2H4	**	~	1.06E+01	~	**	~
10PPM C2H6	**	~	· **	~	9.80E+00	~
Lower Limit of Quantitation	10.0	0.001	10.0	0.003	10.0	0.002

Units for the samples are mg/L dissolved in water. Units for the standards are parts per million.

sample date is represented in() 1999

^{**} denotes None Detected.

^{*} denotes Below Limit of Quantitation.

[~] denotes Not Applicable.



MEMORANDUM MANTECH ENVIRONMENTAL RESEARCH SERVICES CORP. Environmental Science

In reply refer to: 99-28LP/lp Contract # 68-C-98-138

To: Dr. Don Kampbell

Thru: D.D. Fine Bows

From: Lynda Pennington & KP

Subject: SR # SF-0-68

Ref:

Copies: R.L. Cosby

G.B. Smith

Date: April 19, 1999

J.L. Seeley

Attached are inorganic results for 16 Tinker AFB samples submitted to MERSC under Service Request # SF-0-68. The samples were received April 12, and were analyzed April 13, 1999. The methods used for analysis were Waters capillary electrophoresis method N-601 for chloride and suifate and Lachat FIA methods 10-107-04-2-A for nitrate+ nitrite and 10-107-06-1 for ammonia.

Quality control measures performed along with your samples included analysis of blanks, duplicates, spikes, known WPO samples and check standards.

If you have any questions concerning this data, please feel free to contact me.

SAMPLE	NH ₃ (N)	NO ₂ *+NO ₃ * (N)	CI ⁻	SO ₄ -2
	mg/L	mg/L	mg/L	mg/L
4/8 43-D	<0.10	3.60	55.7	38.3
4/8 43-S	<0.10	4.31	160	46.0
4/7 2-62A	<0.10	4.80	5.55	3.43
4/7 2-62B	<0.10	5.52	176	111
4/7 2-62C	<0.10	3.00	7.17	16.5
4/8 2-63B	<0.10	3.09	(112) (112)	(84.2) (84.3)
4/8 2-64B	<0.10	4.27	53.3	26.3
4/7 2-65B	<0.10	0.52	56.4	23.8
4/8 2-168B	(<0.10) (<0.10)	(0.41) (0.41)	3.24	36.1
4/8 2-272B	<0.10	2.83	106	28.9
4/8 2-273B	<0.10	0.60	18.6	9.75
4/7 2-274A	<0.10	8.0	10.7	10.1
4/7 2-274B	<0.10	4.32	82.6	27.2
4/7 2-355B	<0.10	3.32	202	15.5
4/8 2-392B	(<0.10) (<0.10)	0.43	(0.64) (0.69)	(6.39) (6.49)
4/7 2-393B	<0.10	(0.43) (0.41)	187	13.9
Blank	<0.10	<0.10	<.50	<.50
AQC	4.92	12.5	10.4	58.3
AQC T.V.	4.80	12.0	10.8	58.0
Cneck Std.	2.49	2.46	4.74	4.76
Check Std. T.V.	2.50	2.50	5.00	5.00
Spike Recovery	100%	102%	101%	101%



MEMORANDUM MANTECH ENVIRONMENTAL RESEARCH SERVICES CORP Environmental Science

> In reply refer to: 99-SH38

> > Contract #68-C-99-138

To: Dr. Don Kampbell From: Sharon Hightower

Dennis Fine Dome Thru:

April 15, 1999 Subject: SF-0-68 Date:

Copies: R.L. Cosby

G.B. Smith

J.L. Seeley & L.K. Pennington &

Attached are TOC results for 16 Tinker samples submitted April 11, 1999 under Service Request #SF-0-68. Sample analysis was begun April 14, 1999 and completed April 14, 1999 using

Blanks, duplicates, and AQC samples were analyzed along with your samples, as appropriate, for quality control. If you have any questions concerning this data, please feel free to ask me.

KAMPBELL TINKER LIQUIDS SF-0-68 SAMPLES RECEIVED 4/11/99 SAMPLES ANALYZED 4/14/99 BY SHARON HIGHTOWER

SAMPLE	MG/L TOC
2-62B, 4/7/99 2-62C, 4/7/99 2-63B, 4/8/99 2-64B, 4/7/99 2-65B, 4/7/99	3.33 3.71 1.36 2.84 4.38 8.40 2.72 2.58 4.42 3.66 3.69 22.3 5.13
2-273B, 4/8/99 2-274A, 4/7/99 2-274B, 4/7/99 2-355B, 4/7/99	3.99 2.85 3.49 2.65 8.30 2.34 5.21 22.7

Tinker AFB-March 1999-FTA-2 SiTe

•.										
	Sample	TOC	Temp.	D.O. mg/L	Redax	Cond.	PH SH	Fett ma/L	Sulfide ma/L	AIR Mg/L
<u></u>	2-64B		19.8	1.0	+177	921	7.3	4.1	<.1	360
	2-62 8	13.9	18,2	1, 2	+70	1435	7.2	0,1	<.)	360
	2-62A	57.0	wellpur	aped di	y-did	not recha	rg 2	0.1	<./	360
	2-620		17.9	5.4	+315	848	7.12	4.1	<./	380
	2-274A	:	19,4	3.4	+264	382	8.0	<,)	<.1	200
	2-2748	16.9	18.6	4,4	+247	970	7,4	<./	۷.1	380
	2-3558	15,9	19.0	6.1	+353	1323	7.0	0.1	۷.1	340
	2-658	17.0	17.5	0.2	+ 347	749	7.1	<./	4.	400
	12-65A	61.2	Lowys	eld-V	DA sam	ples only	/		:	
	2-3938	16.2	17.5	4.7	+270	: 620	7.2	<./	<.1	320
-	12-2728	17.2	18.9	1.7	+195	942	7.2	۷.۱	۷.)	300
	43-D	17.4	17.7	3. /	+ 313	1089	7. /	<.	<.1	40
	43-5	119.4	17.7	3.5	724	1293	7.0	<.1	۷./	280
	12-3928	: 15.)	17.4	5,5	+151	236	9,9	<.1	<.1	100
	2-2738	7.2	19.3	2.0	+277	.779	7.2	۷.۱	<.1	320
	2-63 B	10.9	119.4	0, 2	7244	11693	7.1	۷.۱	۷.۱	600
	2-63 A	54.0				Asonly				
		11,2	17.2	0.9	+283		6.9	۲۰۱	۷.1	520
	2-64A	56.6	Low wal	er yreld.	Took VOX	1's only				
			1							
	1			· · · · · · · · ·						
								:		
. !										-
•					1					

Attachment B Responses to Comments

PARSONS

Parsons Engineering Science, Inc. • A Unit of Parsons Infrastructure & Technology Group Inc.

1700 Broadway, Suite 900 • Denver, Colorado 80290 • (303) 831-8100 • Fax: (303) 831-8208 • www.parsons.com

27 December, 1999

Mr. Jerry Hansen AFCEE/ERT 3207 North Road, Bldg. 532 Brooks AFB, TX 78235-5363

Subject: Responses to AFCEE Comments on the Draft Addendum to the Remediation

by Natural Attenuation Treatability Study for Groundwater at Site FTA-2,

Tinker AFB, Oklahoma

Dear Mr. Hansen:

This letter provides responses to comments received on the Draft Addendum to the Remediation by Natural Attenuation (RNA) Treatability Study (TS) for Groundwater at Site FTA-2, Tinker Air Force Base (AFB), Oklahoma. The draft Addendum to the TS was prepared by Parsons Engineering Science, Inc. (Parsons ES) for the Air Force Center for Environmental Excellence Technology Transfer Division (AFCEE/ERT) and Tinker AFB. The intent of the draft Addendum to the TS was to determine the role of natural attenuation in remediating chlorinated solvent and fuel contamination in groundwater at FTA-2 over time. The draft Addendum was submitted to AFCEE in December 1999. Comments on the draft Addendum were received from AFCEE as reviewed by Jon Atkinson dated 22 December 1999. Responses to these comments were prepared by Parsons ES and are presented herein.

Responses to AFCEE Comments: 22 December 1999

Comment 1) Page 6, Table 2: Suggest rounding northing and easting values to one decimal place because most surveyed horizontal locations are accurate only to 0.1 foot.

<u>Parsons ES Response:</u> Northings and eastings were surveyed to two decimal places. Therefore, the reported surveying accuracy on page 6, Table 2 is correct, and no change is required.

Comment 2) Page 24, Sec 3.2.1, Para 2, Last Sent:

a. Suggest stating whether a slug test or other aquifer test has been conducted at well 2-392B, and if so, what was the measured hydraulic conductivity value.



b. Recommend depicting the areal extent of the stated low-permeability zone on Figure 5 or another appropriate figure and adding a representative east-west hydrogeologic cross section through well 2-392B.

Parsons ES Response:

- a. The text will be changed to indicate a slug test was conducted at well 2-393B, and that the measured hydraulic conductivity was 1.1 feet per day (ft/day). This information is also provided in Section 2.1, Upper Saturated Zone.
- b. A discussion of the low permeability zone is included in the TS (Section 3-3), along with an east-west cross-section (Cross-Section E-E'). References to these items will be added to the Addendum.
- Comment 3) Page 27, Sec 3.2.1, Para 1, Line 3: Recommend correcting the typo "re.".

Parsons ES Response: The text will be corrected as recommended.

- Comment 4) Page 30, Sec 3.2.1, Para 2, Sent 6: A Celsius temperature of 13.5 degrees corresponds to a Fahrenheit temperature of 56 degrees, not seven degrees.
- Parsons ES Response: The text will be corrected to indicate that the boiling point of vinyl chloride (VC) is -13.5 degrees Celsius, which corresponds to a Fahrenheit temperature of 7 degrees.
- Comment 5) Page 30 Sec 3.2.2, Line 2: Suggest writing out the "2" in "2 of the 14 wells" and other numbers less than 10 that occur throughout the report unless they describe units of analytical measurement (e.g., mg/L) or are components of dates. This practice of writing numbers less than 10 as words is consistent with the Air Force *The Tongue and Quill*.
- Parsons ES concurs with the practice of writing out numbers less than 10 for quantities that do not specify time, money, or measure. However, a noted exception to this practice is the use of numerals for numbers less than 10 when numbers greater than 10 describing the same item are used in the same sentence. Therefore, the sentence is correct as written, and no change is required.

Comment 6) Page 55, Sec 5.0, Para 1: Recommend briefly discussing if the noted hotspots support a conceptual site model of multiple sources with linear contaminant transport in a northwest direction..

Parsons ES Response: The noted hotspots do support a conceptual model of multiple sources, and this will be noted in the text. However, they do not necessarily support linear contaminant transport in a northwest direction. Linear transport in a northwest direction is suggested by the geometry of the CAH plume at the site.

Comment 7) Page 56, Para 1, Last Sent: Suggest that more details be provided regarding further source and site characterization. Specifically, recommend proposing number and location for additional soil borings and for temporary and/or permanent monitoring wells and media-specific analytes.

<u>Parsons ES Response:</u> The details of recommended source and site characterization are included the TS. A sentence will be added to the text referencing the recommended characterization described in the TS.

If you have any questions, have additional comments, or require additional information, please call me at (303) 831-8100.

Sincerely,

PARSONS ENGINEERING SCIENCE, INC.

Bruce M. Henry, P.G.

Bru M. Henry

Project Manager

cc: File 733939.20010